



Retrofitting and re-powering as a control strategies for curtailment of exposure of underground miners to diesel aerosols

Aleksandar D. Bugarski¹ · Jon A. Hummer¹ · Shawn Vanderslice¹ · Teresa Barone¹

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Abstract

A study was conducted to examine the potential of diesel emissions control strategies based on retrofitting existing power packages with exhaust aftertreatment devices and repowering with advanced power packages. The retrofit systems, a diesel oxidation catalyst (DOC) and diesel particulate filter (DPF), were evaluated individually using a US EPA tier 2 (ter 2) engine operated under four steady-state conditions and one transient cycle. The DOC effectively curtailed emissions of CO, and to some extent organic carbon (OC), elemental carbon (EC), and aerosol number concentration. The DPF system offered substantially higher reductions in OC and EC mass and aerosol number concentrations. Both, the DOC and DPF achieved reductions in the aforementioned emissions without adversely affecting emissions of NO₂ and nano-sized aerosols. The strategy of repowering with an advanced system was examined using a US EPA tier 4 final (tier 4f) engine equipped with a cooled exhaust gas recirculation system and diesel exhaust fluid-based selective catalytic reduction system, but not with a DPF system. The tier 4f engine contributed substantially less than the tier 2 engine to the EC and OC mass, aerosol number, and CO, NO, and NO₂ concentrations. The tier 4f engine was very effective in reducing aerosol mass, NO, and NO₂ concentrations, but it was not equally effective in reducing aerosol number concentrations. The implementation of viable exhaust after treatment systems and advanced diesel power packages could be instrumental to the underground mining industry to secure a clean, economical, and dependable source of power for mobile equipment.

Keywords Diesel · Exhaust after treatment · Advanced engines · Underground mining

Acronyms

| | | | |
|-----------------|---|-----------------|---|
| ACGIH | American Conference of Governmental Industrial Hygienists | DEF | Diesel exhaust fluid |
| AIOH | Australian Institute of Occupational Hygienist | DPF | Diesel particulate filter |
| ASC | Ammonia slip catalyst | DPM | Diesel particulate matter |
| ASTM | ASTM International, an international standards organization that develops and publishes voluntary consensus technical standards | EC | Elemental carbon |
| CMD | Count median diameter | EGR | Exhaust gas recirculation |
| CO | Carbon monoxide | FMPS | Fast Mobility Particle Sizer |
| CO ₂ | Carbon dioxide | HC | Hydrocarbons |
| DOC | Diesel oxidation catalytic converter | I50 | Intermediate speed 50 percent load (ISO M8) |
| D _{em} | Electrical mobility diameter | I100 | Intermediate speed 100 percent load (ISO M6) |
| | | IARC | International Agency on Research on Cancer |
| | | MSHA | Mine Safety and Health Administration |
| | | N | Number |
| | | NIOSH | National Institute for Occupational Safety and Health |
| | | NO | Nitric oxide |
| | | NO ₂ | Nitrogen dioxides |
| | | NO _x | Nitric oxides (NO _x = NO + NO ₂) |
| | | OC | Organic carbon |
| | | OEM | Original equipment manufacturer |
| | | ON | Ontario |
| | | Pd | Palladium |

✉ Aleksandar D. Bugarski
 abugarski@cdc.gov

¹ Office of Mine Safety and Health Research, National Institute for Occupational Safety and Health, Pittsburgh Mining Research Division, 626 Cochran's Mill Rd, Pittsburgh, PA 15236, USA

| | |
|----------|---|
| PM | Particulate mass |
| PN | Particulate number |
| QFF | Quartz fiber filters |
| R50 | Rated speed 50% load (ISO M3) |
| R100 | Rated speed 100% load (ISO M1) |
| SAE | Society of Automotive Engineers |
| SCR | Selective catalyst reduction |
| SCR-In | Before DOC/SCR/ASC system |
| SCR-Out | After DOC/SCR/ASC system |
| SS | Steady-state |
| TC | Total carbon |
| TD | Thermodenuder |
| TLV | Threshold limit values (ACGIH) |
| TOT-EGA | Thermal optical transmittance-evolve gas analysis |
| TR | Transient |
| ULSD | Ultralow sulfur diesel |
| US EPA | US Environmental Protection Agency |
| σ | Log-normal distribution spread |

1 Introduction

Diesel-powered equipment is extensively used in the underground mining industry [1, 2]. As a result, some underground miners that work in the areas with limited quantities of ventilation air are exposed to diesel aerosols and gases [3–7]. Exposures to diesel exhaust have been linked to various adverse health outcomes including those on the pulmonary system [6, 8–10], bladder [8], cardiovascular system [11, 12], and brain [13]. Nitrogen oxides (especially NO₂), various forms of organic compounds, and nanometer and ultrafine particulates are considered biologically active components of diesel exhaust emitted by traditional and contemporary engines [14–17]. The International Agency on Research on Cancer (IARC) categorized diesel engine exhaust as a carcinogen to humans (group 1) [18, 19].

Occupational exposure to diesel aerosols is limited directly or indirectly in several jurisdictions including the USA [20, 21], European Union [22], and Australia [23]. Exposures of underground miners to particulate matter and gases emitted by diesel engines can be controlled through the implementation of a variety of integrated, multifaceted control strategies—elimination, substitution, engineering controls, administrative controls, and the use of personal protective equipment [24]. In recent years, the substitution of diesel power with electric equipment has been discussed [25]. It appears that in spite of concerted efforts to substitute some heavy- and light-duty diesel-powered vehicles with electric (in particular lithium ion battery)-powered vehicles [25, 26], important limitations remain in electric equipment flexibility, quality of the build, durability, maintenance programs, and other engineering and economic factors. These limitations indicate that diesel equipment will continue to power an important

part of underground mining process for some time. Therefore, engineering control strategies that allow the industry to curtail diesel pollutants at their source, prior to their release into the underground environment, should remain central to efforts to reduce exposures. Potential strategies for exposure reduction include (1) re-powering existing and powering new vehicles with advanced engine technologies, (2) retrofitting existing vehicles with exhaust aftertreatment technologies and crankcase emission controls, and (3) using alternative fuels such as fatty acid methyl ester biodiesel and hydrotreated vegetable oil renewable diesel.

Regulatory frameworks pertinent to the use of diesel engines [27, 28] resulted in the rapid development of engine, aftertreatment, and fuel technologies. Improvements in engine combustion technologies [29–31], exhaust aftertreatment technologies [32–35], and alternative fuels [36–38] have had profound effects on the levels of diesel emissions and the physical and chemical properties of aerosols emitted by diesel engines [32, 39–42]. Those advancements could be instrumental to the underground mining industry's efforts to comply with currently enforced or recommended personal exposure limits for diesel particulate matter (DPM) and diesel exhaust [20–23].

Exhaust aftertreatment devices such as diesel oxidation catalyst (DOC), selective catalyst reduction (SCR) systems, and diesel particulate filter (DPF) systems are instrumental to efforts to substantially reduce diesel gaseous and particulate emissions. The DOCs are extensively used to control emissions of carbon monoxide (CO) and gas phase hydrocarbons (HC) [43, 44]. DOCs are also integrated into advanced DOC/DPF systems to promote generation of NO₂ to assist passive regeneration of DPF substrates [44–46] and support active regeneration of DPF substrates via catalytic combustion [44], and into DOC/SCR systems to maintain an NO₂/NO ratio needed to optimize performance of SCR systems [47, 48]. Full-flow DPF systems are a widely recognized technology for the reduction of solid diesel particulate mass (PM) and particulate number (PN) to levels required by current regulations [31, 49, 50]. The use of DPFs and filtration systems with disposable filter elements (DFEs) are critical in efforts to curtail DPM emissions from underground coal mining equipment in the USA [1]. For nitrogen oxide reduction, diesel exhaust fluid (DEF)-based SCR systems emerged as a most effective technology [35, 50, 51].

The use of exhaust aftertreatment devices was associated with the generation of secondary emissions in some cases [52, 53]. Increases in secondary emissions of highly toxic NO₂ [54] are of particular concern to the underground mining industry and have been observed for DOCs and DPFs coated with platinum group metals [55–57]. Due to the potential for secondary emissions, the US Mine Safety and Health Administration included additional requirements for using existing DPF systems and banned the acquisition of new retrofit type DPF systems that increase NO₂ concentrations beyond raw exhaust levels [58]. A potential solution to the

problem is the use of alternative base metal-palladium (Pd) coatings. Those types of coatings were found to be effective in oxidizing CO and HC at exhaust temperatures above 190 °C and were also effective in removing NO₂ at temperatures between 170 and 330 °C but promoted the limited formation of NO₂ at temperatures above 330 °C [59, 60]. Emissions of potentially high concentrations of nucleation mode aerosols for certain DPF-equipped engines [32, 39, 40] and DPF/SCR-equipped engines [32, 42, 51, 61] are of additional concern.

The objective of this study was to examine the potential of selected engineering controls based on retrofitting existing power packages with DOCs and DPF systems or repowering equipment with advanced power packages to reduce contribution of diesel emissions to the concentrations of aerosols in underground mines. A comparison of emissions was made for the retrofitted and repowered systems to assess the potential benefits of each control strategy on the reduction of miners' exposures.

2 Methodology

The results of laboratory evaluations of aerosol and gaseous emissions for two electronically controlled turbocharged diesel engines with similar power ratings but from different generations were used to support the evaluation. The first evaluated engine, Engine 1, is a 2004 4.3-liter Mercedes-Benz Model OM 904 LA (family 4MB XL4.25RJA) rated at 130 kW (174 bhp) @ 2200 rpm and 675 Nm (498 lb-ft) @ 1400 rpm and is typical of those currently used in underground mining in the USA [1]. That engine complied with the US Environmental Protection Agency (US EPA) tier 2 emission standards [27], and it was approved by the US Mine Safety and Health Administration (MSHA) for use in underground mines in the USA (Approval number: 7E-B098). Engine 1 was tested in three different configurations: (1) without exhaust aftertreatment (Engine 1), (2) retrofitted with a DOC Model MinNoDOC from AirFlow Catalyst Systems, Rochester, NY (Engine 1 DOC), and (3) retrofitted with full-flow DPF system Model Green Trap 1100 from NETT Technologies, Mississauga, ON (Engine 1 DPF). The results were used to examine the potential of retrofit-type exhaust

after treatment devices as a control strategy, to curtail emissions of aerosol and selected gaseous emissions from the previous generations of engines. It is important to note that the washcoat on the metal substrate of MinNoDOC was impregnated with a catalyst formulation intended to allow for the effective control of CO and HC emissions while also controlling NO₂ emissions [56]. The DPF substrate was also impregnated with a catalyst formulation that was designed to suppress generation of secondary NO₂ emissions. The DOC cylindrical canister was 622 mm (24.5") long and 305 mm (12") in diameter. The DPF cylindrical canister was 787 mm (31") long and 305 mm (12") in diameter.

The second engine, Engine 2, was 5.1-liter 2014 Mercedes Benz Model OM 934 LA (family EMBXL07.7RJA) rated at 129 kW (173 bhp) @ 2200 rpm and 750 Nm (535 lb-ft) @ 1400 rpm. That engine complied with US EPA tier 4 final emission standards [27] but did not have MSHA approval. Engine 2 is representative of the group of US EPA tier 4 final compliant advanced non-road engines that meet emissions standards through (1) implementation of combustion improvements, (2) use of cooled exhaust gas recirculation (EGR), and (3) use of an exhaust aftertreatment system that consisted of the DOC, DEF-based SCR, and ammonia slip catalyst (ASC) [31, 34, 35, 62].

The engines were coupled to a 400-kW water-cooled eddy-current dynamometer (SAJ, AE400) and evaluated at four steady-state (SS) operating conditions and one transient (TR) cycle. The SS modes, (1) rated speed 100 percent load (R100), (2) rated speed 50 percent load (R50), (3) intermediate speed 100 percent load (I100), and (4) intermediate speed 50 percent load (I50), were the subset of the International Organization for Standardization (ISO) 8-mode test cycle [63]. The selected engine operating parameters for the SS modes are shown for both engines in Table 1.

Back-to-back repetitions of a 900-s duration TR cycle (Fig. 1), custom-designed to recreate the duty cycle of an engine in underground mining load-haul-dump vehicles, were used for the evaluation of both engines over TR conditions. It is important to note that Engine 2 produced higher torque and power outputs at all test conditions than Engine 1 [Table 1].

Table 1 Engine operating conditions for Engine 1 and Engine 2

| Engine operating conditions | Engine 1 | | | Engine 2 | | |
|-----------------------------|------------------|-------------------|---------------|------------------|-------------------|---------------|
| | Engine Speed rpm | Torque Nm (lb-ft) | Power kW (hp) | Engine Speed rpm | Torque Nm (lb-ft) | Power kW (hp) |
| R100 | 2200 | 515 (380) | 119 (159) | 2200 | 542 (400) | 125 (168) |
| R50 | 2200 | 258 (190) | 59 (80) | 2200 | 271 (200) | 63 (84) |
| I100 | 1400 | 637 (470) | 93 (125) | 1400 | 719 (530) | 105 (141) |
| I50 | 1400 | 319 (235) | 47 (63) | 1400 | 359 (265) | 53 (71) |

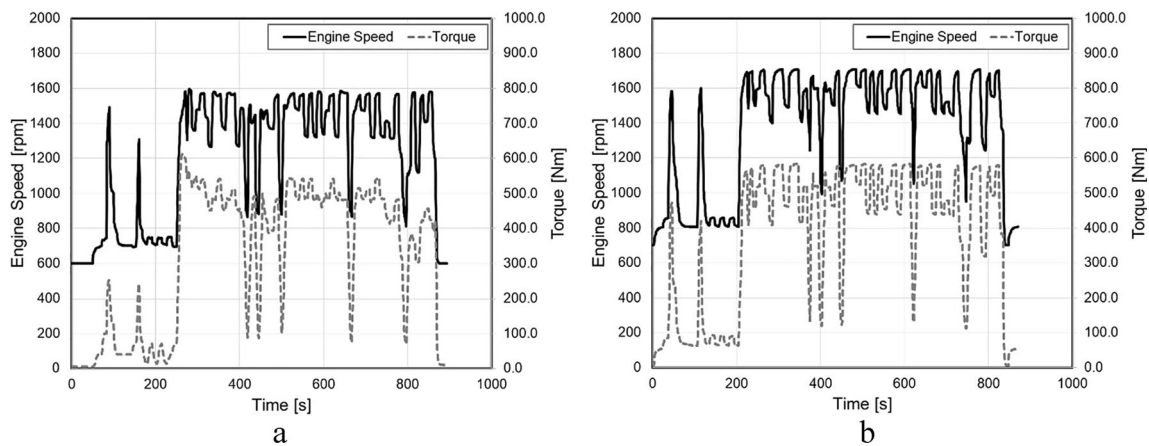


Fig. 1 TR mining cycles for **a** Engine 1 and **b** Engine 2

Test duration was determined based on requirements for the collection of adequate quantities of diesel aerosols on filters used for carbon analysis. The SS tests conducted to evaluate Engine 1, when tested without after treatment and retrofitted with a DOC, were 3600 s long. The tests conducted to evaluate Engine 1 when retrofitted with DPF and Engine 2 were 21,300 s and 14,400 s long, respectively. The TR cycle was repeated the necessary number of times to ensure that total duration of the TR tests was similar to the duration of the corresponding SS tests.

Additional testing was done on Engine 2 in order to assess (1) the volatility of SCR-Out aerosols and (2) the effects of exhaust temperature on CO, NO, and NO₂ emissions before (SCR-In) and after (SCR-Out) DOC/SCR/ASC system. The volatility tests were performed for Engine 2 operated at four SS engine operating conditions shown in Table 1 and three additional low engine speed and load operating conditions: LI (700 rpm and 0 Nm [0 lb-ft]), R12 (2200 rpm and 68 Nm [50 lb-ft]), and I9 (1400 rpm and 68 Nm [50 lb-ft]). In order to characterize CO, NO, and NO₂ emissions over the majority of its operating temperature range, Engine 2 was operated at several SS conditions at the rated and intermediate engine speeds (Fig. 2). For each of the engine speeds, the across spectrum exhaust temperatures were achieved by keeping engine speed constant while gradually intensifying engine load. The increased variabilities in SCR-In and SCR-Out temperatures for the cases when Engine 2 was operated at rated speed and torque outputs of 115 Nm (85 lb-ft), 136 Nm (100 lb-ft), and 163 Nm (120 lb-ft) (Fig. 2) were result of the attempts of the engine/exhaust aftertreatment management systems to keep SCR-Out exhaust temperatures above approximately 200 °C (392 °F) by managing additional fuel injections and EGR rate (Fig. 3).

Throughout the study, both engines were fueled with ultra-low sulfur diesel (ULSD) obtained from a single batch. The results of the analyses performed on that fuel are shown in Table 2. Engine 2 was supplied with DEF that meets ISO standard (32.5 percent urea) [64].

Measurements with direct reading instruments and filter samplings for aerosol characterizations were executed in exhaust diluted approximately 30 times using a two-stage partial dilution system (Dekati, Tampere, Finland, Model FPS4000). In order to account for test-to-test variations in dilution rates, the results were normalized to a nominal dilution ratio of 30. Triplicate filter samples for carbon analysis were collected from the dilution system on tandem 37-mm quartz fiber filters (QFFs, Pall Corporation, Ann Arbor, MI, 2500QAT-UP) enclosed in five-piece cassettes (SKC, Eighty Four, PA, 225-3050LF and 225-304). In order to minimize OC contamination of the media, the QFFs were pre-baked in a muffle furnace at 800 °C for 4 h. A nominal sampling flow rate of 1.7 lpm was maintained by subsonic critical orifices, installed in the manifolds coupled to a single vacuum pump (Oerlikon Leybold Vacuum GmbH, Cologne, Germany, Sogevac SV25B). The actual sampling flow rates were determined using results of flow verifications with a primary flow calibrator (Mesa Laboratories, Lakewood, CO, Bios Defender 530). The carbon analysis was performed at NIOSH PMRD using the thermal optical transmittance-evolve gas analysis (TOT-

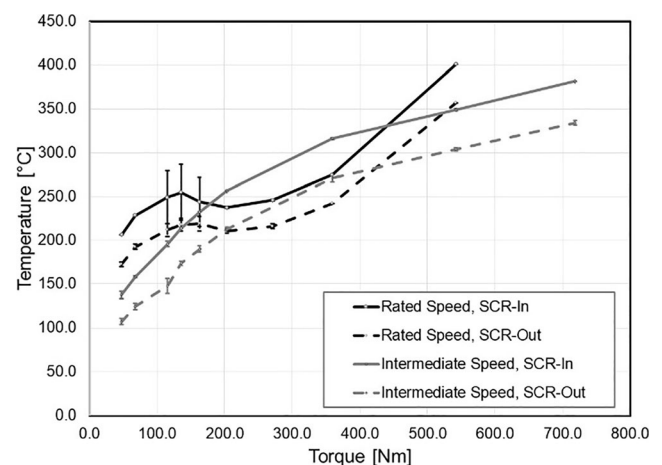


Fig. 2 Exhaust temperatures as a function of generated torque for Engine 2 operated at rated and intermediate engine speeds

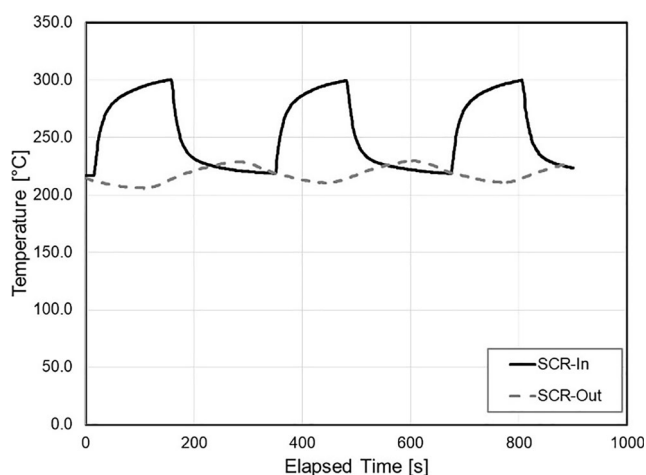


Fig. 3 SCR-In and SCR-Out exhaust temperatures for Engine 2 operated at rated engine speed and 136 Nm (100 lb-ft) of load

EGA) method NIOSH Method 5040 [65]. The analysis was performed using an OC/EC Aerosol Analyzer from Sunset Laboratory Inc. (Portland, OR). The results of the analysis performed on the secondary QFFs were used as a dynamic blank correction for the primary QFFs [65].

The fast mobility particle sizer spectrometer (FMPS, TSI, Minneapolis, MN, Model 3091) was used to measure, at 1 Hz frequency, the number concentrations and size distributions of non-volatile and volatile aerosols in diluted exhaust with an electrical mobility diameter between 5.6 and 560 nm. In order to enhance the clarity of the figures, the aerosol size distributions were fitted with log-normal curves using DistFit software from Chimera Technologies (Forest Lake, MN). The volatility of diesel aerosols in the diluted exhaust of Engine 2, operated at the selected SS engine operating conditions, was assessed. The assessment was made using a scanning mobility particle sizer spectrometer (SMPS, TSI, Model 3936), which measures the number concentrations and size distributions of aerosols with electrical mobility diameters between 10 and 400 nm. The SMPS measurements were performed on (1) exhaust diluted in partial dilution tunnel and (2) exhaust diluted in partial dilution tunnel and subsequently treated in the low-flow thermodenuder (TD, TSI, Model 3065). In the first

section of the TD, the diluted exhaust was preheated to 400 °C (752 °F) and subsequently, in the second section of the TD, the semi-volatile compounds were denuded via diffusion toward activated charcoal adsorbent. The concentrations of CO, NO, and NO₂ in undiluted exhaust were measured in 20-s intervals using a Fourier transform infrared (FTIR) spectrometer (Gasmeter Technologies Oy, Vantaa, Finland, DX-4000).

It is important to note that Engine 2 was equipped with closed crankcase breather while Engine 1 was equipped with an open filtered crankcase breather. As a result, the crankcase emissions were included in the assessed emissions for Engine 2, but not for Engine 1.

3 Results

The results of carbon analysis performed on the QFF samples collected from the 30:1 diluted exhaust are summarized in Fig. 4. The dilution factor was selected because it is representative of conditions in underground mines. The results showed that, depending on the SS operating mode, the evaluated DOC reduced, on average, 20 to 83% of OC and 24 to 49% of EC mass concentrations in exhaust emitted by Engine 1. At TR conditions, the use of the same DOC resulted in slight increase, within measurement error range, in average OC and EC mass concentrations.

The evaluated DPF removed on average over 92% of OC and 98% of EC emitted by Engine 1. Engine 2 emitted between 23 and 93% less OC and between 43 and 88% less EC than Engine 1 without aftertreatment. The OC and EC emissions for Engine 2 were particularly lower than the corresponding emissions for Engine 1 for high load conditions. However, it is important to note that Engine 2 emitted much more OC and EC than Engine 1 retrofitted with a DPF. For the selected test conditions, EC comprised between 66 and 92% of total carbon (TC) emitted by Engine 1, 77 to 91% of TC emitted by Engine 1 retrofitted with a DOC, and 76 to 85% of TC emitted by Engine 2. The uncertainty of measurements of relatively low OC and EC concentrations in the diluted exhaust of Engine 1 retrofitted with DPF was too high to assess the carbon makeup of those aerosols.

Figure 5 shows the results of number concentration measurements of aerosols in 30:1 diluted exhaust. For the SS tests, the DOC reduced the average number concentrations of aerosols emitted by Engine 1 by 22 to 52%. In the case of TR tests, the DOC increased average number concentrations of aerosols emitted by Engine 1 by 62%. For all test conditions, the DPF captured over 99% of particles emitted by Engine 1. The average number concentrations of aerosols emitted by Engine 2 (with DOC/SCR/ASC system) at all test conditions, were between 78 and 95% lower than the corresponding number concentrations of aerosols emitted by Engine 1. However, the average number concentrations of aerosols emitted by Engine 2 were between 88 and 99%

Table 2 Properties of the ULSD fuel used for this study

| Property | Test method | Unit | Value |
|--------------------|-------------|----------|-------|
| Specific gravity | ASTM D1298 | — | 0.830 |
| Aromatics content | ASTM D1319 | % volume | 21.7 |
| Olefins content | ASTM D1319 | % volume | 3.1 |
| Parafins content | ASTM D1319 | % volume | 75.2 |
| Cetane number | ASTM D613 | — | 47.3 |
| Flash point | ASTM D93 | K | 340 |
| Heat of combustion | ASTM D240 | MJ/kg | 45.9 |
| Sulfur content | ASTM D5453 | ppm | 5.6 |

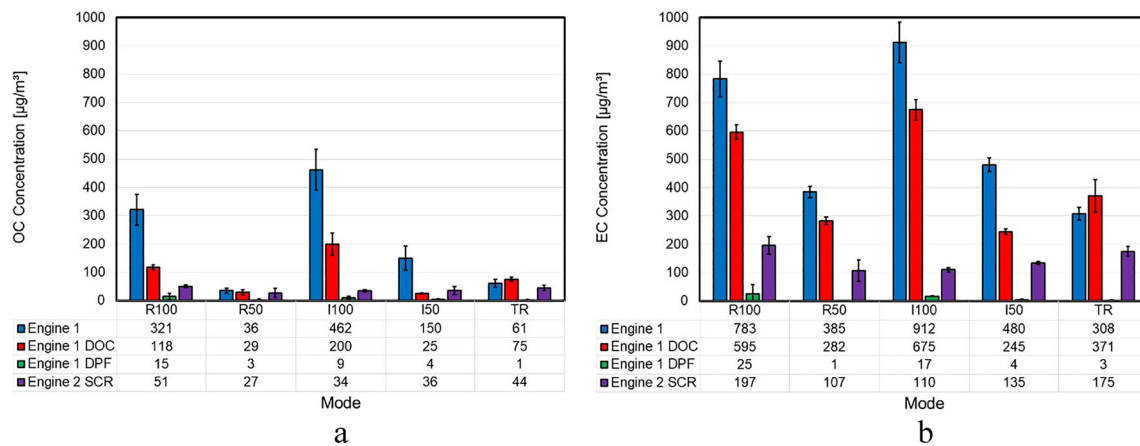


Fig. 4 Effects of the evaluated engine/exhaust aftertreatment technologies on concentrations of (a) OC and (b) EC in diluted exhaust (30 times)

higher than the corresponding average concentrations in the exhaust of the DPF-filtered Engine 1.

The size distribution of aerosols in the diluted exhaust for SS and TR engine operating conditions are shown in Figs. 6 and 7, respectively. Engine 1, when operated without after treatment at the R100 and R50 conditions, produced aerosols distributed predominantly in the accumulation mode with count median diameters (CMDs) around 60 nm and 70 nm, respectively, with the remaining aerosols distributed in the weaker nucleation mode with count median diameters around 10 nm (Fig. 6; Table 3). When operated at the I100 and I50 conditions, the same engine emitted aerosols distributed in single accumulation mode. When retrofitted with the DOC and operated in all SS modes, Engine 1 emitted aerosols distributed in single accumulation mode (Fig. 6) with the count median diameters similar to those observed for the size distributions of aerosols emitted the same engine operated without aftertreatment (Table 3). The peak concentrations emitted by Engine 1 retrofitted with DOC were found to be somewhat lower than those of the

agglomeration aerosols emitted by Engine 1 when operated without aftertreatment at SS modes (Table 3). The earlier mentioned increases in the average number concentrations observed after Engine 1 was retrofitted with DOC and operated over TR conditions can be primarily attributed to the increase in concentrations of nucleation mode aerosols (Fig. 7b).

The relatively less abundant aerosols in the DPF-treated exhaust of Engine 1 were distributed between two or three modes (Figs. 6 and 7). Nucleation mode aerosols with count median diameters around 10 nm were found in filtered exhaust for all SS and TR operating conditions. The concentrations of nucleation mode aerosols were comparable or less than those of accumulation mode aerosols (Figs. 6 and 7). Slight increases in concentrations of aerosols in the sub-20 nm range over engine-out levels were only found for I100 and I50 conditions. For R50 and I100 conditions, one of the remaining modes appeared to be a carbon-based accumulation mode with count median diameter around 80 nm.

The distributions of aerosols emitted by Engine 2 were bimodal with the majority of aerosols in the accumulation mode with count median diameters around 49 nm and 57 nm and remaining aerosols distributed in less pronounced nucleation modes with count median diameters between 12 and 14 nm (Fig. 6). It is important to note that the count median diameters of the agglomeration aerosol emitted by Engine 2 were 10 to 25 nm smaller than those of the agglomeration aerosols emitted by Engine 1, when operated without after treatment and with DOC (Table 3).

Volatility tests performed on aerosols emitted by Engine 2 showed that the TD heated to 400 °C (752 °F) removed 45 to 85% of aerosols emitted at selected steady state conditions (Fig. 8). In the case of LI conditions, the majority of the removed aerosols were sub-40 nm nucleation mode aerosols. In other cases, the TD removed aerosols in whole spectrum of sizes.

The 3600-s averages of CO, NO, and NO₂ concentrations in the raw exhaust of Engine 1 and Engine 2 operated over four steady-state and TR conditions are shown in Fig. 9. When retrofitted with the DOC, Engine 1 emitted less CO (33 to

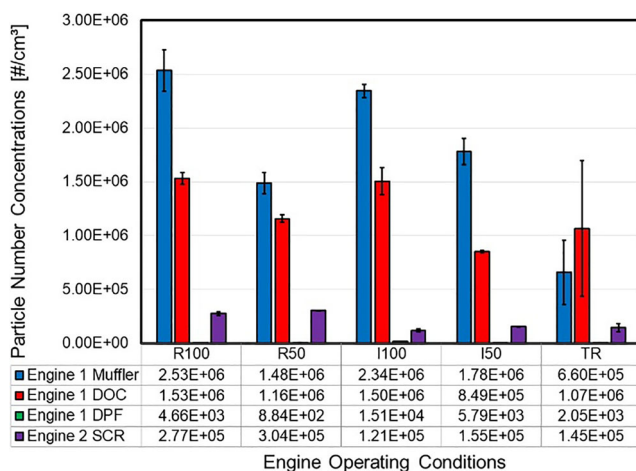


Fig. 5 Effects of the evaluated engine/exhaust aftertreatment technologies on average number concentrations of aerosols in diluted exhaust (30 times) for SS (R100, R50, I100, and I50) and TR operating conditions

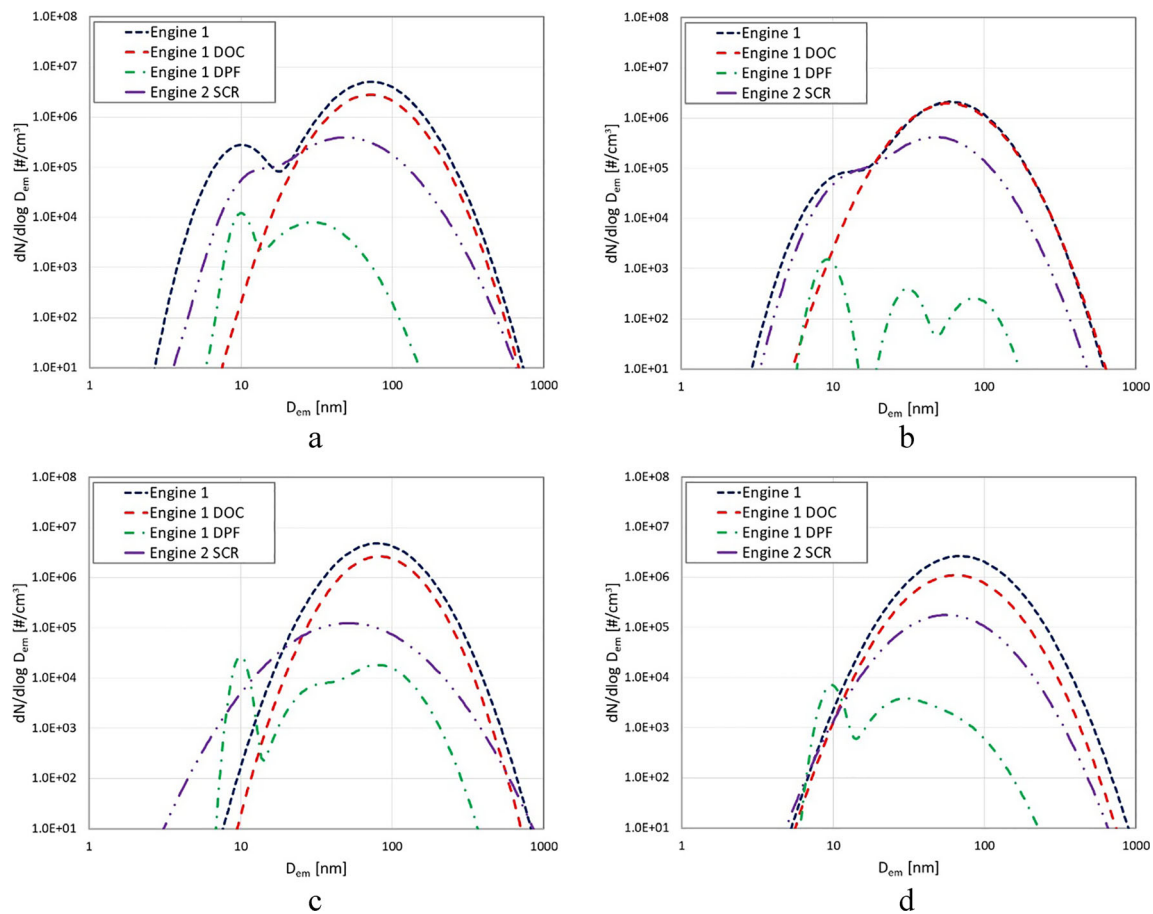


Fig. 6 Effects of the evaluated engines and exhaust aftertreatment technologies on size distribution of aerosols in diluted exhaust (30 times) for SS engine operating conditions. **a** R100. **b** R50. **c** I100. **d** I50

86%), more NO (3 to 31%), and less NO₂ (52 to 89%) than without the retrofit. When operated with the DPF, Engine 1 emitted less CO (32 to 87%), more NO (3 to 31%), and less NO₂ (52 to 89%) than without the retrofit. At the corresponding engine operating conditions, Engine 2 with DOC/SCR/ASC after treatment emitted on average between 59 and

99% less CO, 70 and 93% less NO, and 30 to 97% less NO₂ than Engine 1 operated without aftertreatment.

The results of the additional 900-s tests conducted at rated and intermediated speeds (Fig. 2) were used to examine the effects of exhaust temperature on SCR-Out CO, NO, and NO₂ emissions from Engine 2 (Fig. 10). Due to low catalytic activity and the

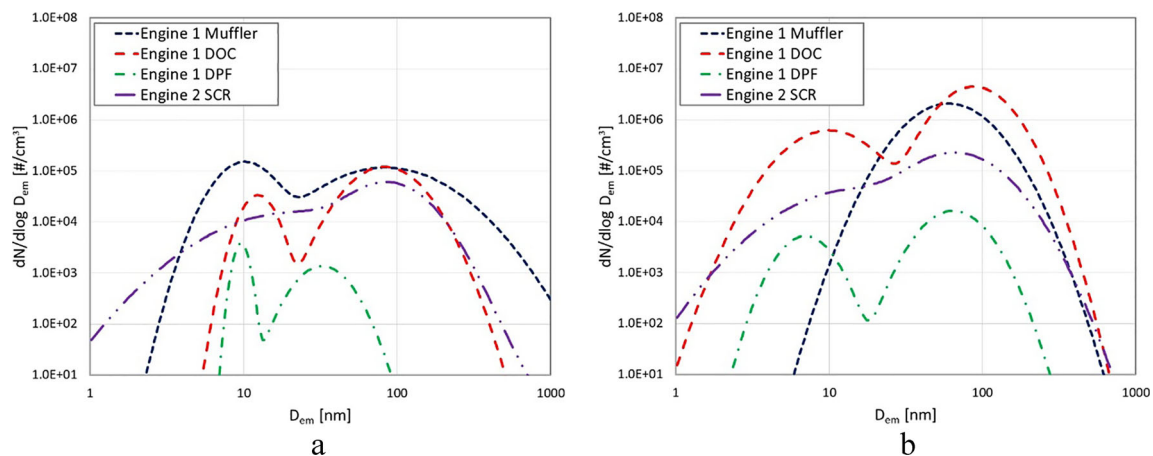


Fig. 7 Effects of the evaluated engines and exhaust aftertreatment technologies on size distribution of aerosols in diluted exhaust (30 times) for TR engine operating conditions. **a** Selected instance of low number concentrations (TR low). **b** Selected instance of high number concentrations (TR high)

Table 3 Statistical parameters for the size distributions of aerosols emitted by the engines operated at SS and TR conditions

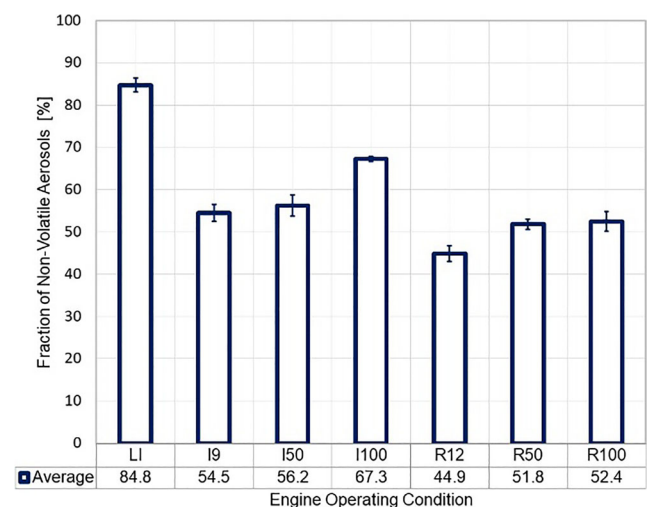
| Mode | Exhaust Aftertreatment | Nucleation | | | Accumulation 1 | | | Accumulation 2 | | |
|---------|------------------------|------------|---------------|----------------------------------|----------------|---------------|----------------------------------|----------------|---------------|----------------------------------|
| | | CMD nm | σ – | Total Conc. #/cm ³ | CMD nm | σ – | Total Conc. #/cm ³ | CMD nm | σ – | Total Conc. #/cm ³ |
| R100 | Engine 1 Muffler | 10.1 | 1.340 | 8.94E+04 | 72.6 | 1.570 | 2.46E+06 | | | |
| | Engine 1 DOC | | | | 71.5 | 1.570 | 1.36E+06 | | | |
| | Engine 1 DPF | 10.0 | 1.130 | 1.57E+03 | 29.7 | 1.560 | 3.87E+03 | | | |
| | Engine 2 SCR | 12.4 | 1.310 | 1.94E+04 | 48.7 | 1.760 | 2.44E+05 | | | |
| R50 | Engine 1 Muffler | 12.1 | 1.400 | 3.50E+04 | 60.3 | 1.600 | 1.32E+06 | | | |
| | Engine 1 DOC | | | | 58.6 | 1.620 | 1.27E+06 | | | |
| | Engine 1 DPF | 9.3 | 1.160 | 3.07E+02 | 31.3 | 1.190 | 9.47E+01 | 85.6 | 1.310 | 1.69E+02 |
| | Engine 2 SCR | 13.7 | 1.400 | 3.04E+04 | 47.6 | 1.650 | 2.83E+05 | | | |
| I100 | Engine 1 Muffler | | | | 79.5 | 1.580 | 2.40E+06 | | | |
| | Engine 1 DOC | | | | 82.2 | 1.540 | 1.25E+06 | | | |
| | Engine 1 DPF | 10.0 | 1.100 | 3.50E+03 | 31.4 | 1.350 | 2.57E+03 | 80.5 | 1.480 | 2.10E+04 |
| | Engine 2 SCR | 12.4 | 1.400 | 2.20E+04 | 51.5 | 1.910 | 8.70E+04 | | | |
| I50 | Engine 1 Muffler | | | | 68.7 | 1.670 | 1.82E+06 | | | |
| | Engine 1 DOC | | | | 64.5 | 1.660 | 7.49E+05 | | | |
| | Engine 1 DPF | 9.7 | 1.140 | 1.26E+03 | 26.8 | 1.380 | 1.09E+03 | 45.9 | 1.630 | 2.99E+03 |
| | Engine 2 SCR | 12.1 | 1.560 | 2.03E+04 | 56.5 | 1.740 | 1.32E+05 | | | |
| TR Low | Engine 1 Muffler | | | | 83.9 | 2.050 | 1.12E+05 | | | |
| | Engine 1 DOC | | | | 82.8 | 1.520 | 6.78E+04 | | | |
| | Engine 1 DPF | 9.6 | 1.100 | 5.00E+02 | 32.5 | 1.390 | 6.10E+02 | | | |
| | Engine 2 SCR | 21.6 | 2.460 | 1.85E+04 | 87.8 | 1.580 | 3.44E+04 | | | |
| TR High | Engine 1 Muffler | | | | 86.1 | 1.590 | 1.70E+06 | | | |
| | Engine 1 DOC | | | | 88.1 | 1.490 | 2.41E+06 | | | |
| | Engine 1 DPF | 7.0 | 1.370 | 2.23E+03 | 64.3 | 1.460 | 8.27E+03 | | | |
| | Engine 2 SCR | 15.4 | 2.229 | 4.54E+04 | 67.2 | 1.690 | 1.55E+05 | | | |

absence of urea injections, the CO and NO emissions were highest for the engine operating conditions that generated exhaust temperatures below 200 °C (392 °F) (Fig. 10). At temperatures above 200 °C (392 °F), the system was found to effectively convert CO and NO. Evidence of limited NO₂ formation was found for an engine operating condition that generated temperatures above 300 °C (572 °F) (I100).

4 Discussion and Conclusion

The results indicated that the evaluated DOC effectively curtailed US EPA Tier 2 engine emissions of CO, and to some extent OC, EC, and aerosol number concentrations. The reductions were achieved while avoiding secondary emissions of acutely toxic NO₂, which suggests that the DOC is a viable option for reducing exposures in underground mines. Nevertheless, substantially higher reductions of OC and EC mass concentrations and total number concentrations of aerosols were achieved by retrofitting the Tier2 engine with a DPF system. EC concentrations were about 100 times lower for the

DPF-treated exhaust. The results showed that both evaluated systems, the DOC and DPF, achieved reductions in aforementioned emissions without adversely affecting emissions of

**Fig. 8** Average fractions of non-volatile aerosols in the exhaust of Engine 2 (TD temperature 400 °C [752 °F])

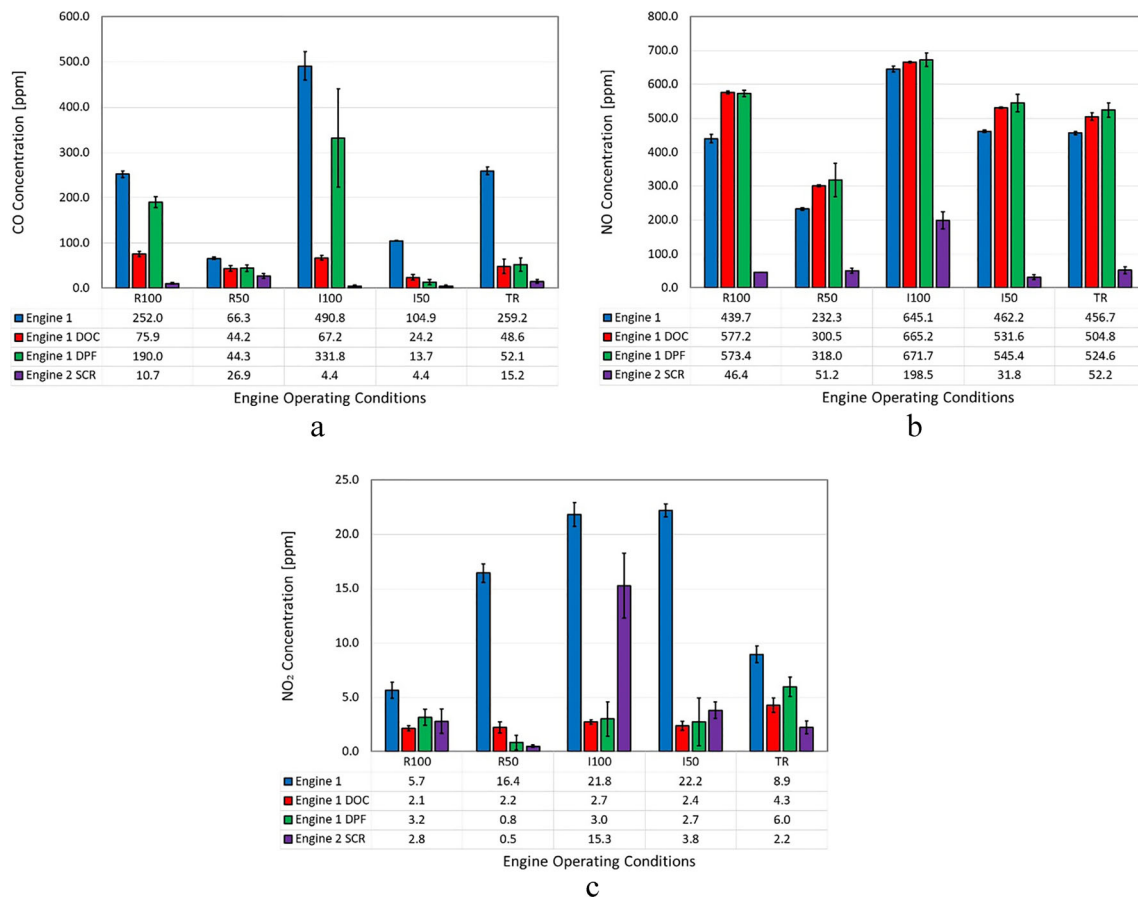


Fig. 9 Effects of the evaluated engines and exhaust aftertreatment technologies on **a** CO, **b** NO, and **c** NO₂ emissions

NO₂ and nucleation mode aerosols, and therefore, those two devices could be suitable for addressing selected emissions from diesel-powered vehicles from existing fleets. The suitability of the DPF systems for the specific applications would depend on a number of engineering parameters and primarily on the viability of the applied DPF regeneration strategy. Establishing a relationship between exhaust temperature

profiles for the specific vehicle duty cycles and balance point temperatures for the passive DPF regeneration is critical to the success of those applications. In terms of developing adequate DPF regeneration strategies and optimizing the performance of other exhaust aftertreatment devices, retrofit solutions have disadvantages over OEM solutions that allow for the integration of engine and exhaust aftertreatment systems. One example was the manipulation of fuel injection and EGR rates to achieve desired SCR-Out exhaust temperatures.

The evaluations showed that, for all test conditions, the EPA Tier 4f engine contributed substantially less than the US EPA Tier 2, when operated without exhaust aftertreatment, to the mass concentrations of EC and OC, number concentrations of aerosols, and concentrations of CO, NO, and NO₂. It is important to note that the relatively low DPM mass emissions were achieved without the use of a DPF system and that reductions in NO and NO₂ emissions were achieved in part using cooled EGR and SCR systems. The results of size distribution measurements indicated that the reductions in mass concentrations of aerosols may have been partially achieved through better in-cylinder fuel and air mixing and hence reductions in the size of emitted aerosols. Therefore, repowering existing vehicles currently powered by US EPA Tier 2 (and US EPA Tier 3) with US EPA Tier 4 final engines fitted with

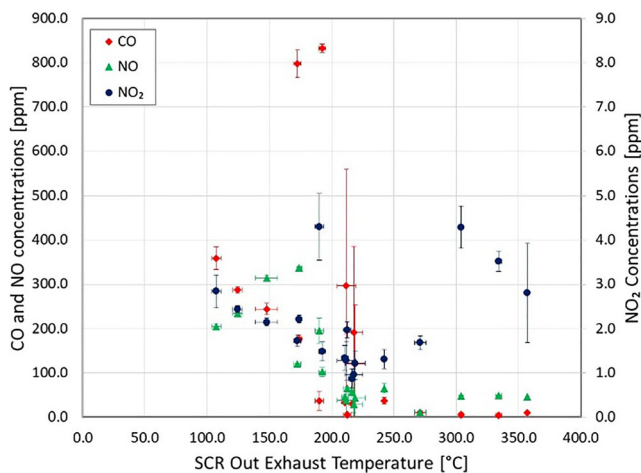


Fig. 10 Effects of exhaust temperatures of CO, NO, and NO₂ SCR-Out emissions

DEF-based SCR systems, similar to the one tested in this study, could substantially reduce the contribution of existing and newly introduced diesel-powered underground mining vehicles to mass concentrations of submicron aerosols and criteria gases in underground mines. Tier 4f engines may prove to be a viable technology for helping the underground mining industry to comply with current mass-based regulations limiting the exposure of underground metal/nonmetal miners to DPM [21].

In addition, the results demonstrated that tested US EPA Tier 4 engine was effective in reducing particulate mass and cumulative emissions of nitrogen oxides, but not equally effective in reducing particulate number emissions. The results confirmed findings of Lucachick et al. [31] and Fiebig et al. [66] that the use of DPFs in advanced exhaust aftertreatment systems would be critical to efforts to reduce contributions of such engines to aerosol number concentrations. Exhaust aftertreatment systems with DOC, DPF, and SCR systems are expected to be an integral part of European Stage V non-road power packages with power outputs between 56 and 560 kW (75 and 750 hp) [67, 68]. The results demonstrate that the effects of control technologies and strategies on physical and chemical properties of emitted aerosols should not be neglected during the selection process.

A number of economic and technical aspects need to be taken into consideration prior to the industrywide implementation of advanced exhaust aftertreatment systems and engine systems in underground mining. Some of the parameters affecting implementation of these technologies in underground mining operations are of additional technical complexity, space requirements, higher capital and operational costs, and fluid requirements. Implementation of technologies needed to meet stringent standards are associated with an increase in technical complexity and capital cost [67]. The higher costs are associated with improvements in a number of engine systems including fuel injection, turbocharging, EGR, and engine and exhaust aftertreatment control. Dallmann et al. [68] suggested that the increase in incremental cost is primarily driven by the adoption of DPF and SCR systems. The space requirements for the installation of the DPF system, similar to the one retrofitted to the Tier 2 engine, and the SCR system and associated hardware, similar to the one on the Tier 4f, could be considerable. The installation of those in the engine bays of existing and new underground mining equipment might prove to be difficult and only possible with extensive redesign and optimization of the equipment and systems.

Increased maintenance costs are associated with increased complexity of engine and exhaust aftertreatment systems. The operation of advanced engines depends strongly on the use of pressure, temperature, and gas sensors, and special precautions should be taken to avoid potential problems associated with the overexposure of underground miners to NO and NO₂ due to the failure of the SCR systems to inject DEF [69].

Implementation of engines fitted with DEF-based SCR systems would require establishing procedures for managing the supply of DEF. The availability of high-quality fuels with low sulfur content (< 15 ppm) and lubricants with low ash content (API CJ-4 and CK-4) are critical to the operation of catalyzed DPF systems [46]. The cleanliness of the fuel is critical to protecting the fuel system components of modern engines equipped with high-pressure common rail systems [70].

Implementation of viable exhaust aftertreatment systems and advanced diesel power packages would be instrumental to the underground mining industry worldwide to secure an economical and dependable source of power for mobile equipment that does not generate emissions that adversely affect miner's health.

Compliance with Ethical Standards

Conflict of Interest The authors declare that they have no conflict of interest.

Disclaimer The findings and conclusions in this paper are those of the authors and do not necessarily represent the official position of the National Institute for Occupational Safety and Health, Centers for Disease Control and Prevention. Mention of company names or products does not constitute endorsement by NIOSH.

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