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RESEARCH ARTICLE



Maternal levels of endocrine disrupting chemicals in the first trimester of pregnancy are associated with infant cord blood DNA methylation

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ABSTRACT

Endocrine disrupting chemicals (EDCs) pose a public health risk through disruption of normal biological processes. Identifying toxicoepigenetic mechanisms of developmental exposure-induced effects for EDCs, such as phthalates or bisphenol A (BPA), is essential. Here, we investigate whether maternal exposure to EDCs is predictive of infant DNA methylation at candidate gene regions. In the Michigan Mother-Infant Pairs (MMIP) cohort, DNA was extracted from cord blood leukocytes for methylation analysis by pyrosequencing ($n = 116$) and methylation changes related to first trimester levels of 9 phthalate metabolites and BPA. Growth and metabolism-related genes selected for methylation analysis included imprinted (*IGF2*, *H19*) and non-imprinted (*PPARA*, *ESR1*) genes along with LINE-1 repetitive elements. Findings revealed decreases in methylation of LINE-1, *IGF2*, and *PPARA* with increasing phthalate concentrations. For example, a log unit increase in Σ DEHP corresponded to a 1.03 [95% confidence interval (CI): -1.83 , -0.22] percentage point decrease in *PPARA* methylation. Changes in DNA methylation were also inversely correlated with *PPARA* gene expression determined by RT-qPCR ($r = -0.34$, $P = 0.02$), thereby providing evidence in support of functional relevance. A sex-stratified analysis of EDCs and DNA methylation showed that some relationships were female-specific. For example, urinary BPA exposure was associated with a 1.35 (95%CI: -2.69 , -0.01) percentage point decrease in *IGF2* methylation and a 1.22 (95%CI: -2.27 , -0.16) percentage point decrease in *PPARA* methylation in females only. These findings add to a body of evidence suggesting epigenetically labile regions may provide a conduit linking early exposures with disease risk later in life and that toxicoepigenetic susceptibility may be sex specific.

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Introduction

Endocrine disrupting chemicals (EDCs) are ubiquitous in modern society due to their widespread use in consumer products (e.g., food and beverage containers), personal care products (e.g., lotion), and medical supplies (e.g., plastic tubing). Phthalates and bisphenol A (BPA) are two of the most commonly studied EDCs, and both human and animal studies suggest these chemicals can disrupt normal biological processes [1]. BPA can impact endocrine function by acting as an estrogenic [2] and anti-androgenic compound [3]. In addition, BPA and some phthalates can alter the thyroid hormone balance [4], which can contribute to metabolic dysfunction in adults [5]. While exposure at any life stage may be impactful, the developing human fetus is particularly vulnerable to EDC exposure as this is a critical period of rapid growth, and perturbations during this time can have lasting effects on disease susceptibility [6]. To date, the exposure-induced phenotypes reported in the literature, with regard to BPA and phthalates, have been inconsistent with the direction of association differing by population, age, and sex [7]. Nevertheless, multiple cohort studies have

found that prenatal exposures to these toxicants are associated with growth-related outcomes [8–11]. A growing body of animal data bolsters these epidemiological findings. Specifically, studies in rodents have demonstrated that perinatal exposure to BPA is associated with persistent changes in body weight, food intake or preference, and hormone levels [12–16]. Therefore, with a global obesity epidemic as a backdrop, it is imperative to identify aberrant exposure-induced programming events that have the potential to shape metabolic outcomes later in life.

Epigenetic modifications such as DNA methylation are mitotically heritable alterations capable of modulating gene expression without changing the underlying DNA sequence [17]. As such, DNA methylation is at the interface of genetics and the fetal environment with the ability to act as an adaptive layer of regulatory control for gene expression and retrotransposon repression [18,19]. Further, evidence suggests that toxicoepigenetic changes can occur in response to BPA and phthalates [20]. Using the viable yellow agouti mouse model, we have demonstrated that perinatal BPA exposure alters coat color distribution with associated shifts in DNA methylation profiles in the

offspring [21,22]. Observations from human cohort studies suggest BPA can influence epigenetic programming of fetal liver enzymes [23] and imprinted genes [24]. Interestingly, perinatal BPA-induced epigenetic effects are commonly sex specific [25] and linked to DNA methylation changes at a number of loci including genes involved in liver beta oxidation [26], energy homeostasis [27], and growth and metabolism [24,28]. Maternal exposure to phthalates averaged across pregnancy has been found to be inversely associated with methylation profiles of retrotransposons Alu and LINE-1 in a population of Mexican-American children [29]. Studies with the ELEMENT Mexico City birth cohort also found maternal phthalate exposure in the third trimester is associated with altered methylation of *H19* [24], which is involved in body composition and growth [30]. Given the growing evidence demonstrating the toxicoepigentic potential of EDCs along with the potential developmental contribution to metabolic-related disease risk, there exists a clear need to identify molecular biomarkers of exposure. Identification of EDC sensitive epigenetic biomarkers will aid in elucidating the developmental origins of metabolic diseases. Therefore, using a biologically accessible source of surrogate DNA and RNA, the goal of this study was to determine whether first trimester maternal exposures to EDCs are predictive of newborn DNA methylation and whether DNA methylation is associated with gene expression levels at growth and metabolism-related candidate regions.

Methods

Study population

Women were recruited between 2010 and 2015 during their first trimester of pregnancy as part of the Michigan Mother Infant-Pairs (MMIP) project, an ongoing birth cohort. Prospective participants were informed of the study during their first prenatal visit at a University of Michigan clinic, and were eligible if they were 18 years of age or older, conceived naturally, and had a singleton pregnancy. Women provided spot urine and venous blood samples during their first trimester prenatal visit (8–14 weeks). Infant cord blood was collected at time of birth. The University of Michigan Medical School Institutional Review Board approved this study, and all women provided written informed consent prior to participation.

Phthalate metabolite and BPA measurement

Spot urine samples were collected into polypropylene urine collection containers, aliquoted into glass vials, and frozen at -80°C until analysis. Urinary BPA (56 out of the cohort of 116) and nine phthalate metabolites (109 out of the cohort of 116), comprising monoethyl phthalate (MEP), mono-n-butyl phthalate (MnBP), monoisobutyl phthalate (MiBP), monobenzyl phthalate (MBzP), mono-3-carboxypropyl phthalate (MCPP), mono-2-ethylhexyl phthalate (MEHP), mono-2-ethyl-5-hydroxyhexyl phthalate (MEHHP), mono-2-ethyl-5-oxohexyl phthalate (MEOHP), and mono-2-ethyl-5-carboxypentyl phthalate (MECPP), were measured at NSF International in two batches (Ann Arbor, MI) using isotope dilution–liquid chromatography–tandem mass spectrometry (ID–LC–MS/MS) as previously

described [31]. Summary measures for parent compounds di-(2-ethylhexyl) phthalate (ΣDEHP) and dibutyl phthalate (ΣDBP) for each sample were calculated by dividing their respective individual metabolite concentrations by their molar mass and summing them. The ΣDEHP measure comprised MEHP, MEHHP, MEOHP, and MECPP, while the ΣDBP measure comprised MnBP and MiBP. Specific gravity (SG) was measured using a handheld digital refractometer (Atago Co., Ltd., Tokyo, Japan) at the time of sample analysis. Values below the limit of detection (LOD) were replaced with $\text{LOD}/\sqrt{2}$. While the maternal participants recruited in the second half of the MMIP study had total urinary BPA measured at NSF International, the participants in the first half of the study had unconjugated plasma BPA ($n = 60$ out of the cohort of 116) measured in plasma at the Wadsworth Center (Albany, NY) as previously described [32]. Plasma BPA processing was performed following collection and analysis methods developed and validated by four independent laboratories [33]. The geometric mean and standard deviation for BPA concentrations measured in the two matrices were similar (Table 1); however, because no subject had both matrices analyzed, we have no way of doing a head-to-head comparison. Therefore, plasma BPA and urinary BPA were considered distinct exposure measurements and not one combined exposure measurement.

Data collection for covariates and potentially confounding variables

The infant's sex, gestational age, and birth weight as well as maternal pre-pregnancy weight, height, and age were taken from the medical records. Maternal pre-pregnancy weight and height were used to calculate maternal body mass index (BMI). Self-reported ethnicity, maternal smoking status, as well as household income were collected via survey.

DNA isolation and methylation analysis

Infant cord blood was collected into Paxgene Blood DNA and RNA tubes (PreAnalytix) at the time of birth and stored at -80°C until processing. Total genomic DNA was isolated using the Paxgene Blood DNA kit. Genomic DNA was bisulfite converted using the EZ-96 DNA Methylation Kit (Zymo). Briefly, sodium bisulfite was added to approximately 500 ng of genomic DNA, converting unmethylated cytosines to uracil, which are replaced with thymine during PCR; methylated cytosines remain unchanged [34]. For this study, we selected imprinted (*IGF2*, *H19*) and non-imprinted (*PPARA*, *ESR1*) genes along with LINE-1 repetitive elements as potential candidate regions of epigenetic lability. Given its prevalence across the genome, LINE-1 can be used as a surrogate for global DNA methylation levels [35]. Apart from LINE-1, all of the other interrogated genes play a role in metabolism, growth, or development. Insulin-like growth factor II (*IGF2*) and *H19* are well-characterized imprinted genes, in which parent-of-origin monoallelic expression is involved in the regulation of body composition and growth [30,36–38]. *PPARA* is a non-imprinted gene that encodes the peroxisome proliferator-activated receptor alpha (*PPAR- α*) protein, a nuclear receptor that regulates fatty acid metabolism [39–41]. *ESR1* is a non-imprinted gene that encodes estrogen receptor alpha (*ER- α*), a transcription factor

Table 1. Population statistics.

		n	%	mean (SD)	min	max
Child's Sex	Male	62	53			
	Female	54	47			
Race/Ethnicity	Non-Hispanic White	103	89			
	Hispanic	3	3			
	African American	2	2			
	Multiracial	1	1			
	Native American	1	1			
	Asian	4	3			
	Pacific Islander	2	2			
Gestational age (days)		116		278 [7]	252	292
Birth weight (g)		116		3510 (445)	2270	4685
Maternal pre-pregnancy BMI		110		25.28	19.05	48.41
Maternal age (years)		116		31 [4]	22	42
Maternal smoking	Did not smoke during pregnancy	94	81			
	Did smoke during pregnancy	10	9			
	Missing	12	10			
Family income (\$)	<49,999	28	24			
	50,000 - 99,999	44	38			
	> 100,000	41	35			
	Missing	3	3			
Specific gravity		109		1.02 (0.01)	1.00	1.03
Bisphenol A (BPA) (ng/ml) ^a	Urine	56		0.57 (4.72)	0.04	4.76
	Plasma	60		0.78 (5.47)	0.14	96.43
Phthalates (ng/ml) ^a	Monobenzyl phthalate (MBzP)	109		3.49 (3.78)	0.25	78.8
	Mono(3-carboxypropyl) phthalate (MCPP)	109		1.58 (4.18)	0.35	351.85
	Monoethyl phthalate (MEP)	109		22.03 (4.14)	1.28	1598.00
ΣPhthalates (nMol) ^a	ΣDEHP Metabolites	109		0.09 (0.11)	0.005	0.75
	ΣDBP Metabolites	109		0.08 (0.09)	0.002	0.43
DNA methylation (%)	LINE1	113		79.97 (3.22)	69.59	92.24
	H19	109		55.60 (4.94)	45.75	84.82
	IGF2	102		47.85 (3.68)	38.78	61.76
	ESR1	92		4.31 (0.96)	0.00	7.42
	PPARA	100		17.29 (2.66)	9.29	24.98

^aPhthalates and BPA are presented as geometric mean (SD).

involved in regulation of energy homeostasis [42]. The gene-specific PCR and pyrosequencing primers and conditions for the target regions are listed in Supplemental Table 1. PCR amplification was performed after bisulfite conversion using HotStarTaq master mix (Qiagen), forward primer (50 pmol), and reverse biotinylated primer (50 pmol) in a 30 μ l reaction. PCR fragments were analyzed by gel electrophoresis or automated capillary electrophoresis using the Qiaxcel Advanced System (Qiagen). DNA methylation quantification of CpG sites was performed using pyrosequencing on a PyroMark ID instrument (Qiagen). To determine percent methylation, PyroMark software calculated the fraction of methylated cytosines (%mC) among the total sum of methylated and unmethylated cytosines. For quality assurance, all pyrosequencing plates included 0, 50, and 100% methylated bisulfite converted human control DNA (Qiagen), as well as at least one no DNA template control. A subset of the samples from every 96-well plate was run in technical duplicate to calculate a coefficient of variation (CV) and if the average %CV was greater than 5% the plate was repeated.

Real-time quantitative PCR (RT-qPCR)

RNA was extracted from whole cord blood samples using the Paxgene Blood RNA kit (PreAnalytix). RNA concentration and purity were checked using a NanoDrop 2000 spectrophotometer (Thermo Scientific, Wilmington, DE). Complementary DNA (cDNA) synthesis was performed on 1 μ g RNA template for each sample with the Bio Rad iScript cDNA Synthesis kit (Hercules, CA) according to

manufacturer's instructions. Each cDNA sample was diluted 1:2 and added to a mixture of gene-specific forward and reverse primer, nuclease-free water, and iQ SYBR Green Supermix followed by 2-step PCR + melt detection on a Bio-Rad CFX96 system (Hercules, CA) using the following parameters: one cycle of 95°C for 3 minutes, followed by 45 cycles of [95°C for 10 seconds, 55°C for 30 seconds, plate read], and finally one cycle of 95°C for 10 seconds. The melt curve for each plate was 65°C - 95°C; 5°C increment for 5 seconds, with plate read at each temperature. For RT-qPCR analysis, individual samples were run in triplicate for *PPARA* and three housekeeping genes [Beta-actin (*B-actin*), Glyceraldehyde 3-phosphate dehydrogenase (*GAPDH*), and Tyrosine 3-Monooxygenase/Tryptophan 5-Monooxygenase Activation Protein Zeta (*YWHAZ*)]. In addition to housekeeping genes, an inter-plate control and genomic DNA control were included to improve quality assurance and facilitate the calculation of relative expression using the $2^{-\Delta\Delta C_t}$ method [43]. The gene-specific primer sets used and their respective literature sources are listed in Supplementary Table 2. Primer pair specificity for all designed primers was checked using the NCBI Primer-BLAST online tool (<https://www.ncbi.nlm.nih.gov/tools/primer-blast/>).

Statistical methods

DNA methylation analysis was conducted using the average methylation value for all measured CpGs for each candidate gene. To adjust for heterogeneity in variance, BPA and phthalate were ln-transformed prior to regression analysis. The

distribution of ln-transformed EDCs was examined across categories of sociodemographic and perinatal characteristics using simple linear regression for continuous variables and ANOVA with post-hoc Tukey HSD for categorical variables to identify potential confounders. To adjust for urine dilution in linear regression analyses, urinary SG was included as a covariate in models using continuous urinary EDC variables, as we have done previously [31]. In fully adjusted models, the child's sex, maternal pre-pregnancy body mass index (BMI) and maternal age were included as potential confounders based on a priori expectations informed by previous studies and current literature. This method of analysis was used to evaluate the relationship between EDC exposure and DNA methylation. Other covariates that were considered included household income and smoking status. When income was included in the model, the effect sizes did not change appreciably (>10%) and therefore was not included in the final model. Similarly, smoking status was not included because the majority of women were non-smokers. The results of the multivariable linear regression model for DNA methylation are presented as an absolute percentage point change in DNA methylation outcome [95% confidence interval (CI)] per log unit increase in continuous EDC value. *PPARA* relative expression values were log-normally distributed and ln-transformed. Spearman's rank correlation test was used to evaluate the relationship between *PPARA* methylation and *PPARA* gene expression. Given that the literature suggests EDC-induced effects may be sexually dimorphic, sex-stratified associations between exposure and methylation outcomes were also investigated.

Results

Demographics, exposure distribution, and DNA methylation

Table 1 summarizes maternal and infant population characteristics, maternal first trimester EDC levels, and infant cord blood DNA methylation. Women in the present analysis represent 116 participants in the MMIP cohort study. This population of women was 31 years of age on average, mostly white, slightly overweight, and non-smoking. The average infant birth weight and gestational age was 3510 grams and 39.7 weeks, respectively. Overall, EDC measurements varied widely across this population. Plasma BPA had a wider range (0.14 to 96.43 ng/ml) relative to urinary BPA (0.04 to 4.76 ng/ml), while the geometric means were

comparable [0.57 ng/ml (SD 4.72) and 0.78 ng/ml (5.47), respectively]. Measured levels of MBzP, MCP, and MEP ranged from 0.25 to 78.8 ng/ml, 0.35 to 351.85 ng/ml, and 1.28 to 1,598.00 ng/ml, respectively. After conversion to the sum of their phthalate metabolites, the geometric mean of Σ DEHP and Σ DBP were similar [0.09 nMol (SD 0.11) and 0.08 nMol (0.09), respectively]. The inter-individual variation across CpG loci for each of the candidate genes [*ESR1* (4 sites), *PPARA* (2 sites), *H19* (4 sites), *IGF2* (3 sites), and *LINE-1* (4 sites)] was consistent; therefore, average methylation level across CpGs for each gene was used in all analysis. The mean DNA methylation of *LINE-1*, *ESR1*, *IGF2*, *PPARA*, and *H19* was 79.97 (SD 3.22), 4.31 (0.96), 47.85 (3.68), 17.29 (2.66), and 55.60 (4.94), respectively. The mean DNA methylation of *LINE-1*, *ESR1*, *IGF2* or *PPARA* did not significantly vary by infant sex (Δ relative to males, *P* value for *t* test; $\Delta = -0.07$, *P* = 0.91; $\Delta = -0.13$, *P* = 0.52; $\Delta = 1.32$, *P* = 0.08; $\Delta = 0.22$, *P* = 0.68, respectively). However, the mean DNA methylation of *H19* was 2.53 percentage points higher (*P* = 0.01) in females relative to males.

Maternal EDC levels are related to cohort-wide infant DNA methylation

Several significant relationships were observed between maternal first trimester EDC levels and infant DNA methylation profiles, adjusting for child's sex, maternal pre-pregnancy BMI, maternal age, and urinary SG (Table 2). Interestingly, all of the significant associations were observed for phthalate metabolites, and these associations were all negative. For instance, a log unit increase in MCP was associated with a 0.60 (95%CI: -1.14, -0.06) percentage point decrease in *LINE-1* methylation and a 0.83 (95%CI: -1.52, -0.15) percentage point decrease in *IGF2* methylation. Similarly, MBzP, MCP and Σ DEHP were each negatively associated with *PPARA* methylation. For example, a log unit increase in Σ DEHP corresponded to 1.03 (95%CI: -1.83, -0.22) percentage point decrease in *PPARA* methylation with similar associations observed for MBzP and MCP (Table 2). Similar to the direction of effect for the urinary phthalate metabolites, there was a borderline significant inverse relationship between urinary BPA and *ESR1* methylation [-0.38 (95%CI: -0.78, 0.01)]. However, BPA measured in either urine or plasma was largely not associated (*P* > 0.05) with all other candidate gene nor *LINE-1* methylation.

Table 2. Absolute change in infant DNA methylation per log unit increase in maternal EDC exposure.

	LINE1			H19			IGF2			PPARA			ESR1		
	β	95%CI	<i>P</i> value	β	95%CI	<i>P</i> value	β	95%CI	<i>P</i> value	β	95%CI	<i>P</i> value	β	95%CI	<i>P</i> value
MBzP	0.12	-0.49, 0.73	0.7	0.44	-0.55, 1.44	0.38	-0.34	-1.13, 0.45	0.4	-0.62	-1.22, -0.02	0.04*	0.03	-0.20, 0.26	0.79
MCP	-0.60	-1.14, -0.06	0.03*	-0.42	-1.34, 0.50	0.36	-0.83	-1.52, -0.15	0.02*	-0.83	-1.38, -0.28	0.004*	-0.02	-0.23, 0.19	0.85
MEP	-0.14	-0.65, 0.37	0.59	-0.08	-0.90, 0.74	0.85	-0.51	-1.13, 0.10	0.1	0.17	-0.32, 0.66	0.49	0.13	-0.05, 0.32	0.15
Σ DEHP	-0.61	-1.43, 0.21	0.14	0.53	-0.83, 1.90	0.44	-0.98	-2.12, 0.17	0.09	-1.03	-1.83, -0.22	0.01*	-0.07	-0.53, 0.22	0.69
Σ DBP	-0.64	-1.35, 0.08	0.079	0.02	-1.15, 1.18	0.98	-0.24	-1.53, 0.68	0.61	-0.02	-0.75, 0.70	0.95	0.1	-0.17, 0.37	0.47
Plasma BPA	0.19	-0.34, 0.72	0.48	0.34	-0.06, 0.75	0.097	-0.02	-0.59, 0.56	0.95	0.11	0.34, 0.56	0.63	0.08	-0.06, 0.22	0.28
Urine BPA	0.55	-0.11, 1.22	0.1	1.08	-0.51, 2.67	0.18	-0.22	-1.22, 0.77	0.65	0.05	-0.65, 0.75	0.88	-0.38	-0.78, 0.01	0.05

**P* < 0.05, adjusted for specific gravity, maternal BMI, maternal age and child's sex.

Infant methylation patterns are correlated with candidate gene expression

Due to the observed hypomethylation in infant *PPARA* methylation with increasing concentrations of several maternal phthalate metabolites, RT-qPCR was performed to determine if *PPARA* methylation corresponded with altered levels of gene expression. A subset of MMIP cohort RNA samples ($n = 50$) was available for expression analysis and also had a DNA methylation value that met stringent quality assurance standards. Figure 1 depicts a plot of the unadjusted DNA methylation and gene expression data along with the associated r and P values for the Spearman Rank Correlation Test. DNA methylation levels at *PPARA* inversely correlate with *PPARA* gene expression ($r = -0.34$, $P = 0.02$).

Sex-stratified analysis reveals sex-specific EDC effects

In sex-stratified analyses sex-specific BPA-related effects were observed for multiple gene regions. Urinary BPA was associated with a 1.35 (95%CI: -2.69 , -0.01) percentage point decrease in *IGF2* methylation and a 1.22 (95%CI: -2.27 , -0.16) percentage point decrease in *PPARA* methylation in

females (Table 3a) but not males (Table 3b). Female-specific inverse associations were also observed between MEP and *IGF2* as well as MCPP and *ESR1* (Table 3a). In contrast to the female-specific inverse associations, plasma BPA and urinary BPA in males tended to be positively related to LINE-1 and *ESR1* (plasma BPA only), but these relationships were only borderline significant (Table 3b).

Discussion

Recently, considerable emphasis has been placed on determining the contribution of early life exposures to the rising tide of obesity [44]. However, there have been few epidemiological studies focused on assessing the epigenetic impact of exposures to EDCs in early pregnancy within the context of growth and metabolic programming. This may in part be due to the difficulty of recruiting women in the first trimester of pregnancy and highlights a unique feature of the MMIP cohort. Since EDCs can profoundly impact developmental pathways during critical windows of development and potentially lead to adult disease, identifying differential epigenetic profiles that act not only as biomarkers of exposure but also as potential markers of disease susceptibility is of great importance.

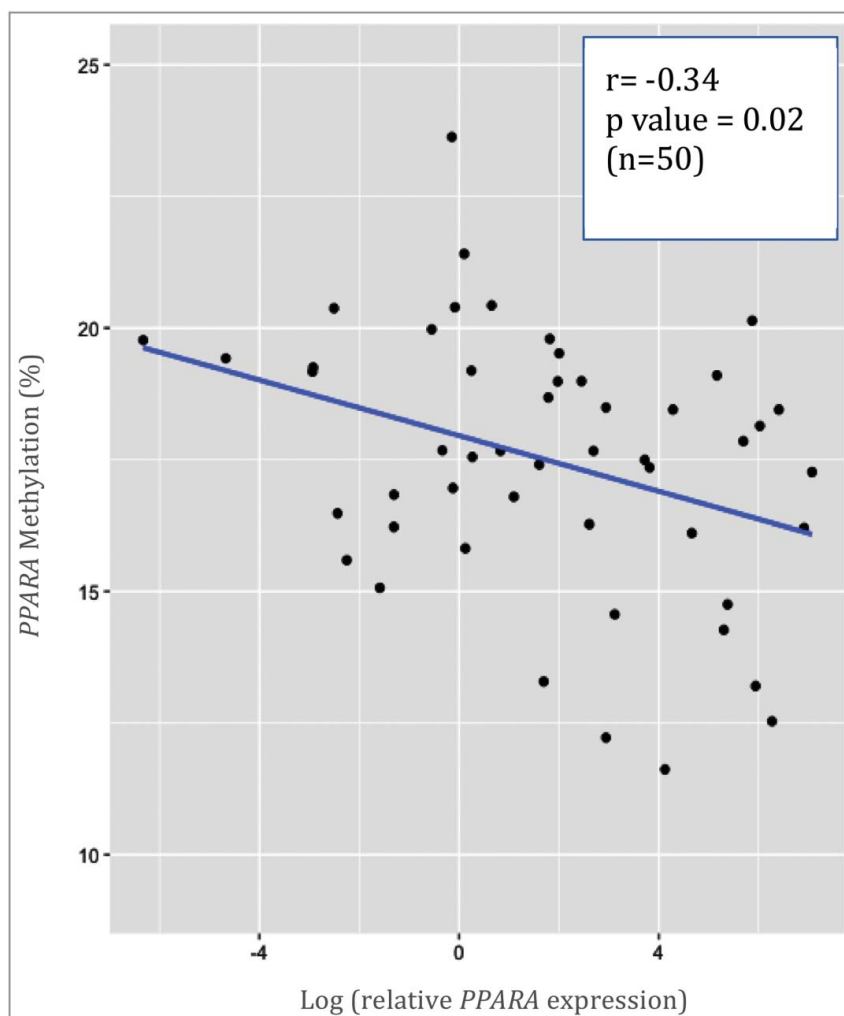


Figure 1. Correlation between *PPARA* gene expression and *PPARA* DNA methylation in infant cord blood. The unadjusted data are plotted, and the associated r and p values for the Spearman Rank Correlation Test are in upper right. Here, we show that *PPARA* expression is inversely correlated with *PPARA* methylation.

Table 3a. Absolute change in infant DNA methylation per log unit increase in maternal EDC exposure (females).

	LINE1			H19			IGF2			PPARA			ESR1		
	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value
MBzP	-0.07	-0.99, 0.85	0.88	0.94	-0.81, 2.70	0.28	-0.56	-2.03, 0.92	0.45	-0.36	-1.30, 0.58	0.44	-0.11	-0.30, 0.08	0.26
MCP	-0.88	-1.84, 0.07	0.07	-0.55	-2.38, 1.28	0.55	-1.75	-3.11, -0.39	0.01	-1.12	-2.00, -0.25	0.01	-0.23	-0.42, -0.04	0.02*
MEP	-0.24	-1.05, 0.56	0.54	-0.16	-1.70, 1.38	0.83	-1.32	-2.39, -0.24	0.02*	-0.29	-1.08, 0.50	0.46	0.02	-0.14, 0.18	0.81
Σ DEHP	-0.76	-2.05, 0.52	0.24	0.99	-1.42, 3.4	0.41	-1.86	-4.09, 0.37	0.1	-0.84	-2.14, 0.46	0.2	-0.16	-0.48, 0.16	0.32
Σ DBP	-0.91	-2.03, 0.20	0.11	0.75	-1.37, 2.88	0.48	-0.45	-2.24, 1.33	0.61	-0.18	-1.30, 0.94	0.75	-0.13	-0.37, 0.12	0.31
Plasma BPA	-5.40	-2.21, 1.14	0.51	-1.19	-1.40, 1.01	0.74	-0.10	-2.13, 1.94	0.92	0.8	-0.32, 1.92	0.15	-0.05	-0.32, 0.22	0.7
Urine BPA	-0.46	-1.60, 0.67	0.41	-0.19	-3.48, 3.09	0.9	-1.35	-2.69, -0.01	0.05*	-1.22	-2.27, -0.16	0.03*	0.11	-0.15, 0.36	0.38

* $P < 0.05$ only when stratified; adjusted for specific gravity, maternal BMI, maternal age.

Using a hypothesis driven candidate gene approach focused on genes related to growth and development as well as the repetitive element, LINE-1, we observed hypomethylation of LINE-1, *IGF2*, and *PPARA* with increasing phthalate concentrations. Two individual phthalate metabolites (MBzP and MCP) as well as the sum of DEHP metabolites were inversely associated with *PPARA* methylation, and DNA methylation was inversely correlated with *PPARA* gene expression, providing evidence of functional relevance. Additionally, a sex-stratified analysis of EDCs and DNA methylation showed that some relationships, including the association of urinary or plasma BPA with methylation, were female-specific.

The negative relationship observed between *PPARA* methylation and Σ DEHP is particularly notable. In addition to being one of the most commonly identified EDCs in the US population [45], observed changes in DNA methylation and expression in adult rats resulting from *in utero* Σ DEHP exposure have raised speculation that the PPAR pathway, and specifically *PPARA*, may participate in epigenetic disruption [46]. The findings from the present study support this hypothesis, as human *PPARA* promoter was inversely associated with Σ DEHP levels. Moreover, epigenetic modification in this region was also correlated with *PPARA* expression in infant cord blood DNA. Given that *PPARA* is a master regulator of lipid metabolism in the liver and predominant PPAR subtype in this organ [47], the identification of a functionally relevant region measured in a surrogate easily obtainable matrix would be extremely valuable as a biomarker, provided it reflects the profile of the liver. Other studies have found that rats perinatally exposed to BPA displayed increased *PAPRA* activation in the liver [48], but this group did not evaluate *PPARA* expression simultaneously in blood. Moving forward, studies investigating *PPARA* expression should consider investigating matched tissue specimens (e.g., liver and a bioavailable tissue) as well as address the issue of cellular heterogeneity, which we did not

account for here. Together, these future studies will aid in determining the biological relevance of our findings.

Our findings in a human birth cohort characterized for EDC levels in maternal first trimester urine and blood adds to a relatively sparse, but growing, body of data on prenatal phthalate-induced epigenetic effects [20]. The finding that MCP had a consistent negative impact on 3 of the 5 candidate loci evaluated, including methylation patterns of the imprinted gene *IGF2* is consistent with the findings of impact from phthalate exposure. A study with 196 women from the Boston, MA area found first trimester MCP had an impact on epigenetic patterns of imprinted genes *IGF2* and *H19* in the placenta [49]. When the sum of all phthalates or the sum of low molecular weight phthalates were considered, both were found to be inversely associated with *IGF2* DMR methylation [49]. Findings from the present study also revealed an inverse relationship between MCP and methylation of LINE-1 repetitive elements. Aberrant repetitive element methylation patterns can impact genome stability and gene expression [50], and previous reports have suggested these repetitive loci are targets of developmental phthalate exposure. For instance, increasing concentrations of MEP in early and late pregnancy in the CHAMACOS cohort were found to be inversely associated with *Alu* methylation in infant cord blood [29]. Zhao et al. found that maternal third trimester levels of Σ DEHP were inversely associated with LINE-1 measured in DNA collected from pooled samples from 4 within individual placental biopsies [51]. Solomon et al. were recently the first to assess epigenome-wide effects in humans from prenatal phthalate exposure. Using the Infinium 450K BeadChip they identified 27 differentially methylated regions (DMRs) in infant cord blood of which more than half of the DMRs were associated with Σ DEHP and some DMRs were found in genes involved in endocrine function [52]. Taken

Table 3b. Absolute change in infant DNA methylation per log unit increase in maternal EDC exposure (males).

	LINE1			H19			IGF2			PPARA			ESR1		
	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value	β	95%CI	P value
MBzP	0.42	-0.41, 1.26	0.31	-0.05	-1.03, 0.94	0.93	0.01	-0.80, 0.82	0.98	-0.77	-1.58, 0.04	0.06	0.08	-0.33, 0.49	0.7
MCP	-0.41	-1.05, 0.24	0.21	-0.16	-0.98, 0.66	0.69	-0.18	-0.85, 0.48	0.58	-0.44	-1.20, 0.31	0.25	0.05	-0.28, 0.39	0.75
MEP	-0.02	-0.69, 0.64	0.95	-0.06	-0.82, 0.70	0.88	0.27	-0.36, 0.89	0.4	0.44	-0.18, 1.05	0.16	0.22	-0.09, 0.53	0.16
Σ DEHP	-0.83	-2.00, 0.34	0.16	0.02	-1.40, 1.44	0.98	-0.07	-1.27, 1.13	0.9	-0.79	-1.97, 0.39	0.18	-0.32	-0.93, 0.29	0.29
Σ DBP	-0.50	-1.46, 0.47	0.31	-0.83	-1.93, 0.27	0.13	0.04	-0.87, 0.95	0.92	0.3	-0.68, 1.29	0.54	0.2	-0.26, 0.65	0.39
Plasma BPA	0.55	-0.02, 1.11	0.06	0.42	-0.08, 0.91	0.09	-0.01	-0.50, 0.47	0.97	-0.15	-0.70, 0.40	0.59	0.18	-0.003, 0.37	0.05
Urine BPA	0.6	-0.04, 1.24	0.07	0.74	-0.17, 1.64	0.11	0.59	-0.37, 1.55	0.21	0.21	-0.46, 0.88	0.51	-0.46	-1.05, 0.13	0.12

* $P < 0.05$ only when stratified; adjusted for specific gravity, maternal BMI, maternal age.

together, the prenatal epigenomic landscape, from gene to genome-wide, appears to be sensitive to several different phthalates or phthalate mixtures.

Because it is well known that EDCs can have sexually dimorphic phenotypic effects, we also performed a sex-stratified analysis for DNA methylation. Interestingly, we found female-specific EDC-induced effects, which were consistently negative. Sex-specific BPA-induced epigenetic effects have been observed previously in human cohorts [24] and rodent models [26–28]. In some instances the sexually dimorphic BPA-induced epigenetic changes have been linked to metabolic phenotypes. For example, Anderson et al. showed that BPA-induced epigenetic effects in female mice mediated metabolic outcomes including body weight and body fat [53]. Although the sex-specific findings from the present study should be considered preliminary in view of the small sample size, they highlight the need to include both sexes in animal and human studies when evaluating EDC-related effects.

A logistical challenge for epidemiological studies is precisely capturing and quantifying environmental exposures. Relating infant cord blood DNA methylation changes to maternal EDC levels in spot urine sample is one limitation of our study. However, while urinary EDC concentrations may vary throughout pregnancy [54,55], it has been suggested that a single urinary measurement may reasonably represent several months of maternal exposure and thus potential fetal exposure [56]. By contrast, a major strength of our study is the exposure assessment of EDCs in the first trimester of pregnancy as this is a critical window of sexual differentiation and a period when perturbation of normal and necessary epigenetic patterns is particularly impactful. Due to the number of comparisons made, we note that the possibility of chance findings in our assessment of the effect of early EDC exposures on candidate gene DNA methylation cannot be ruled out. Finally, the potential biological relevance of such small changes in DNA methylation to offspring health outcome remains to be elucidated; a recent review of literature suggests that such small epigenetic changes during critical windows of differentiation have the potential to induce short and long term gene expression changes [57].

Conclusion


Results from this study further contribute to a growing body of evidence suggesting that *in utero* exposure to EDCs can perturb normal biological functions via the epigenome. The epigenetically labile regions described here, along with others yet to be identified, may act as biomarkers of disease susceptibility if they are linked to childhood or adult phenotypes in subsequent studies. There is a great need to discover such biomarkers and exploit them in an effort to break the cycle of intergenerational obesity and metabolic disorders.


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
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