# Prevalence and associated demographic characteristics of exposure to multiple metals and their species in human populations: the United States NHANES, 2007-2012 

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#### Abstract

Lead $(\mathrm{Pb})$, cadmium $(\mathrm{Cd})$, mercury $(\mathrm{Hg})$, and arsenic (As) are among the top 10 pollutants of global public health concern. Studies have shown that exposures to these metals produce severe adverse effects. However the mechanisms underlying these effects, particularly joint toxicities, are poorly understood in humans. The objective of this investigation was to identify and characterize prevalent combinations of these metals and their species in the U.S. National Health and Nutrition Examination Survey (NHANES) population to provide background data for future studies of potential metal interactions. Exposure to each metal was defined as urine or blood levels $\geq$ respective medians of the NHANES 2007-2012 participants $\Varangle 6$ years ( $n=7408$ ). Adjusted-odds ratios (adj-OR) and 95\% confidence intervals (CI) were determined for demographic factors (age, gender, and race/ethnicity). All models included two additional covariates, cotinine and body mass index (BMI). Species-specific analysis was also conducted for As and Hg including iAs (urinary arsenous acid and/or arsenic acid), met-iAs (urinary monomethylarsonic acid and/or dimethylarsinic acid), and oHg (blood methyl-mercury and/or ethyl-mercury). For combinations of As and Hg species, age- and gender-specific prevalence was determined among NHANES 20112012 participants ( $\mathrm{n}=2342$ ). Approximately $49.3 \%$ of the U.S. population were exposed to a combination of three or more metals. The most prevalent unique specific combinations were $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}, \mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} / \mathrm{Cd}$. Age was consistently associated with these combinations: adj-ORs ranged from $10.9(\mathrm{~Pb} / \mathrm{Cd})$ to $11.2(\mathrm{~Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As})$. Race/ethnicity was significant for $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$. Among women of reproductive age, frequency of oHg/iAs/met-iAS and oHg/metiAs was 22.9 and $40.3 \%$, respectively. These findings of metals and their species in human population may help prioritize efforts to assess joint toxicities and their impact on public health.


## Keywords

Prevalence of multiple metal exposure; exposure patterns; NHANES

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## Introduction

Lead $(\mathrm{Pb})$, cadmium $(\mathrm{Cd})$, mercury $(\mathrm{Hg})$, and the metalloid, arsenic (As), which are referred to as metals from here on, are 4 metals among the top 10 chemicals of major public health global concern [World Health Organization (WHO 2010)]. These metals contaminate the air, surface waters, and soil through natural and human actions (Jarup 2003; Scheckel et al, 2013; Kah et al, 2012; Bradham et al, 2016; Diamond et al, 2016; Tagne-Fotso et al, 2016) and may enter the human body singly or in combinations from the same or different sources. These ubiquitous and persistent metals exert some common mechanisms of toxicity, such as production of oxidative stress (Jomova and Valko 2011; Bridges and Zalups, 2017), and release of stress proteins and interference with the functions of essential metals (Goyer 1997; Davidson et al. 2015). Exposures to multiple metals demonstrated enhanced toxicities compared to single metals, even at low dose levels (Wang and Fowler 2008). In humans, adverse effects of the individual metals have been widely studied (Sweet and Zelikoff, 2001; Ratcliffe et al 1996, Bernstam and Nriagu 2000; ATSDR 2015b), however their potential effects in combinations and associated effect-dose levels have not been well examined, particularly in higher combinations. Such studies are resource and time intensive because the available data are sparse.

One of the challenging issues in investigating potential adverse health effects of multiple chemical exposure are the uncertainties involved. In order to understand potential health impacts from multiple metals and associated dose levels, one needs to account for many parameters, including routes and durations of exposure, bioavailability, pharmacokinetics, metabolism, and half-lives. Given the complexities in calculating internal doses, for future research, it is necessary to prioritize metal combinations that are prevalent in human populations [National Institute of Environmental Health Sciences (NIEHS) 2012].

Exposures to individual chemicals in the non-institutionalized U.S. civil population have been characterized, mainly using data from the biomonitoring program, the National Health and Nutrition Examination Survey (NHANES). This program includes limited assessment of the sources or routes of exposure, but identifies individual, rather than combinations of, environmental chemicals of significance in blood and urine specimens [Centers for Disease Control and Prevention (CDC) 2009]. Recent studies showed that the NHANES data may also be used for further research on co-exposure to multiple chemicals in the U.S. population (Qian et al. 2015; Sobus et al. 2015; Woodruff et al. 2011). The objective of this investigation was to examine co-exposure to multiple metals and their species in the NHANES population to provide background data for future studies of potential metal interactions. The specific aims of our study were to (1) estimate the prevalence of all possible unique combinations of the 4 metals $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$, and As detected at or above the population median levels in blood and urine specimens of the 2007-2012 NHANES participants and identify most prevalent combinations, (2) identify the population profiles (i.e., age, gender, race), and (3) determine unique species combinations of blood organic- Hg and urinary inorganic-related As species.

## Methods and Materials

## Study population and Metal analytes

Briefly, NHANES is a nationally representative cross-sectional survey conducted by the National Center for Health Statistics, CDC. The survey aims to assess the health and nutritional status of the civilian, non-institutionalized U.S. population using a complex multi-stage probability sample design (NHANES 2015). NHANES collects information through a combination of interviews and physical examinations. Blood and urine specimens are collected for lab tests.

Various environmental chemical analytes were determined in subsamples of NHANES participants. Blood metals were measured in participants aged one year and older, whereas urine metals were determined only in subsamples of participants aged 6 years and older. Participants in three NHANES cycles (2007-2008-2009-2010, and 2011-2012) who were tested for both blood and urinary $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$ as well as urinary As were included. NHANES measured As in urine but not in blood, because of its short half-life in blood. The cycles were combined to improve statistical power and obtain more reliable estimates.

Data on blood As and Hg species were available for only the 2011-2012 cycle. Urinary As species included arsenous acid, arsenic acid, monomethylarsenic acid (MMA), and dimethylarsinic acid (DMA); blood Hg species included methyl- Hg and ethyl- Hg . The lab methods used to measure the 4 metals ( $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$, and As ) and As and Hg species in blood and urine specimens have been described in NHANES (2015). NHANES urine samples are 'spot' samples that are collected one time at the time of physical examination (NHANES 2015; Pleil and Sobus, 2016). To adjust for variations in urinary dilution associated with changes in hydration status regression-based creatinine adjustment was performed for all urinary metal levels (Cole et al. 1995; Thompson et al. 1990).

## Statistical analysis

All statistical analyses were conducted using SAS Version 9.3 survey procedures (SAS Institute Inc., Cary, NC, 2010) and SAS-callable SUDAAN Version 11.0 (Research Triangle Park, NC). SUDAAN's Descript procedure was used for estimation of percentiles for subdomains and for estimation of geometric means (GM) and confidence intervals (CI). SURVEYFREQ, SURVEYREG, and SURVEYLOGISTIC procedures were used for weighted \% and means, linear regressions, and logistic regressions, respectively. NHANES assigned each participant a sample weight to account for his or her probability of selection as well as for non-response (Johnson et al. 2013). Subsample weights, modified for combined cycles, and design variables were used to account for NHANES's complex sample design. Taylor series linearization methods were used for variance estimation.

For all analytes, the sample \% participants was presented with a concentration above the limit of detection (LOD). For each analyte with a detection rate above $50 \%$, median and GM concentrations were also calculated with their respective $95 \%$ CI. The medians were calculated using data from all 3 cycles, except for blood methyl- Hg for which only the 2011-2012 data were available. When the concentration of an analyte was below the LOD, a value assigned by NHANES (i.e., LOD divided by the square root of 2 ) was employed.

Our main analyses included two parts. The first part involved the 4 metals and the second part was an advances analysis of the As and Hg species data. In the first part of the analysis that included 2007-2012 NHANES participants, a dichotomous variable was created separately for each of the urinary total As, blood and urinary Pb , blood and urinary Cd , and blood and urinary total Hg , indicating whether a person's metal concentrations were greater than or equal to the respective medians of the entire study population. All possible unique combinations of the 4 metals (i.e., None, As, $\mathrm{Cd}, \mathrm{Hg}, \mathrm{Pb}, \mathrm{As} / \mathrm{Cd}, \mathrm{As} / \mathrm{Hg}, \mathrm{As} / \mathrm{Pb}, \mathrm{Cd} / \mathrm{Hg}$, $\mathrm{Cd} / \mathrm{Pb}, \mathrm{Hg} / \mathrm{Pb}, \mathrm{As} / \mathrm{Cd} / \mathrm{Hg}, \mathrm{As} / \mathrm{Cd} / \mathrm{Pb}, \mathrm{As} / \mathrm{Hg} / \mathrm{Pb}, \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$, and $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$ ) were identified, where each metal concentration was at or above the population median in urine or blood, or both. Then the prevalence for each of the unique combinations of the metals was calculated. In addition, as a summary measure, the set of two metals that are most common across the unique combinations was also identified. For example, we estimated the prevalence of the two-metal set, As and Cd, was estimated by summing the prevalence estimates of all unique combinations containing As and Cd (i.e., $\mathrm{As} / \mathrm{Cd}, \mathrm{As} / \mathrm{Cd} / \mathrm{Hg}$, $\mathrm{As} / \mathrm{Cd} / \mathrm{Pb}$, and $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$ ). In the same manner, the prevalence of the two-metal set, As and Hg , was determined by summing the prevalence estimates of $\mathrm{As} / \mathrm{Hg}, \mathrm{As} / \mathrm{Cd} / \mathrm{Hg}$, $\mathrm{As} / \mathrm{Hg} / \mathrm{Pb}$, and $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$. Similarly, the set of three metals that are most commonly detected together across the unique combinations was identified.

To examine the demographic factors associated with the three most prevalent unique combinations, adjusted odds ratios (adj-OR) and their 95\% CIs were calculated by employing logistic regression analyses where the odds of having the combination was regressed on the covariates. The controls consisted of NHANES participants who had no or only one of the 4 metals detected at or above the population median levels. The covariates of interest included age (in 10 year increments), gender, and race/ethnicity (MexicanAmericans, Non-Hispanic Whites, Non-Hispanic Blacks, and Others). Serum cotinine concentration ( $<0.015,0.015-9.999, \geq 10 \mathrm{ng} / \mathrm{mL}$ ) and body mass index (BMI) were included in all models as potential confounders. These two potential confounders were selected because of the availability of objective quantifiable measurements and their known associations with both the metals and demographic variables (Chiba and Masironi 1992; CDC 2016; Padilla et al 2010). Serum cotinine has been established as a valid biomarker for active tobacco smoking ( $\geq 10 \mathrm{ng} / \mathrm{mL}$ ) (Pirkle et al. 1996) and secondhand tobacco smoking ( $<10 \mathrm{ng} / \mathrm{mL}$ ) (Benowitz 1996; Pirkle et al. 1996). Non-exposure to tobacco smoke was defined as $<$ LOD (i.e., $<0.015 \mathrm{ng} / \mathrm{mL}$ ). For participants 21 years and older, BMI was calculated as weight in kg divided by height-squared in $\mathrm{m}^{2}$. For participants younger than 21 years, CDC BMI-for-Age Growth Chart coefficients were used for BMI calculation (CDC 2002). BMI was categorized into underweight, normal, overweight, and obese. For participants 21 years and older, the corresponding cut points were $<18.5,18.5-24.9$, and 2529.9 , and $\geq 30$. For participants younger than 21 years, the corresponding cut points were the $<5^{\text {th }}, 5^{\text {th }}-<85^{\text {th }}, 85^{\text {th }}-<95^{\text {th }}$, and $\geq 95^{\text {th }}$ percentiles obtained from the age- and gender specific growth charts.

For the second part of the analysis that included only the 2011-2012 cycle participants, a dichotomous variable was created for each of the urinary inorganic-related As species (arsenous acid, arsenic acid, MMA, and DMA) and blood oHg species (methyl-Hg and ethyl -Hg ), indicating whether an individual's concentration was greater than or equal to the
respective LOD values. LOD values were selected for cut points, because stable estimates of the population medians could not be obtained for several species (e.g., urinary arsenous acid and arsenic acid) due to a small number of participants with a level $\geq$ LOD as shown in Table 1. The As species was grouped into iAs (urinary arsenous acid and arsenic acid) and met-iAS (urinary MMA and DMA). The Hg species consisted of oHg (blood methyl- Hg and ethyl- Hg ). Unique combinations of iAs , met-iAs, and oHg were presented, where each species group had at least one species detected at or above the respective LOD level. The gender- and age-prevalence estimates were obtained by categorizing the ages into 3 groups ( $6-14,15-44$, and $\geq 45$ years). Women of the reproductive age group ( $15-44$ years) were considered a vulnerable subgroup, since these metals cross the placental barrier (Baranowska 1995; Rudge et al. 2009; Piasek et al, 2016) and can also be detected in breast milk (Ettinger et al. 2014; Garcia-Esquinas et al. 2011).

## Results

## Prevalence of $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$, and As combinations

Most NHANES participants 6 years and older in 2007-2012 exhibited detectable levels of the 4 metals in urine ( $85.8-98.5 \%$ ) and blood specimens ( $69.8-99.5 \%$ ) (Table 1). The weighted creatinine-adjusted median values for $\mathrm{As}, \mathrm{Cd}, \mathrm{Hg}$, and Pb in urine specimens were $7.91 \mu \mathrm{~g} / \mathrm{L}, 0.2 \mu \mathrm{~g} / \mathrm{L}, 0.44 \mu \mathrm{~g} / \mathrm{L}$, and $0.5 \mu \mathrm{~g} / \mathrm{L}$, respectively. The weighted median values in blood specimens were $0.27 \mu \mathrm{~g} / \mathrm{L}$ for $\mathrm{Cd}, 0.76 \mu \mathrm{~g} / \mathrm{L}$ for Hg , and $1.07 \mu \mathrm{~g} / \mathrm{dL}$ for Pb (Table 1).

In $8.4 \%$ of the U.S. population 6 years and older, none of the 4 metals were detected in urine or blood at or above the respective population medians. Among the remaining participants, $16.2 \%$ exhibited only a single metal, $26.1 \%$ two metals, and $49.3 \% 3$ or 4 metals (Table 2). The most common unique combination among all 16 possible combinations of 4 metals was the quaternary combination of $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}(22.1 \%)$, followed by the ternary combination of $\mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}(10.6 \%)$ and binary combination of $\mathrm{Cd} / \mathrm{Pb}(8.4 \%)$ (Table 2). To summarize the findings, the set of two metals most commonly detected was also examined together across the unique combinations. In $45.5 \%$ of the population, Cd and Pb were detected together in various unique forms [i.e., $8.4(\mathrm{Cd} / \mathrm{Pb})+4.4(\mathrm{As} / \mathrm{Cd} / \mathrm{Pb})+10.6(\mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb})$ $+22.1(\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb})$ ], while only $8.4 \%$ of the population displayed the unique $\mathrm{Cd} / \mathrm{Pb}$ combination. The set of three metals most commonly detected together was $\mathrm{Cd}, \mathrm{Hg}$, and Pb ; $32.7 \%$ of the population had the three metals in various unique forms [i.e., $10.6(\mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb})$ $+22.1(\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb})]$ versus $10.6 \%$ had the unique $\mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$ combination.

## Demographic characteristics of the most prevalent metal combinations

The participants who had one of the three most prevalent unique combinations $(\mathrm{Pb} / \mathrm{Cd}$, $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ ) were older (weighted GM age: 51.6, 52.8 , and 53.7 years for $\mathrm{Pb} / \mathrm{Cd}, \mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$, respectively) than the control group which consisted of participants who did not exhibit any or only one metal at or above the respective median concentration (weighted GM age: 22.9 years) (Table 3). These individuals were also more likely to be an active tobacco smoker as indicated by a serum cotinine level $\geq 10 \mathrm{ng} / \mathrm{mL}$ (weighted \%: 40.1, 37.6, and $23.8 \%$ for $\mathrm{Pb} / \mathrm{Cd}, \mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$, respectively) than controls (weighted \%: 9.3\%). In the multivariate logistic regression analyses, age was
the most consistent independent risk factor among the demographic variables (age, gender, and race/ethnicity), for all three unique combinations (Figure 1). The adj-ORs for age in 10 year increments were 10.9 for $\mathrm{Pb} / \mathrm{Cd}, 11.1$ for $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and 11.2 for $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$, after taking into account the effects of gender, race/ethnicity, cotinine, and BMI. Race/ethnicity was a significant risk factor for the $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ combination. Compared to Non-Hispanic Whites, the odds of displaying the quaternary combination was 1.8 -fold higher for NonHispanic Blacks, 2.6-fold for Mexican Americans, and 5.6-fold for "Other" race/ethnicity group that includes Asian ethnicity, after adjusting for the effects of all other covariates. Non-Hispanic Blacks also presented a significantly increased odds of having the $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$ combination, while Mexican-Americans a significantly increased odds of the $\mathrm{Pb} / \mathrm{Cd}$ combination, compared to non-Hispanic Whites. Males were less likely to display the $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$ combination, but there was no significant gender effect for the other two prevalent combinations.

## Prevalence of oHg, iAs, and met-iAs combinations

Our analysis of the species combinations included the 2011-2012 participants (sample $\mathrm{n}=2342$ ). The most prevalent combination of the species at a detectable level was the binary combination of oHg/met-iAs ( $42.1 \%$ ), followed by the ternary combination of $\mathrm{oHg} / \mathrm{met}-$ iAs/iAs (24.7\%) (data not shown). The prevalence was lower for the binary combinations of iAs/met-iAs (3.4\%) and oHg/iAs ( $0.5 \%$ ) (data not shown). The age- and gender-specific prevalence of unique combinations of oHg , met-iAs, and iAs is illustrated in Figure 2. Among males, the ternary combination of $\mathrm{oHg} /$ met-iAs/iAs was detectable at or above the LOD levels in $22.6 \%$ of the 6-14 years group (sample $\mathrm{n}=239$ ), reaching a peak ( $38.4 \%$ ) in the 15-44 years group (sample $\mathrm{n}=482$ ), then declined to $22.9 \%$ in the 45 years or older group (sample $\mathrm{n}=450$ ). In females, the prevalence of $\mathrm{oHg} /$ met-iAs/iAs peaked ( $25.7 \%$ ) in the $6-14$ years group (sample $n=242$ ), followed by $22.9 \%$ in the reproductive age group (15-44 years, sample $n=477$ ), and $15.3 \%$ in the 45 years or older group (sample $n=452$ ). In contrast, the prevalence of the binary combination of $\mathrm{oHg} /$ met-iAs rose with age in both females ( $26.0,40.3$, and $53 . \%$ for $6-14,15-44$, and $>44$ years, respectively) and males ( $24.7,34.2$, and $50 . \%$, for $6-14,15-44$, and $>44$ years, respectively).

## Discussion

Individuals may be exposed to metals singly or in combination simultaneously, depending upon their environment. Subjects may also be exposed sequentially to the individual metals. Our study found that an estimated $49.3 \%$ of the U.S. population 6 years and older in 20072012 had a ternary or quaternary combination of $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$, and As, detected in their blood or urine at or above the population median concentrations. These 4 metals are often examined individually by biomonitoring programs in many countries because of their wellrecognized biological and physicochemical properties; however, their joint exposure profiles have not been widely monitored. Our study identified that the most prevalent unique combination of the 4 metals in the U.S. population was $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}(22.1 \%)$, followed by $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}(10.6 \%)$ and $\mathrm{Pb} / \mathrm{Cd}(8.4 \%)$.

Our findings provide baseline information for prioritizing efforts in assessing potential adverse health effects and the effect-doses of the metal combinations. The idea of prioritizing chemicals and their mixtures to reduce a complex problem to a more manageable one has been used effectively for over two decades. These efforts resulted in identifying mixtures of concern based upon frequency of occurrences in various environmental media and in completed exposure pathways (Fay and Mumtaz 1996; Toccalino et al. 2012). The identified mixtures were employed for experimental toxicity testing to fill data gaps. In the current analysis data on combinations of 4 metals commonly detected are provided for in the U.S. population.

These 4 metals of major public health concern (WHO 2010) are also among the top 7 hazardous substances that pose human health threats at Superfund sites in the United States [Agency for Toxic Substances and Disease Registry (ATSDR) 2015a]. Although data are limited regarding their combination effects in humans, some investigators suggested complex interactions between these metals. de Burbure et al. (2006) noted Hg co-exposure exacerbated the effect of Cd on urinary homovanillic acid in children, one of the endproducts of dopamine metabolism, whereas Pb co-exposure appeared to antagonize the effect of Hg on the dopaminergic marker. In a study of 6-11 years old U.S. girls who participated in NHANES III, higher blood Pb was associated with lower levels of inhibin B, a marker of follicular development, with the strongest effect seen in the presence of high concentrations of both blood Pb and urinary Cd (Gollenberg et al. 2010).

Assessing health effects of exposure to chemical combinations is a challenging area, hampered by a paucity of relevant data as well as the complexities involved. Often multiple mechanisms act with such exposures, sometimes simultaneously or sequentially in the overall expression of toxicity. The nature and severity of toxicity may change as a function of combinations of chemicals and/or their ratios (e.g., chemical/chemical and chemical/ metabolites) (ATSDR 2015b). Similarly, variability in toxicity exists between individuals due to polymorphisms and differences in metabolism (ATSDR 2015b). Most available information of human toxicity for chemical combinations has been derived from animal experimental studies and many of these investigations were not designed to actually quantify chemical interactions that contribute to greater than expected joint toxicity of a given mixture (ATSDR 2004). To achieve this, a weight of evidence methodology based on pairwise comparison was developed (Mumtaz and Durkin 1992; ATSDR 2004). The binary studies have been particularly useful in understanding both mechanisms of toxicity and interactions as a function of dose and chemical combinations that play a critical role in toxicity of multiple chemicals.

Our study identified unique ternary and quaternary combinations of the 4 metals, and the binary and ternary combinations of their species. These results may be utilized in designing future experimental toxicity studies that might enable one to move away from dependence on pair-wise comparison methodology.

Our observations also confirmed age as the strongest independent demographic factor associated with all three most common combinations (i.e., $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}, \mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} /$ Cd ). The strength of associations was similar between the three combinations (adj-ORs,
10.9-11.1 for each 10 year increment). Previous investigators reported nonlinear relationships between age and individual metals, including a quadratic relationship with gender-adjusted blood Hg concentrations among NHANES participants (Caldwell et al. 2009). A general population study in Belgium reported similar nonlinear and nonmonotonic relationships between age and urinary Cd concentrations (Chaumont et al. 2013). Ruiz et al (2010) suggested that life stage-related anatomical and physiological changes may play an important role in the absorption and excretion of Cd. Our observations on the race/ethnicity differences in the odds of having metal combinations may be partly explained by differences in exposure, such as diet (deCastro et al. 2014; Martorell et al. 2011). For the $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ combination, subgroup analyses by age suggested potential interaction effects between age and gender, as well as between age and race/ethnicity, although the interaction terms were not significant (data not shown). If the interaction effect is confirmed, this information might help to identify specific subgroups for further assessment of exposure sources and potential adverse health effects.

Our study is the first to report the prevalence of As and Hg species combinations in the U.S. general population. Our results showed that $22.9 \%$ of females in the reproductive age group (15-44 years) exhibited the combination of $\mathrm{oHg} / \mathrm{met}-\mathrm{iAs} / \mathrm{iAs}$ at detectable levels in 20112012. The prevalence was even higher ( $25.7 \%$ ) in younger females (6-14 years). These metals may be transferred from mother to fetus (Baranowska 1995; Rudge et al. 2009). Hence, further studies are needed to understand the potential adverse health effects and effect-dose relationships between mother and infant (Ettinger et al. 2014; Garcia-Esquinas et al. 2011). In the U.S. population, iAs was frequently detected together with met-iAs (i.e., MMA and/or DMA), as a binary (met-iAs/As, 3.4\%) or ternary (oHg/met-iAs/iAs) combination ( $24.7 \%$ ), but was rarely detected alone $(0.1 \%)$ or as $\mathrm{oHg} / \mathrm{iAs}(0.5 \%)$. In contrast, met-iAs occurred alone ( $6.6 \%$ ) more frequently or as $\mathrm{oHg} /$ met-iAs $(42.1 \%)$ without detectable iAs. iAs is rapidly metabolized to MMA and DMA in the body, which may partly explain our findings. It is also possible that the presence of met-iAS without detectable iAs may be partly explained by direct intake of met-iAs from exogenous sources as described previously (Aylward et al. 2014; deCastro et al. 2014).

Our study has several strengths and limitations. A large number of participants ( $\mathrm{n}=7408$ ) was included by combining three NHANES cycles, affording stable estimates of the prevalence. To our knowledge, this is the first study to report the age- and gender-specific prevalence of exposure to As and Hg species combinations in the U.S. general population. Our study is, however, limited by the small number of chemicals considered. Development of new statistical methods is urgently needed to simultaneously examine a large number of chemicals and non-chemical agents (e.g., radiation) and to interpret the results in a biologically meaningful way. It was not possible to examine age-cohort effect because of the limited number of age-cohort combinations in the data used. Further, other potential confounders (e.g., diet and occupation) were not included in examining demographic factors associated with prevalent combinations. Future studies examining exposure risk factors need to consider using most reliable methods for exposure measurements, such as long-term average daily intake for dietary factors. The population median concentrations were employed as the cut points to provide baseline information for the population. Future
investigations assessing potential adverse health effects of metal combinations need to take steps to establish exposure levels that are sufficient to cause health concerns.

## Conclusions

Data for metals ( $\mathrm{Pb}, \mathrm{Cd}, \mathrm{Hg}$, and As) exposure reported by NHANES 2007-2012 were analyzed. Approximately in $50 \%$ of the population 6 years or older a ternary or quaternary combination of the 4 metals was detected in their blood or urine specimens at or above the respective population median levels. The most prevalent unique combination was $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$, followed by $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$, and $\mathrm{Pb} / \mathrm{Cd}$. Age was confirmed as the most consistent independent demographic factor associated with exposure to these combinations, after adjusting for the effects of other demographic variables (gender and race/ethnicity) and two selected confounders (serum cotinine and BMI). Compared to Non-Hispanic Whites, all other race/ethnicity groups were associated with a significantly increased risk of $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ combination.

The prevalence of inorganic-related As and organic Hg species combinations is noteworthy. In addition, $63 \%$ of females of reproductive age ( $15-44$ years) were exposed to the specific unique combination of $\mathrm{oHg} / \mathrm{iAs} /$ met-iAs $(23 \%)$ or $\mathrm{oHg} / \mathrm{met}-\mathrm{iAs}(40 \%)$. The prevalence of oHg/iAs/met-iAs was higher in younger females aged 6-14 years (25.7\%) than other age groups. Our findings fill a data gap by providing baseline data on exposure patterns of multiple metal combinations and their species that demonstrated high relevance to public health in the U.S. population. These baseline data might help (1) prioritize efforts in assessing potential adverse health effects, (2) establish exposure levels that are sufficient to cause health concerns, (3) identify exposure sources, and (4) implement preventive measures.

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| Variable | (Pb/Cd) <br> Adjusted OR (95\%Cl) |  | ( $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$ ) <br> Adjusted OR (95\%Cl) |  | ( $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ ) <br> Adjusted OR (95\%Cl) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Gender |  |  |  |  |  |  |
| Male | 1.0 ( 0.6-1.5) | 4 | 0.7 (0.5-0.9) | (1) | 0.8(0.6-1.1) | " |
| Female | Reference |  | Reference |  | Reference |  |
| Age (10 years) | 10.9 (10.8-11.1) | ¢ | 11.1 (11.0-11.3) | ¢ | 11.2 (11.1-11.4) | ¢ |
| Race/ethnicity |  |  |  |  |  |  |
| Non-Hispanic Blacks | 1.5 (0.9-2.6) | 1 | 1.8 (1.1-3.0) | . | 1.8(1.1-2.9) | d |
| Mexican-Americans | 2.3 (1.1-4.6) | H | 1.0 (0.5-2.0) | d | 2.6 (1.6-4.3) | H |
| Others | 1.2 (0.6-2.5) | (1) | 1.6 (0.9-2.8) | 4 | 5.6 ( 3.6-8.5) | \| ${ }^{\prime \prime}$ |
| Non-Hispanic Whites | Reference |  | Reference |  | Reference |  |
| Serum Cotinine ( $\mathrm{ng} / \mathrm{mL}$ ) $\geq 10$ <br> 0.015-9.999 $<0.015$ | $\begin{gathered} 19.6(11.3-34.0) \\ 2.4(1.5-3.7) \\ \text { Reference } \end{gathered}$ |  | $\begin{gathered} 14.7(7.5-28.7) \\ 2.0(1.1-3.5) \\ \text { Reference } \end{gathered}$ | $\longmapsto=1$ | $\begin{gathered} 6.6(4.4-10.0) \\ 1.7(1.2-2.4) \\ \text { Reference } \end{gathered}$ | $\mathrm{lm}$ |
| BMI |  |  |  |  |  |  |
| Obese | 0.4 (0.3-0.7) | 中 | 0.4 (0.2-0.6) | 4 | 0.3 (0.2-0.4) | + |
| Overweight | 0.7 (0.4-1.3) | 4 | 0.9 (0.6-1.4) | 4 | 0.6 (0.4-0.8) | 4 |
| Normal/Below | Reference |  | Reference |  | Reference |  |
|  |  | $\begin{array}{llll} 0 & 10 & 20 & 30 \end{array}$ |  | $\begin{array}{llll} 0 & 10 & 20 & 30 \end{array}$ |  | $\begin{array}{llll} 0 & 10 & 20 & 30 \end{array}$ |

Figure 1.
Demographic factors associated with most prevalent unique combinations of $\mathrm{As}, \mathrm{Cd}, \mathrm{Pb}$, and Hg among U.S. population 6 years and older, NHANES 2007-2012.
Abbreviation: As, arsenic; Cd, cadmium; Pb, lead; Hg, mercury; OR, odds ratio; 95\% CI, $95 \%$ confidential interval; SE, standard error; BMI, Body mass index. Age in 10 year increments. Serum cotinine levels $\geq 10 \mathrm{ng} / \mathrm{mL}$ indicate active tobacco smoking (Pirkle et al. 1996) and $<10 \mathrm{ng} / \mathrm{mL}$ indicate second hand tobacco smoking (Benowitz 1996; Pirkle et al. 1996). We defined non-exposure to tobacco smoke as <LOD (i.e., $0.015 \mathrm{ng} / \mathrm{mL}$ ).


Figure 2.
Unique combinations of blood organic Hg and urinary inorganic-related As species measured at or above the respective detection limits among U.S. population 6 years and older: NHANES 2011-2012.
iAs, urinary arsenous acid and/or arsenic acid; met-iAs, urinary monomethylarsonic acid and/or dimethylarsinic acid; oHg, blood methyl-Hg and/or ethyl-Hg.
Median and geometric mean concentrations of $\mathrm{As}, \mathrm{Cd}, \mathrm{Pb}$, and Hg in urine and blood for the U.S. population 6 years and older, NHANES 2007-2012.

| Chemical analytes | LOD level ${ }^{\boldsymbol{a}}$ |  | Sample N | Detection rate Weighted \% | Weighted Concentration ${ }^{\text {b }}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2007-2010 | 2011-2012 |  |  | GM (SE) | Median (95\% CI) |
| Urine ${ }^{\text {c }}$ ( $\mu \mathrm{g} / \mathrm{L}$ ) |  |  |  |  |  |  |
| As, total | 0.74 | 1.25 | 7969 | 98.5 | 9.78 (0.29) | 7.91 (7.44, 8.44) |
| Arsenous acid | 1.2 | 0.48 | 7964 | 13.0 | - | - |
| Arsenic acid | 1.0 | 0.87 | 7945 | 2.1 | - | - |
| Arsenobetaine | 0.4 | 1.19 | 7963 | 54.2 | 1.80 (0.06) | 1.20 (1.11, 1.28) |
| Arsenocholine | 0.6 | 0.28 | 7964 | 2.6 | - | - |
| Dimethylarsinic acid | 1.7 | 1.8 | 7964 | 82.2 | 4.27 (0.07) | 3.92 (3.80, 4.07) |
| Monomethylarsonic acid | 0.9 | 0.89 | 7964 | 31.2 | - | - |
| Trimethylarsine oxide | 1.0 | 0.25 | 7964 | 1.3 | - | - |
| Cd | 0.042 | 0.56 | 7979 | 85.8 | 0.21 (0.00) | 0.20 (0.19, 0.21) |
| Hg | 0.08 | 0.05 | 8006 | 94.3 | 0.47 (0.01) | 0.44 (0.42, 0.47) |
| Pb | 0.1 | 0.08 | 7979 | 97.1 | 0.52 (0.01) | 0.50 (0.49, 0.52) |

Blood ( $\mu \mathrm{g} / \mathrm{L}$ )

| Cd | 0.2 | 0.16 | 7603 | 69.8 | $0.31(0.00)$ | $0.27(0.26,0.28)$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Hg, total | 0.33 | 0.16 | 7603 | 85.2 | $0.81(0.03)$ | $0.76(0.72,0.81)$ |
| Inorganic | 0.35 | 0.27 | 7586 | 19.1 | - | - |
| Ethyl $^{d}$ | - | 0.16 | 2376 | 3.6 | - | - |
| Methyl $^{d}$ | - | 0.12 | 2376 | 84.9 | $0.50(0.03)$ | $0.49(0.42,0.57)$ |
| $\mathrm{Pb}(\mu \mathrm{g} / \mathrm{dL})$ | 0.25 | 0.25 | 7603 | 99.5 | $1.10(0.02)$ | $1.07(1.03,1.11)$ |

Abbreviations: As, arsenic; Cd , cadmium; Pb , lead; Hg , mercury; GM , geometric mean; SE , standard error; $95 \% \mathrm{CI}, 95 \%$ confidence interval.
${ }^{\text {a }}$ NHANES survey cycles 2007-2008 and 2009-2010 had the same LOD levels.
${ }^{b}$ Weighted geometric mean and median concentrations were calculated for each analyte with a detection rate above $50 \%$. When the concentration of an analyte was below the limit-of-detection (LOD), we used a value assigned by NHANES (i.e., LOD divided by the square root of 2 ).
${ }^{c}$ All urinary concentrations were adjusted per gram of urine creatinine using regression methods (Thompson et al. 1990).

Table 2
Specific unique combinations of $\mathrm{As}, \mathrm{Cd}, \mathrm{Pb}$, and Hg detected at or above the respective median concentrations in urine or blood among the U.S. population 6 years and older, NHANES 2007-2012.

|  |  |  |
| :--- | :---: | :---: |
| Metal combination ${ }^{\boldsymbol{a}}$ | Sample $\mathbf{N}^{\boldsymbol{b}}$ | ${\text { Prevalence }{ }^{\boldsymbol{c}}}_{$$}$ |
| None | 590 | $8.4(7.0-9.7)$ |
| As | 206 | $2.7(2.2-3.1)$ |
| Cd | 320 | $4.7(3.9-5.6)$ |
| Pb | 347 | $3.6(3.1-4.2)$ |
| Hg | 333 | $5.2(4.6-5.8)$ |
| $\mathrm{As} / \mathrm{Cd}$ | 119 | $1.4(1.0-1.7)$ |
| $\mathrm{As} / \mathrm{Hg}$ | 369 | $5.0(4.3-5.8)$ |
| $\mathrm{As} / \mathrm{Pb}$ | 236 | $2.2(1.8-2.7)$ |
| $\mathrm{Cd} / \mathrm{Hg}$ | 317 | $5.4(4.6-6.3)$ |
| $\mathrm{Cd} / \mathrm{Pb}$ | 632 | $8.4(7.3-9.5)$ |
| $\mathrm{Pb} / \mathrm{Hg}$ | 294 | $3.7(3.1-4.3)$ |
| $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg}$ | 449 | $6.5(5.5-7.4)$ |
| $\mathrm{As} / \mathrm{Cd} / \mathrm{Pb}$ | 381 | $4.4(3.7-5.1)$ |
| $\mathrm{As} / \mathrm{Hg} / \mathrm{Pb}$ | 448 | $5.7(4.9-6.5)$ |
| $\mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$ | 696 | $10.6(9.3-11.9)$ |
| $\mathrm{As} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{Pb}$ | 1671 | $22.1(20.3-23.9)$ |

Abbreviation: As, arsenic; Cd, cadmium; Pb , lead; Hg, mercury; 95\% CI, $95 \%$ confidence interval.
${ }^{a}$ As and Hg represent total As and total Hg .
$b_{\text {All participants }}(\mathrm{n}=7408)$ were tested for urinary and blood $\mathrm{Cd}, \mathrm{Pb}$, and Hg as well as urinary As.
${ }^{c}$ Detected in blood and/or urine specimens.
Demographic characteristics associated with most prevalent unique combinations of $\mathrm{As}, \mathrm{Cd}, \mathrm{Pb}, \mathrm{and} \mathrm{Hg}$ at or above the respective population medians among U.S. population 6 years and older, NHANES 2007-2012.

| Characteristics | None or one metal | $\mathrm{Pb} / \mathrm{Cd}$ | $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg}$ | $\mathrm{Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) |


|  | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) | $\mathrm{N}^{a}($ Weighted \%) |
| :--- | :--- | :--- | :--- | :--- |
| Gender |  |  |  |  |

$340(55.4) \quad 330(48.9) \quad 813(49.8)$
819 (50.2)
$52.8 \pm 0.70 \quad 53.7 \pm 0.43$
340 (20.8)
479 (29.4)
580 (35.5)
389 (23.8)
863 (52.9)
380 (23.3)
496 (30.4)
574 (35.2)
562 (34.4)
Abbreviation: As, arsenic; Cd , cadmium; Pb , lead; Hg , mercury; GM , geometric mean; SE , standard error; BMI, Body mass index.
${ }^{a}$ Sample N: None or one metal $=1751, \mathrm{~Pb} / \mathrm{Cd}=614, \mathrm{~Pb} / \mathrm{Cd} / \mathrm{Hg}=675, \mathrm{~Pb} / \mathrm{Cd} / \mathrm{Hg} / \mathrm{As}=1632$.
${ }^{b}$ Serum cotinine levels $\geq 10 \mathrm{ng} / \mathrm{mL}$ indicate active tobacco smoking (Pirkle et al. 1996) and $<10 \mathrm{ng} / \mathrm{mL}$ indicate second hand tobacco smoking (Benowitz 1996; Pirkle et al. 1996). Non-exposure to tobacco smoke defined as <LOD (i.e., $0.015 \mathrm{ng} / \mathrm{mL}$ ).


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