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# Early Life Exposure to Air Pollution and Autism Spectrum Disorder: Findings from a Multisite Case–Control Study

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# Abstract

**Background:** Epidemiologic studies have reported associations between prenatal and early postnatal air pollution exposure and autism spectrum disorder (ASD); however, findings differ by pollutant and developmental window.

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**Data and Code availability:** All code used for the analyses described in this paper can be available upon request. The outcome data are not available for replication due to the sensitive nature of the outcome. Air pollution exposure data are available upon request.

**Objectives:** We examined associations between early life exposure to  $PM_{2.5}$  and ozone in association with ASD across multiple US regions.

**Methods:** Our study participants included 674 children with confirmed ASD and 855 population controls from the Study to Explore Early Development, a multi-site case–control study of children born from 2003 to 2006 in the United States. We used a satellite-based model to assign air pollutant exposure averages during several critical periods of neurodevelopment: 3 months before pregnancy; each trimester of pregnancy; the entire pregnancy; and the first year of life. Logistic regression was used to estimate odds ratios (OR) and 95% confidence intervals (CIs), adjusting for study site, maternal age, maternal education, maternal race/ethnicity, maternal smoking, and month and year of birth.

**Results:** The air pollution–ASD associations appeared to vary by exposure time period. Ozone exposure during the third trimester was associated with ASD, with an OR of 1.2 (95% CI: 1.1, 1.4) per 6.6 ppb increase in ozone. We additionally observed a positive association with  $PM_{2.5}$  exposure during the first year of life [OR = 1.3 (95% CI: 1.0, 1.6) per 1.6 µg/m<sup>3</sup> increase in  $PM_{2.5}$ ].

**Conclusions:** Our study corroborates previous findings of a positive association between early life air pollution exposure and ASD, and identifies a potential critical window of exposure during the late prenatal and early postnatal periods.

#### Keywords

Air pollution; autism spectrum disorder; particulate matter; ozone

# Introduction

Autism spectrum disorder (ASD) is a group of neurodevelopmental disorders marked by impairments in social interaction and communication, and repetitive behaviors. The Centers for Disease Control and Prevention (CDC) estimates that 1 in 59 children has been identified with ASD.<sup>1</sup> The etiology of ASD is poorly understood; however, studies suggest contributions from both genetic and environmental factors.<sup>2-4</sup> Studies have suggested the prenatal and early postnatal periods to be critical windows of susceptibility for environmental toxicant exposure in relation to ASD.<sup>5,6</sup>

Particulate matter and ozone are among the most ubiquitous criteria air pollutants and have been shown to induce inflammation and oxidative stress,<sup>7-9</sup> both of which have been implicated in the development of ASD.<sup>10,11</sup> Several epidemiologic studies have reported associations between prenatal and early postnatal air pollution exposure and ASD; however, findings have differed by pollutant and developmental window.<sup>12</sup> Particulate matter 2.5 and

10  $\mu$ m in diameter (PM<sub>2.5</sub> and PM<sub>10</sub>) exposure during the prenatal<sup>13-16</sup> and postnatal<sup>15,17</sup> developmental periods has been associated with ASD in studies in the US, with two studies suggesting the third trimester may be a potential critical window of susceptibility.<sup>14,16</sup> In addition, one previous study found a positive association between prenatal ozone exposure and ASD.<sup>13</sup> Further, there is geographic variation in the composition of air pollution and may be geographic variation in associations between air pollution and ASD, however few studies have had the capacity to explore this potential heterogeneity. Additionally, few

studies to date have assessed associations between air pollution and ASD across different geographic locations using uniform exposure and outcome assessment methods.

To address these limitations, we investigated the association between early life exposure to  $PM_{2.5}$  and ozone in association with ASD using data from the Study to Explore Early Development (SEED). We assessed associations with exposures during critical periods of neurodevelopment, including the period three months before pregnancy, the entire pregnancy period, by trimester, and during the first year of life. We additionally investigated whether the association with  $PM_{2.5}$  and ozone differed for ASD with or without a co-occurring intellectual disability, by ASD severity, and by state of birth (SEED study site).

# Methods

#### **Study Population**

SEED is a multi-site case–control study with study sites for the first phase located within California, Colorado, Georgia, Maryland, North Carolina, and Pennsylvania.<sup>18</sup> Both cases and controls were eligible to participate in SEED if they were born in a study catchment area between September 2003 and August 2006 and resided there at 30-68 months of age. Children with possible ASD were ascertained through multiple sources serving or evaluating children with developmental problems, including: early intervention programs, special education programs, clinics, and individual providers. Population controls were identified by randomly sampling state birth records of children born in the specified date range to mothers that resided in the study catchment area at the time of delivery. Eligibility was based on residence, birthdate, and availability of a knowledgeable caregiver to participate in English or Spanish (California and Colorado only). An introductory letter was first sent to potential participants and those reached were then screened for eligibility through a follow-up recruitment call. For most, we were unable to determine the validity of their address, thus whether the target sample received mailed recruitment materials and were eligible, therefore our ability to determine response rates was limited.

Institutional review boards at each study site and at CDC approved SEED. Informed consent was obtained from all enrolled participants.

## **Outcome Ascertainment**

Upon telephone enrollment, caregivers of all children were administered the Social Communication Questionnaire (SCQ).<sup>19</sup> Any children who screened positive (SCQ score 11) or reported a previous ASD diagnosis received a comprehensive developmental assessment to determine final ASD classification. This assessment included two gold standard instruments, 1) the Autism Diagnostic Observation Schedule (ADOS),<sup>20</sup> 2) the Autism Diagnostic Interview-Revised (ADI-R).<sup>21,22</sup> Final ASD case classification was based on the results from the ADOS and ADI-R. ASD severity was additionally assessed using the ADOS calibrated severity score, which was calculated using the ADOS total score, ADOS language level, and chronological age at evaluation.<sup>23</sup> The ten-point scale was dichotomized into low/moderate (scores 4-7) and more severe (8-10).

All children participated in a general developmental assessment that included Mullen Scales of Early Development. Children who did not have an indication of possible ASD (negative SCQ screen, no previous ASD diagnosis and no ASD-specific service classification) received only this general developmental assessment. Children with a Mullen Early Learning Composite Score of <70 were classified as having an intellectual disability.<sup>24</sup>

#### **Exposure Assessment**

Exposure to air pollution was determined using data from satellite-based models linked to address at birth. Study participants' dates of birth and addresses at birth were identified using birth certificates. To ensure participant's privacy, all dates, including dates of birth, were randomly shifted by 0-14 days in either direction. For a given participant, the shift was the same for all dates in order to maintain the relation among dates. We calculated start date of pregnancy by subtracting the child's gestational age from their date of birth, using calculated LMP as a proxy for time of conception. Analyses used the clinical estimate of gestational age reported on birth certificates. Addresses at birth were geocoded in ArcGIS using the ESRI Maps street database. Geocoding match rates ranged from 95%-100% across study sites.

Average  $PM_{2.5}$  and ozone concentration estimates were derived at a daily temporal resolution and a 1×1 km spatial resolution using an exposure prediction model for the study exposure period years (2002-2007). This prediction model has been thoroughly described elsewhere<sup>25,26</sup> and has previously been used in a study of air pollution and mortality in the US.<sup>27</sup> Briefly, the prediction model incorporated data from a chemical transport model that simulates atmospheric chemistry, and is driven by meteorologic input from the Goddard Earth Observing System of the NASA Global Modeling and Assimilation Office (GEOS-Chem).<sup>28</sup> GEOS-Chem predictions were calibrated using monitored data. Land use terms (percentage of urban areas, population density, road density, and elevation), meteorologic variables (such as air temperature, precipitation, and wind speed), and satellite data were used to help calibrate GEOS-Chem outputs and to aid in downscaling to a finer spatial resolution. This previously developed model used a neural network to calibrate all the predictors to monitored PM<sub>2.5</sub> and ozone and was trained and validated with ten-fold cross-validation.

Participants were matched to the centroid of the nearest grid cell based on their residence at birth. Daily  $PM_{2.5}$  and ozone concentrations were averaged for several periods, including the  $1^{st}$  (weeks 1-13 of pregnancy),  $2^{nd}$  (weeks 14-26), and  $3^{rd}$  trimesters (weeks 27 to birth), preconception (3 months prior to conception), entire pregnancy, and year post birth.

#### Covariates

We obtained information to assess potential confounders from the caregiver interview and birth certificates. We used a directed acyclic graph to identify the covariate adjustment set to be included in the model that would result in the least biased estimate (see eFigure 1).<sup>29</sup> The final adjustment set consisted of the following variables: maternal age (continuous), maternal race/ethnicity (non-Hispanic white, other race/ethnicity), maternal education (< bachelor's degree, bachelor's degree), maternal smoking (any smoking three months

before conception or during pregnancy), and year of birth (2003-2004, 2005-2006). Season of birth is related to the levels of air pollution and has been linked with ASD in some previous studies. Therefore, analyses additionally included an indicator variable for month of birth to control for seasonal trend. All models included an indicator for study site (California, Colorado, Georgia, Maryland, North Carolina, and Pennsylvania). We excluded participants from the current analyses if they did not have a geocoded address (*n*=44) or were missing data on key covariates (*n*=32). In total, 16 cases and 16 controls were missing covariate data on maternal tobacco use and race/ethnicity.

#### Statistical Analyses

We used logistic regression to estimate adjusted odds ratios (OR) and corresponding 95% confidence intervals (CI) for the associations between  $PM_{2.5}$  and, separately, ozone with ASD, with the population group serving as the control group for all analyses. We first modeled the exposure–response function between air pollution and ASD graphically using natural cubic spline models (eFigure 2). We analyzed  $PM_{2.5}$  and ozone exposures as continuous measures because a continuous term fit better than categorical coding (quartiles), per the lower Akaike Information Criterion value and because continuous coding allows comparability with the results from previous studies. Effect estimates are scaled to the interquartile range (IQR) value for the entire pregnancy period, averaged across study sites. This approach allows the comparison of effect estimates across pollutants, study sites, and developmental windows. We report results for each developmental window modeled separately and also mutually adjusted for the other windows, in order to isolate potentially critical windows of susceptibility. Specifically, results are presented for each trimester mutually adjusted for the other trimesters and, for the preconception, pregnancy, and postnatal windows, mutually adjusted for the other developmental windows.

We assessed effect measure modification by study site for each of the developmental windows in order to assess possible heterogeneity in the  $PM_{2.5}$  and ozone and ASD associations by geographic location. We conducted likelihood ratio tests by comparing models with and without interaction terms for study site and used a p<0.10 as a threshold for presence of modification on the multiplicative scale.

We additionally investigated if  $PM_{2.5}$  and ozone and ASD associations differed for children with ASD depending on the presence of a co-occurring intellectual disability. Further, we compared associations by ASD severity using a dichotomized ASD severity score.<sup>23</sup> We compared the two severity subgroups to the population controls using multinomial logistic regression.

We conducted several sensitivity analyses. Ozone may confound relationships between PM<sub>2.5</sub> and ASD.<sup>30</sup> Therefore, we report results from two-pollutant models for each developmental window by controlling for the other pollutant in each exposure model. We additionally report results adjusted for monitored levels of nitrogen dioxide (NO<sub>2</sub>), as a marker of traffic-related air pollution. Air pollution exposure has been associated with preterm birth<sup>31</sup> and preterm birth is a risk factor for ASD.<sup>32</sup> Therefore, we did not control for preterm birth was due to its potential mediating role.<sup>33</sup> Instead, we assessed the potential mediating role of preterm birth by comparing air pollution and ASD models without an

indicator for preterm birth status to those adjusted for preterm birth. Finally, we used ordinal logistic regression to assess associations between air pollution exposure and ASD severity using the ordinal measure of the severity score. Since controls were not given a severity score, these analyses were among ASD cases only.

# Results

The final sample consisted of 1,529 mother–child pairs with complete outcome classification information based on a completed developmental assessment and geocoded address data available. This included 674 cases of children with ASD and 855 population controls. Compared to the population controls, children with ASD were more likely to be boys, born preterm, and born to non-white mothers and mothers with lower level of education (Table 1). These differences were consistent across SEED study sites (eTable 1).

 $PM_{2.5}$  and ozone levels varied both spatially and temporally within and across the SEED study areas, which were generally centered around large urban areas (Figure 1; also see eFigure 3). During the study years,  $PM_{2.5}$  peaked during the summer months for the Eastern study areas and during the winter months for the California study site. There was less temporal variation in  $PM_{2.5}$  in the Colorado study area. In contrast, ozone concentrations peaked in the summer months for all of the study areas.

The overall  $PM_{2.5}$  average for the pregnancy period was 12.7 µg/m<sup>3</sup> (range: 4.9 – 18.6 µg/m<sup>3</sup>) among all controls (Table 2), and varied considerably across study sites, ranging from 9.0 µg/m<sup>3</sup> in CO to 15.5 µg/m<sup>3</sup> in GA. The overall ozone average for the pregnancy period was 37.1 ppb, and was lowest in California (29.5 ppb) and highest in Colorado (39.8 ppb). Mean PM<sub>2.5</sub> levels were strongly correlated across several of the developmental periods (eTable 2), particularly for pregnancy and first year of life (correlation of 0.92). Pollutant levels were often inversely correlated between the first and third trimester; e.g. for PM<sub>2.5</sub> in California (-0.64), Georgia (-0.65), and North Carolina (-0.71), and for ozone in Colorado (-0.78), Georgia (-0.86), Maryland (-0.85), North Carolina (-0.89), and Pennsylvania (-0.86).

Adjusted associations between air pollutants and ASD differed across developmental periods. For both ozone and  $PM_{2.5}$ , we observed null associations with ASD for exposure during preconception, the pregnancy average, and the first and second pregnancy trimesters (Table 3). We observed elevated associations with ASD for ozone exposure during the third trimester: OR=1.2 per the interquartile range width (IQR<sub>W</sub>) of 6.6 ppb (95% CI: 1.1, 1.4), and for PM<sub>2.5</sub> exposure during the first year of life: OR=1.3 (95% CI: 1.0, 1.6) per 1.6 µg/m<sup>3</sup> increase in PM<sub>2.5</sub> (Table 3; also see eTable 3 for unadjusted results). Results were similar, but with wider CIs, comparing two-pollutant to single pollutant models (eTable 4). Results were additionally similar in models adjusted for preterm birth (eTable 3) and monitored levels of NO<sub>2</sub> (eTable 5). Finally, results were generally similar across the developmental windows and pollutants when adjusting for sex of the child (eTable 3).

Table 3 also includes results from models for trimester-specific associations mutually adjusted for the other trimesters, as well as from models mutually adjusted for the

preconception, entire pregnancy, and first year of life exposure periods. The elevated OR for third trimester ozone exposure increased in models mutually adjusted for exposure during trimesters 2 and 3 [OR=1.3 (95% CI: 1.1, 1.5)] (Table 3). In models mutually adjusted for the preconception period and the first year of life, an association between ozone exposure during pregnancy and ASD was observed [OR=1.4 (95% CI: 1.0, 1.8). Adjusting for exposures during preconception and pregnancy, the OR for PM<sub>2.5</sub> during the first year of life remained similar [OR: 1.3 (95% CI: 0.98, 1.7)].

The association between ASD and PM<sub>2.5</sub> exposure during the first year of life did not vary across study sites (P = 0.14) (Figure 2; also see eTable 6 for numeric data). However, there did appear to be effect modification by site for the associations with PM<sub>2.5</sub> during the first (P = 0.05) and third (P = 0.03) trimesters, as well entire pregnancy exposures (P = 0.07). As results for the Pennsylvania site appeared to be outliers, we re-ran models excluding them (eTable 7); the overall association of first year of life PM<sub>2.5</sub> exposure and ASD was somewhat driven by this site, but less so for the model that also adjusted for the preconception and pregnancy exposure periods. The association observed between third trimester ozone exposure and ASD did not vary across sites (Figure 3) and was slightly strengthened by excluding PA data. Modification by study site was suggested for associations with ozone exposure during the preconception period (P = 0.02), first trimester (P = 0.06), and first year of life (P < 0.001). The inverse associations of ozone exposure observed during the first year of life (overall or adjusted for preconception and pregnancy exposure) appeared to be largely driven by the Pennsylvania study site (eTable 7).

When exploring whether associations differed for phenotypic case groups: severe versus mild ASD symptoms, and, separately, the co-occurrence of an intellectual disability, result patterns were generally similar across the groups (Table 4 and eTable 8). Additionally, there were no consistent associations between air pollution and ASD severity among ASD cases (eTable 9).

# Discussion

In this US-based multisite case–control study,  $PM_{2.5}$  exposure during the first year of life was associated with increased odds of autism spectrum disorder. Our trimester-specific analyses also revealed the third trimester as a potentially critical window of susceptibility for ozone. These findings held in models mutually adjusted for the other developmental windows, as well as two-pollutant models. Associations for third trimester ozone and first year of life  $PM_{2.5}$  exposure were fairly consistent across study sites. A growing number of studies have noted associations with air pollution exposure during the third trimester or early postnatal period.<sup>16,34</sup> This reflects a period of development where the brain structures are formed, but neuronal maturity (myelination and synaptic density) is rapidly increasing.<sup>35</sup>

There are several plausible mechanistic pathways for late prenatal and early postnatal air pollution to contribute to the development of ASD. Synapse formation and neurotransmitter receptor formation both occur during the third trimester of development.<sup>35</sup> Additionally, particulate matter and ozone are established immune toxicants.<sup>36,37</sup> Inflammatory cytokines, produced from the mother in response to both ozone and PM<sub>2.5</sub> exposure, can cross the

placenta, enter fetal circulation, and induce systemic inflammation and oxidative stress.<sup>38</sup> Pollutants directly inhaled by the infant postnatally can additionally induce a systemic inflammatory response and perturb normal development of the immune system. These proinflammatory cytokines may reach the developing brain, cross the blood brain barrier, resulting in neuroinflammation, neuron damage/loss, microglia activation, and DNA damage.<sup>9</sup> Of emerging interest is the mediating role of microglia activation in the association between early life toxicant exposure and ASD.<sup>39</sup> Alterations in microglial development by late prenatal and early postnatal inflammation may alter synaptic pruning, <sup>40,41</sup> resulting in altered neuronal connectivity, and disruption of normal brain development.

The finding that  $PM_{2.5}$  exposure specifically during the first year of life is associated with increased odds of ASD is consistent with recent studies in California,<sup>15</sup> Pennsylvania,<sup>17</sup> Ohio,<sup>42</sup> and Denmark<sup>43</sup> despite some differences in exposure levels (see eFigures 4 and 5 for comparison of studies). The observation that the first year of life may be a specifically susceptible window for exposure during development is consistent with a recent study based in Israel that found associations with postnatal NO<sub>2</sub> exposure, specifically when also adjusting for pregnancy averages.<sup>34</sup> There were additionally slight increases in odds of ASD for PM<sub>2.5</sub> exposure during the third trimester, though results were imprecise; this finding is consistent with two recent US-based studies of PM<sub>2.5</sub>.<sup>14</sup> and PM<sub>10</sub><sup>16</sup> exposure in relation to ASD. The association with ozone exposure was higher when exposure occurred in the third trimester. In a previous California-based study, Becerra et al.<sup>13</sup> found associations between ozone exposure during the entire pregnancy and ASD in single and multi-pollutant models additionally adjusted for PM<sub>2.5</sub>.

We assessed the association between  $PM_{2.5}$  and ozone and ASD using similar exposure assessment and outcome ascertainment methods across study sites from different geographic regions in the US. Although associations for third trimester ozone and first year of life  $PM_{2.5}$ exposure did not vary across different geographic locations in our study, there was heterogeneity in site-specific findings for several of the other windows. For instance, for entire pregnancy exposure averages, a few of the study sites had inverse associations, while we observed elevated odds ratios for sites such as Georgia and Pennsylvania. Exposure levels varied between sites, with the highest  $PM_{2.5}$  levels in Georgia, and the lowest in Colorado and California. In addition, on the east coast  $PM_{2.5}$  is dominated by sulfates from power plants and other regional sources; in contrast, nitrates make up the largest composition of  $PM_{2.5}$  on the west coast.<sup>44</sup> Thus, the regional differences in particulate matter composition and exposure levels could contribute to differences in the estimated associations between air pollution and ASD from different geographic regions described in the literature, and across sites in our analysis.

In order to correct for the inherent correlation of exposure among developmental windows,<sup>45</sup> we present results mutually adjusted for the other developmental windows of interest. Our finding of a direct association for ozone exposure during pregnancy and an inverse association during the first year of life, particularly in mutually adjusted models, may be partially explained by the complex correlation structure between different developmental windows of both the same and different pollutants, i.e. the negative correlation between ozone and PM<sub>2.5</sub> and markers of traffic-related air pollution. This was particularly true for

the Pennsylvania study site for the year post birth (correlation of -0.54 between ozone and PM<sub>2.5</sub> and -0.69 between ozone and monitored NO<sub>2</sub>). These findings additionally highlight the complexity of identifying critical windows of susceptibility across pollutants with differing spatial and temporal variability.

The stronger signal of association during the late prenatal and early postnatal developmental windows could also reflect more accurate exposure measures during these windows. We assigned exposure from just before pregnancy to one year after birth based on only the residential address at birth, which assumes limited mobility during pregnancy and the first year of life. Further, if women in this study spent more time away from home during the earlier pregnancy periods,<sup>46</sup> then the exposure estimates may not have been an accurate proxy for their actual exposure during these periods.<sup>47</sup> Therefore, it is possible that exposure misclassification for earlier pregnancy periods may have attenuated effect estimates for these windows. Previous studies have assessed the extent of air pollution exposure assignment<sup>48</sup> and risk effect estimation<sup>49</sup> when the full address history was incorporated. In addition, Warren et al.<sup>50</sup> assessed the impact of incomplete residential history on identification of critical windows of susceptibility for air pollution exposure and found minimal exposure misclassification and impact on the identification of critical windows.

We were unable to explore shorter windows of exposure because of procedures introduced to ensure privacy of participants. The dates of birth of SEED participants were shifted by up to two weeks in either direction; this shift in date of birth did not impact longer exposure average periods (trimester, year post birth, etc.), based on the sensitivity analyses in the one site for which both actual and shifted data were available, but prevented us from confidently studying shorter exposure periods (such as weeks).

ASD represents a heterogeneous group of neurodevelopmental disorders. Because of this heterogeneity, we attempted to refine the case definition of ASD by conducting analyses with individuals grouped by their severity score level and also by presence of a co-occurring intellectual disability. Our goal in these analyses was to compare the refined case definitions of ASD (moderate/ high severity and co-occurring ID) to the population control group, similar to our approach when using the standard dichotomized ASD measure. These analyses did not reveal marked differences by level of severity or presence of a co-occurring intellectual disability. We additionally assessed associations between  $PM_{2.5}$  and ozone and ASD severity among ASD cases only. These analyses revealed no consistent associations between air pollution and the continuous calibrated severity score. This finding is consistent with a recent study that found associations between early life air pollution exposure and deficits in cognitive and adaptive function among children with ASD, but not ASD severity. 51

Although we adjusted for various markers of socioeconomic status in our analyses, including maternal age and education, there is still the potential for residual confounding by socioeconomic status or place of residence. Our adjustment for study site produced estimates

from models that accounted for the multi-site study design, which also further adjusts for any residual confounding due to between-site variability in air pollution exposure.

SEED aimed to include mother–child pairs from diverse population subgroups by recruiting children with disabilities who were recruited from multiple clinical and education sources and sampling population-based controls from birth certificates. A limitation of this approach is that many targeted families could not be located, potentially because they had moved, possibly outside of the study area, and information about these families was limited. We recognize the importance of considering the influence of selection bias when interpreting results. While participants in all groups were recruited from defined geographic catchment areas, population controls were, in general, more educated and affluent than ASD participants.<sup>52</sup> Maternal age and education were associated with participation in the Georgia site; the only site with the ability to fully assess predictors of non-response.<sup>53</sup> Consequently, we adjusted for maternal age and education in our analyses; still some residual influence of sample selectivity is possible and the direction of such bias is indeterminable.

Our approach involved numerous statistical models, across developmental windows, pollutants, study sites, etc. Some of the conclusions, for example, whether associations with air pollutants differed by autism severity, were limited by a modest sample size in subgroups. Furthermore, the large number of statistical comparisons increases the overall chance of false-positive findings, and for this reason we emphasize results that were robust across several modeling approaches rather than relying solely on statistical significance testing.

Finally, we assessed associations with satellite-based modeled estimates of  $PM_{2.5}$  and ozone in relation to ASD. Given that previous studies have found associations between NO<sub>2</sub> and ASD<sup>15,34</sup>, it would have been of interest to incorporate modeled NO<sub>2</sub> estimates in the current study. Unfortunately, that data was not available at the time of our analyses.

This study has several strengths. First, it is the first to assess associations between  $PM_{2.5}$  and ozone and ASD using an air pollution model that incorporates satellite-based data at a fine spatial resolution, as well as land-use terms, meteorologic variables, and data from a chemical transport model. Land-use regression models have adequate spatial resolution, but often have poor temporal resolution since the land-use terms are usually time-invariant.<sup>25</sup> The use of satellite data helps to improve temporal resolution in the resulting exposure predictions, and allows the inclusion of rural locations further away from monitors.<sup>25</sup>

SEED study participants were drawn from a variety of settings serving children with disabilities across six different study sites located in the western, central, and eastern US. This ensures greater representation of ASD than previous clinically oriented studies. The geographic variability also provides the necessary exposure variability to assess health-related associations from exposure to  $PM_{2.5}$  and ozone. SEED additionally provided rich covariate data, allowing thorough adjustment for potential confounding. In addition to standard confounders, we additionally adjusted for year and month of birth in all analyses to account for seasonal trends related to both air pollution exposure and ASD.<sup>54,55</sup> Finally, SEED used gold standard outcome assessment tools, including ADOS and ADI-R, therefore

reducing any potential outcome misclassification. This allowed us to investigate the heterogeneity of ASD and to distinguish associations by core ASD symptom severity.

# Conclusions

In summary, our study corroborates previous findings of a positive association between early life exposure to  $PM_{2.5}$  and ozone and ASD, and also confirms prior findings of potential late prenatal and early postnatal windows of susceptibility. Given some heterogeneity in results across our study sites representing multiple regions of the United States, future research to explore the influence of variable PM composition by geographic region is warranted.

# **Supplementary Material**

Refer to Web version on PubMed Central for supplementary material.

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#### Figure 1.

Monthly (A)  $PM_{2.5}$  (µg/m<sup>3</sup>) and (B) ozone (ppb) concentrations from 2002-2007, by SEED study site.

Developmental W Study Site	lindow,	Cases (N)	Controls (N)	OR (95% CI)	<b>P</b> <sub>int</sub>
Preconception					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	┄╺╌ └╼╌ └╼╌ └┲╴ └╼╴	96 139 130 107 100 102 674	134 185 160 126 146 104	1.1 (0.96, 1.3) 0.98 (0.76, 1.3) 1.1 (0.89, 1.2) 0.95 (0.79, 1.2) 1.1 (0.90, 1.3) 0.91 (0.71, 1.2)	0.70
Entiro Progranov		0/4	000	1.0 (0.97, 1.1)	0.76
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 100 −1 102 674	134 185 160 126 146 104 855	0.71 (0.51, 1.0) 0.89 (0.64, 1.2) 1.5 (1.0, 2.1) 0.93 (0.66, 1.3) 1.1 (0.73, 1.6) 3.1 (1.8, 5.3) 1.0 (0.88, 1.2)	0.07
First Trimester				(,,	
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 100 102 674	134 185 160 126 146 104 855	0.86 (0.76, 0.97) 0.98 (0.77, 1.2) 1.2 (1.0, 1.5) 1.0 (0.83, 1.3) 0.99 (0.82, 1.2) 1.3 (1.0, 1.7) 0.98 (0.91, 1, 1)	0.05
Second Trimester					0.00
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 100 102 674	134 185 160 126 146 104 855	0.93 (0.82, 1.0) 0.94 (0.75, 1.2) 1.1 (0.90, 1.2) 1.0 (0.84, 1.2) 0.95 (0.79, 1.2) 1.1 (0.87, 1.5) 0.97 (0.91, 1.1)	0.89
Third Trimester					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	├ <del>■</del>	96 139 130 107 100 102 674	134 185 160 126 146 104 855	1.1 (0.94, 1.3) 0.84 (0.66, 1.1) 0.99 (0.85, 1.2) 0.92 (0.77, 1.1) 1.2 (1.02, 1.4) 1.3 (1.06, 1.6) 1.1 (0.98, 1.1)	0.03
Year Post Birth					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 100 —102 674	134 185 160 126 146 104 855	1.5 (0.84, 2.7) 1.1 (0.74, 1.7) 1.5 (0.80, 3.0) 1.0 (0.66, 1.5) 0.90 (0.50, 1.6) 2.8 (1.4, 5.5) 1.3 (1.0, 1.6)	0.14
		_			
	0.2 1	5			
	Odds Ratio				

#### Figure 2.

Site-specific and overall adjusted odds ratios (OR) and 95% confidence intervals (CI) for the associations between  $PM_{2.5}$  and ASD, by developmental window. Results are reported per 1.6-µg/m<sup>3</sup> increase in  $PM_{2.5}$ . All models are adjusted for maternal age, maternal education, maternal race/ethnicity, maternal smoking, month of birth, and year of birth. Interaction p-values for each developmental window are additionally included.

Developmental W Study Site	indow,	Case (N)	s Controls (N)	OR (95% CI)	<b>P</b> <sub>int</sub>
Preconception					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	⊢ ■ - { + ≱-1 + ≠-1 + ■:1 + ■:1 + ■:1 + ■:1	96 139 130 107 100 102 674	134 185 160 126 146 104 855	0.71 (0.52, 0.97) 1.0 (0.88, 1.2) 0.97 (0.81, 1.2) 0.89 (0.74, 1.1) 0.88 (0.71, 1.1) 0.75 (0.61, 0.90) 0.92 (0.82, 1.1)	0.02
Entire Pregnancy					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 100 102 674	134 185 160 126 146 104 855	1.2 (0.72, 2.0) 0.95 (0.70, 1.3) 0.86 (0.58, 1.3) 1.4 (0.89, 2.1) 1.3 (0.74, 2.2) 1.1 (0.76, 1.7) 1.1 (0.85, 1.3)	0.64
First Trimester				(,,	
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	;  ÷=-   =;  +;   -≠;   -≠;   +;	96 139 130 107 100 102 674	134 185 160 126 146 104 855	1.2 (0.92, 1.7) 0.91 (0.77, 1.1) 0.98 (0.82, 1.2) 0.98 (0.79, 1.2) 0.86 (0.69, 1.1) 0.94 (0.78, 1.1) 0.94 (0.83, 1, 1)	0.06
Second Trimester		011	000	0.04 (0.00, 111)	0.00
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	:  -₩-1  +₩-1  +₩-1  +₩-1  +₩-1	96 139 130 107 100 102 674	134 185 160 126 146 104 855	1.0 (0.75, 1.3) 0.90 (0.76, 1.1) 0.88 (0.74, 1.1) 1.1 (0.88, 1.3) 1.0 (0.80, 1.3) 1.1 (0.88, 1.3) 0.96 (0.85, 1.1)	0.43
Third Trimester		0,1	000	0.00 (0.00, 111)	0.40
California Colorado Georgia Maryland North Carolina Pennsylvania Overall	; ;=-1 ;;=-1 ;;=-1 ;;==1 ;;==1 ;;=1	96 139 107 107 100 102 674	134 185 160 126 146 104 855	1.1 (0.78, 1.5) 1.3 (1.1, 1.6) 1.1 (0.95, 1.4) 1.2 (0.96, 1.5) 1.5 (1.2, 1.8) 1.2 (0.99, 1.4) 1.2 (1.1, 1.4)	0.23
Year Post Birth					
California Colorado Georgia Maryland North Carolina Pennsylvania Overall		96 139 130 107 ⊢ 100 102 674	134 185 160 126 146 104 855	$\begin{array}{c} 0.83 \ (0.45, 1.5) \\ 1.4 \ (0.91, 2.1) \\ 0.62 \ (0.30, 1.3) \\ 0.52 \ (0.26, 1.0) \\ 3.4 \ (1.2, 9.4) \\ 0.18 \ (0.08, 0.39) \\ 0.79 \ (0.60, 1.0) \end{array}$	<0.001
		1			
	0.1 1	10			
	Odds Ratio				

#### Figure 3.

Site-specific and overall adjusted odds ratios (OR) and 95% confidence intervals (CI) for the associations between ozone and ASD, by developmental window. Results are reported per 6.6-ppb increase in ozone. All models are adjusted for maternal age, maternal education, maternal race/ethnicity, maternal smoking, month of birth, and year of birth. Interaction p-values for each developmental window are additionally included.

# Table 1.

Characteristics of ASD Cases and Population Controls.

Characteristic	ASD Cases (n=674)	Controls (n=855)
Child sex, N (%)		
Male	551 (82)	453 (53)
Female	123 (18)	402 (47)
Birth year, N (%)		
2003-2004	273 (41)	393 (46)
2005-2006	401 (59)	462 (54)
Maternal race/ethnicity, N (%)		
Non-Hispanic white	378 (56)	611 (71)
Other <sup>a</sup>	296 (44)	244 (29)
Maternal education, N (%)		
<bachelor's< td=""><td>331 (49)</td><td>289 (34)</td></bachelor's<>	331 (49)	289 (34)
Bachelor's	343 (51)	566 (66)
Maternal age at birth (years), N (%)		
<35	486 (72)	589 (69)
35	188 (28)	266 (31)
Maternal smoking, N (%)		
Yes	112 (17)	79 (9)
No	562 (83)	776 (91)
Preterm, N (%)		
Yes	111 (16)	82 (10)
No	563 (84)	773 (90)
SEED study site, N (%)		
California	96 (14)	134 (16)
Colorado	139 (21)	185 (22)
Georgia	130 (19)	160 (19)
Maryland	107 (16)	126 (15)
North Carolina	100 (15)	146 (17)
Pennsylvania	102 (15)	104 (12)
$PM_{2.5}~(\mu g/m^3),$ mean $\pm~SD$		
Pregnancy	$12.8\pm2.7$	12.7 (2.6)
First year of life	$12.7\pm2.5$	12.5 (2.5)
Ozone (ppb), mean ± SD		
Pregnancy	$36.8\pm5.7$	$37.0\pm5.7$
First year of life	$37.8 \pm 4.4$	$38.1 \pm 4.3$

ASD indicates Autism Spectrum Disorder; PM2.5, particulate matter <2.5 µm; SEED, Study to Explore Early Development; SD, standard deviation.

<sup>a</sup>Includes African American, Asian, Hispanic, multiracial, and all others.

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 $PM_{2.5}$  ( $\mu g/m^3$ ) and Ozone (ppb) Exposure Distribution [mean (interquartile range width, IQR<sub>w</sub>)] by Study to Explore Early Development Study Site and Developmental Windows, among Controls Only.

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	All sites Mean (IQR <sub>W</sub> )	California Mean (IQR <sub>W</sub> )	Colorado Mean (IQR <sub>W</sub> )	Georgia Mean (IQR <sub>W</sub> )	Maryland Mean (IQR <sub>W</sub> )	North Carolina Mean (IQR <sub>W</sub> )	Pennsylvania Mean (IQR <sub>W</sub> )
<b>Ozone</b> (ppb)							
Preconception	37.7 (18.8)	31.6 (9.4)	39.8 (20.8)	37.5 (21.5)	37.5 (21.3)	40.3 (13.5)	39.1 (17.9)
Entire pregnancy	37.1 (8.5)	29.5 (4.5)	39.8 (7.9)	39.3 (7.5)	36.8 (7.8)	39.7 (4.4)	35.3 (7.5)
First trimester	36.1 (21.2)	28.8 (13.1)	39.8 (21.3)	36.9 (25.1)	34.2 (16.6)	40.2 (21.7)	34.7 (25.4)
Second trimester	37.2 (20.0)	29.1 (13.9)	40.6 (18.6)	40.5 (18.8)	36.2 (21.9)	40.0 (14.3)	33.7 (17.2)
Third trimester	38.2 (20.8)	30.7 (11.7)	39.2 (21.3)	41.1 (21.9)	40.3 (14.1)	38.6 (21.6)	38.0 (24.1)
First year of life	38.2 (5.0)	31.1 (3.2)	40.0 (5.3)	40.0 (3.4)	38.2 (3.5)	40.5 (2.6)	38.3 (3.3)
$\mathbf{PM}_{2.5}~(\mu g/m^3)$							
Preconception	12.7 (5.0)	10.7 (5.2)	9.1 (2.2)	15.8 (4.8)	15.3 (3.9)	12.9 (3.3)	13.3 (3.2)
Entire pregnancy	12.7 (3.9)	11.6 (1.8)	9.0 (1.6)	15.5 (1.7)	14.8 (1.8)	13.0 (1.5)	12.9 (1.1)
First trimester	12.6 (5.2)	11.9 (7.7)	9.0 (2.3)	15.0 (3.0)	14.3 (2.5)	13.3 (4.8)	13.0 (3.7)
Second trimester	12.7 (4.7)	12.1 (7.9)	8.9 (2.3)	15.6 (4.8)	14.7 (2.4)	13.0 (3.9)	12.6 (2.1)
Third trimester	12.7 (5.6)	10.8 (5.8)	9.1 (2.3)	16.1 (4.8)	15.5 (5.0)	12.9 (4.8)	13.1 (2.7)
First year of life	12.5 (3.9)	11.3 (1.1)	8.7 (1.5)	15.5 (0.8)	14.5 (1.6)	13.1 (0.9)	13.0 (0.8)

## Table 3.

Adjusted<sup>*a*</sup> Associations between PM<sub>2.5</sub> and Ozone and ASD, for Single and Mutual Window Adjustment Models, among Study to Explore Early Development Participants.

	Ozone OR (95%CI)	PM <sub>2.5</sub> OR (95%CI)
First Trimester		
Modeled separately	0.94 (0.83, 1.1)	0.98 (0.91, 1.1)
Mutual adjustment for exposures during trimesters 2 and 3	0.98 (0.85, 1.1)	1.0 (0.93, 1.1)
Second Trimester		
Modeled separately	0.96 (0.85, 1.1)	0.97 (0.91, 1.0)
Mutual adjustment for exposures during trimesters 1 and 3	0.89 (0.77, 1.0)	0.96 (0.90, 1.0)
Third Trimester		
Modeled separately	1.2 (1.1, 1.4)	1.1 (0.98, 1.1)
Mutual adjustment for exposures during trimesters 2 and 3	1.3 (1.1, 1.5)	1.1 (0.98, 1.2)
Preconception		
Modeled separately	0.92 (0.82, 1.1)	1.0 (0.97, 1.1)
Mutual adjustment for exposures during pregnancy and the year post birth	1.0 (0.87, 1.2)	1.0 (0.92, 1.1)
Entire pregnancy		
Modeled separately	1.1 (0.85, 1.3)	1.0 (0.88, 1.2)
Mutual adjustment for exposures during preconception and the year post birth	1.4 (1.0, 1.8)	0.95 (0.77, 1.2)
First year of life		
Modeled separately	0.79 (0.60, 1.0)	1.3 (1.0, 1.6)
Mutual adjustment for exposures during preconception and pregnancy	0.62 (0.41, 0.94)	1.3 (0.98, 1.7)

Effect estimates are scaled to the interquartile range width value for the entire pregnancy period, averaged across study sites  $(1.6-\mu g/m^3 \text{ for PM}_{2.5} \text{ and } 6.6\text{-ppb for ozone})$ .

ASD indicates autism spectrum disorder; CI, confidence interval; OR, odds ratio; PM2.5, particulate matter <2.5 µm.

<sup>a</sup>All models are adjusted for maternal age, maternal education, maternal race/ethnicity, maternal smoking, study site, month of birth, and year of birth.

#### Table 4.

Adjusted<sup>a</sup> Odds Ratios and 95% Confidence Intervals for the Associations between PM<sub>2.5</sub> and Ozone Exposure and ASD Severity.

	Low/moderate Severity OR (95%CI) <sup>b</sup> (n=401)	High Severity OR (95%CI) <sup>C</sup> (n=272)
Ozone		
Preconception	0.93 (0.81, 1.1)	0.91 (0.77, 1.1)
Entire pregnancy	0.99 (0.77, 1.3)	1.2 (0.88, 1.6)
First Trimester	0.91 (0.78, 1.1)	0.98 (0.83, 1.2)
Second Trimester	0.95 (0.82, 1.1)	0.98 (0.83, 1.2)
Third Trimester	1.2 (0.99, 1.4)	1.3 (1.1, 1.6)
First year of life	0.77 (0.55, 1.1)	0.82 (0.57, 1.2)
PM <sub>2.5</sub>		
Preconception	1.0 (0.94, 1.1)	1.1 (0.98, 1.2)
Entire pregnancy	1.1 (0.92, 1.3)	0.92 (0.76, 1.1)
First Trimester	1.0 (0.96, 1.1)	0.90 (0.82, 1.0)
Second Trimester	0.97 (0.90, 1.1)	0.97 (0.88, 1.1)
Third Trimester	1.0 (0.95, 1.1)	1.1 (0.98, 1.2)
First year of life	1.4 (1.1, 1.8)	1.1 (0.82, 1.5)

Effect estimates are scaled to the IQRW value for the entire pregnancy period, averaged across study sites  $(1.6-\mu g/m^3 \text{ for PM}_{2.5} \text{ and } 6.6\text{-ppb for ozone})$ .

ASD indicates autism spectrum disorder; CI, confidence interval; OR, odds ratio; PM2.5, particulate matter <2.5 µm.

<sup>a</sup>Models are adjusted for study site, maternal age, maternal education, maternal race/ethnicity, maternal smoking, month of birth, and year of birth.

<sup>b</sup>Low/moderate severity includes ASD cases with an ADOS Calibrated Severity Score of 4-7.

<sup>c</sup>High severity includes ASD cases with an ADOS Calibrated Severity Score of 8-10.