



Exploring adduct formation between human serum albumin and eleven organophosphate ester flame retardants and plasticizers using MALDI-TOF/TOF and LC-Q/TOF

Shaogang Chu ^{a, b}, Margaret R. Baker ^a, Gladys Leong ^a, Robert J. Letcher ^b, Shirley J. Gee ^c, Bruce D. Hammock ^c, Qing X. Li ^{a,*}

^a Department of Molecular Biosciences and Bioengineering, University of Hawaii at Manoa, 1955 East West Road, Honolulu, HI, 96822, USA

^b Ecotoxicology and Wildlife Health Division, Wildlife and Landscape Science Directorate, Science and Technology Branch, National Wildlife Research Centre, Environment and Climate Change Canada, 1125 Colonel Bay Dr., Ottawa, ON, K1A 0H3, Canada

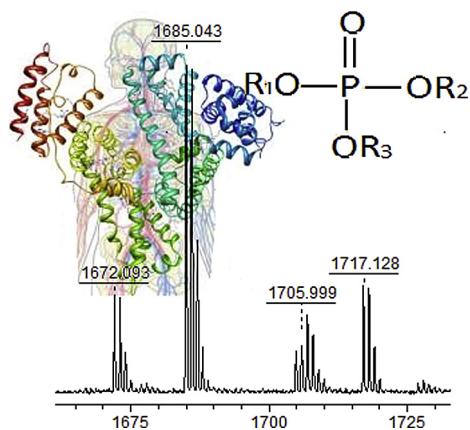
^c Department of Entomology and Nematology, University of California at Davis, One Shields Avenue, Davis, CA, 95616-858, USA



HIGHLIGHTS

- Adduct formation between HSA and OPE flame retardants and plasticizers is explored.
- Profenofos-HSA adduct preparation and analysis verify the methodology.
- L₄₀₈VRY₄₁₁TKKVPQVST₄₂₀ (*m/z* 1685.043) is a HSA-profenofos adduct.
- No OPE-HSA adducts are observed.

GRAPHICAL ABSTRACT



ARTICLE INFO

Article history:

Received 23 December 2016

Received in revised form

8 March 2017

Accepted 28 March 2017

Available online 31 March 2017

Handling Editor: Shane Snyder

Keywords:

Human serum albumin

Protein adduct

Organophosphate ester

Flame retardant

ABSTRACT

Organophosphate (OP) and organophosphate ester (OPE) adducts of albumin are valuable biomarkers for retrospective verification of exposure. In the present study, our goal was to determine whether OPE flame retardants (OPE FRs) and OPE plasticizers can covalently bind to human serum albumin (HSA), which would allow the resulting adducts to be used to evaluate exposure. Eleven OPE FRs and plasticizers were examined in a HSA-adduct *in vitro* assay. Pure HSA was incubated with the target OPEs, as well as with an OP insecticide (profenofos) positive control. After enzymatic cleavage with pepsin or Glu-C, the digested albumin was analyzed by matrix-assisted laser desorption ionization tandem time-of-flight mass spectrometry (MALDI-TOF/TOF-MS) and liquid chromatography-quadrupole time-of-flight mass spectrometry (LC-Q-ToF-MS). Under optimized HSA assay conditions, tyrosine adducts were formed at Y₄₁₁ and Y₁₄₈/Y₁₅₀ with a characteristic mass shift for phosphorylation ($\Delta m/z$ 166) for the profenofos positive control. However, no such phosphorylated peptides were detected for the 11 target OPEs. This

* Corresponding author.

E-mail address: qing@hawaii.edu (Q.X. Li).

negative result suggests that these OPEs have very different affinities from the OP insecticide. They are less reactive or they may specifically interact with other proteins.

© 2017 Elsevier Ltd. All rights reserved.

1. Introduction

Organophosphates (OPs) are a large class of chemicals, and organophosphate esters (OPEs) are widely used as insecticides, herbicides, flame retardants, plasticizers, and even as toxic nerve agents (Reemtsma et al., 2008; Thompson et al., 2010; van der Veen and de Boer, 2012). OPE flame retardants (OPE FRs) and plasticizers are high production-volume chemicals used in a variety of industries, including plastics, furniture, textile, electronics, construction, vehicle, and petroleum industries (Marklund et al., 2005; Reemtsma et al., 2008; Wei et al., 2015; van der Veen and de Boer, 2012; Wang et al., 2010). In 2001, worldwide usage of OPE FRs was estimated at 186,000 tons (Hartmann et al., 2004). OPE FRs used in Western Europe accounted for about 91,000 tons per year in 2006, and there was an increase of 2.5% in 2001–2005 and 7.1% in 2005–2006 (Reemtsma et al., 2008). In China, annual production of OPE FRs reached approximately 70,000 tons in 2007 and is predicted to increase by 15% annually (Wang et al., 2010). The EPA reported that up to 50 million lbs of chlorinated OPE FRs were produced or imported into the United States in 2012 (EPA, 2015). North American annual production of tris(2-chloroisopropyl) phosphate (TCIPP), tris(1,3-dichloro-2-propyl) phosphate (TDCIPP), and tris(2-chloroethyl) phosphate (TCEP) increased from less than 14,000 tons per year in 1986 to 38,000 tons per year in 2012 (Schreder et al., 2016). Production and usage of OPE FRs may even increase due to restrictions and regulations on the use of polybrominated diphenyl ether (PBDE) flame retardants.

The wide array of OP structural types, functional groups and substitutions results in dramatic differences in their chemical, physical and biochemical properties. Usage of OPEs as flame retardants and plasticizers are of increasing concern because of their potential adverse effects on the environment and human health (van der Veen and de Boer, 2012; Wei et al., 2015). Most OPEs are used as additives instead of being chemically bonded to host materials and constitute 1%–30% of the total composition by weight (Morgan and Gilman, 2013). Therefore, they are constantly released into the environment via volatilization, abrasion and leaching from host materials during their lifetime. A growing number of reports has shown the detection and measurement of OPE FRs and OPE plasticizers in various environmental samples including indoor and outdoor air, indoor dust, water, sediments, soils and landfill leachates, biota samples and human plasma and breast milk (Marklund et al., 2003, 2005; Hartmann et al., 2004; Reemtsma et al., 2008; Mercier et al., 2011; van der Veen and de Boer, 2012; McGoldrick et al., 2014; Brommer and Harrad, 2015; Hallanger et al., 2015; Wei et al., 2015; He et al., 2016; Greaves et al., 2016a).

Compared to OP insecticides, knowledge is limited on the toxicity and biological effects of OPE FRs and OPE plasticizers, as was recently reviewed by Wei et al. (2015) and Greaves and Letcher (2016b). Among these OPEs, some chlorinated OPE FRs, such as TCEP and TDCIPP are suspected carcinogens (EPA, 2015). Animal studies have shown that OPE elicited effects on kidneys, liver and neurological systems, while thyroid effects and developmental and reproductive toxicity were more variable across studies (EPA, 2015; Wei et al., 2015). Neurotoxic effects have been observed for TCEP, tri-*n*-butyl phosphate (TNBP) and triphenyl phosphate (TPHP) (Meeker and Stapleton, 2010). Some compounds, such as TPHP, are

suspected of being sensitizers for allergies (Camarasa and Serrabaldrich, 1992) and are potent human blood monocyte carboxylesterase inhibitors (Wei et al., 2015). However, paradoxical conclusions were reported on toxicity of OPE FRs and OPE plasticizers, and these mostly resulted from a problem in quantitative risk assessment and a number of data gaps (de Ree et al., 2014).

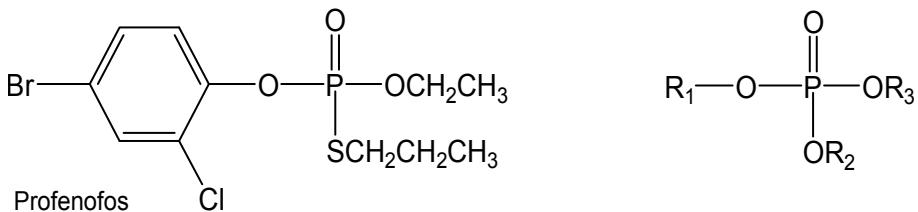
Various enzymatic studies show that OP insecticides inhibit acetylcholinesterase (AChE) (McDonough and Shih, 1997). However, there is still a debate on the toxic effects and current thought suggests that OP toxicity cannot be attributed entirely to inhibition of AChE (Casida and Quistad, 2004; EPA, 2015). Recently, investigators have begun to investigate covalent OP adduction of non-AChE protein targets as a possible causative step in other toxic responses (Thompson et al., 2010; Schopfer et al., 2010; Marsillach et al., 2013; Crow et al., 2014). Serum albumin was found to form a covalent bond with OPs (Casida and Quistad, 2004; Peebles et al., 2005; Schopfer et al., 2010; Crow et al., 2014) and these covalent adducts with blood proteins might provide a new biomarker of OP exposure (Li et al., 2007; Williams et al., 2007; John et al., 2008, 2010; Crow et al., 2014). In contrast to the original OPs that often have quite short half-lives *in vivo*, OP adducts to albumin are relatively stable and more resistant to oxime therapy (Peebles et al., 2005).

Human serum albumin (HSA) adducts can be used as a biomarker for OP insecticides and chemical warfare agent exposure (Li et al., 2007; Schopfer et al., 2010; John et al., 2010; Crow et al., 2014). However, to our knowledge, there is no report in the literature on phosphorylation of HSA by OPE FRs and OPE plasticizers. The objective of the present study was to determine whether common commercial OPE FRs and OPE plasticizers can covalently form adducts with HSA via nucleophilic sites similar to that reported for OP insecticides and chemical warfare agents. In order to achieve sensitive and reproducible results in the present study, the conditions for the specific analysis of stable OPE-HSA adducts formed *in vitro* were optimized with an OP insecticide as the positive control assay. The results, both positive and negative, were confirmed by matrix-assisted laser desorption ionization tandem time-of-flight mass spectrometry (MALDI-TOF/TOF-MS) and liquid chromatography-quadrupole time-of-flight mass spectrometry (LC-Q-TOF-MS).

2. Experimental

2.1. Materials

Table 1 shows the chemical structures, names and abbreviations of OPE FRs and OPE plasticizers that were examined in the present study. TCEP, TNBP, TPHP, tripropyl phosphate (TPP), tris(2-butoxyethyl) phosphate (TBOEP), tris(2-ethylhexyl) phosphate (TEHP), triethyl phosphate (TEP) and 2-ethylhexyl-diphenyl phosphate (EHDPP) were purchased from Sigma-Aldrich (St. Louis, MO, USA). TDCIPP and tri(methylphenyl) phosphate (TMPP) were purchased from TCI America (Portland, OR, USA). Tris(2-chloroisopropyl) phosphate (TCIPP) was purchased from AK Scientific (Union City, CA, USA). The organophosphate insecticide profenofos, which was used as a positive control and to optimize the methods, was purchased from Sigma-Aldrich and its molecular

Table 1Profenofos and 11 organophosphate ester (OPE) flame retardants and plasticizers.²

| No. | Chemical names | Abbreviation | Substituents | CAS |
|-----|---|--------------|--|------------|
| 1 | Tris(2-chloroethyl) phosphate | TCEP | $R_1 = R_2 = R_3 = -CH_2CH_2Cl$ | 115-96-8 |
| 2 | Tripropyl phosphate | TPP | $R_1 = R_2 = R_3 = -CH_2CH_2CH_3$ | 513-08-6 |
| 3 | Tris(2-chloroisopropyl) phosphate | TCIPP | $R_1 = R_2 = R_3 = -CH(CH_3)CH_2Cl$ | 13674-84-5 |
| 4 | Tris(1,3-dichloro-2-propyl) phosphate | TDCIPP | $R_1 = R_2 = R_3 = -CH(CH_2Cl)_2$ | 13674-87-8 |
| 5 | Triphenyl phosphate | TPHP | $R_1 = R_2 = R_3 = -C_6H_5$ | 115-86-6 |
| 6 | Tri-n-butyl phosphate | TnBP | $R_1 = R_2 = R_3 = -CH_2(CH_2)_2CH_3$ | 126-73-8 |
| 7 | Tris(methylphenyl) phosphate ^a | TMPP | $R_1 = R_2 = R_3 = -C_6H_4CH_3$ | 1330-78-5 |
| 8 | Tris(2-butoxyethyl) phosphate | TBOEP | $R_1 = R_2 = R_3 = -CH_2CH_2OCH_2(CH_2)_2CH_3$ | 78-51-3 |
| 9 | Tris(2-ethylhexyl) phosphate | TEHP | $R_1 = R_2 = R_3 = -CH_2CH(C_2H_5)CH_2(CH_2)_2CH_3$ | 78-42-2 |
| 10 | Triethyl phosphate | TEP | $R_1 = R_2 = R_3 = -CH_2CH_3$ | 78-40-0 |
| 11 | 2-Ethylhexyl-diphenyl phosphate | EHDPP | $R_1 = -CH_2CH(C_2H_5)CH_2(CH_2)_2CH_3$ $R_2 = R_3 = -C_6H_5$ | 1241-94-7 |

^a Mixture of *o*, *p*, *m* isomers.

structure was shown in Table 1. HSA (fatty acid free, purity > 99%), endoproteinase Glu-C from *Staphylococcus aureus* V8 (Glu-C, sequencing grade) and pepsin from porcine gastric mucosa were purchased from Sigma–Aldrich. α -Cyano-4-hydroxycinnamic acid (HCCA) and peptide calibration standard II were from Bruker (Billerica, MA, USA). DL-dithiothreitol (DTT) was from Acros Organics (New Jersey, USA) and iodoacetamide (IAA) was from BioRad (Hercules, CA, USA). Millipore ziptips C18 (10 μ L pipette tip) and VivaspinTM 500 centrifugal concentrators VS0102 (cut-off 10 kDa) were obtained from Sigma–Aldrich and Sartorius Stedim North America Inc. (New York, USA), respectively. HPLC grade acetonitrile, ammonium hydroxide (NH_4OH), ammonium bicarbonate (NH_4HCO_3) and trifluoroacetic acid (TFA, >99% pure) were from Fisher Scientific (Waltham, MA, USA). Water was purified on a Milli-Q Advantage A10 system (Millipore, Billerica, MA, USA).

2.2. Sample preparation and digestion

Individual standard solutions of profenofos and the target OPEs were diluted to 2 mg/mL with acetonitrile and kept in a refrigerator at 4 °C until further use. Fatty acid-free HSA solution was prepared at a concentration of 0.5 mg/ml in water. An aliquot of 10 μ L of the OPE acetonitrile solution (2 mg/mL) was added to 100 μ L of fresh HSA water solution (0.5 mg/ml). This mixture was incubated at 37 °C under gentle shaking for 24 h. The incubation mixtures were then subjected to ultrafiltration using a Vivaspin 500 centrifugal concentrator, which was pre-washed three times with 400 μ L of water, for a period of 5 min at 14,000 \times g to concentrate the protein and separate unreacted OPE from the mixture solution. After the filtrate was discarded, the HSA residue was washed seven times with 400 μ L of water. Concentrated HSA fractions (about 25 μ L) were transferred into a vial. The concentrator was rinsed with 25 μ L of water and the rinsate was combined with the HSA fractions. This adduct assay HSA solution (about 50 μ L) was stored at –20 °C until further processing.

Prior to digestion, 25 μ L of 4 mM DTT in 50 mM ammonium bicarbonate buffer (pH 8.2) was added to 25 μ L of the adduct assay solution and incubated at 55 °C for 40 min to rupture the disulfide bonds in HSA. After cooling, 12.5 μ L of 55 mM IAA in 25 mM

ammonium bicarbonate buffer was added and the solution was incubated in the dark for 1 h at room temperature to alkylate the free cysteine groups generated in HSA. A 2.5 μ L of 150 mM DTT in 50 mM ammonium bicarbonate buffer was then added into the solution to remove any remaining IAA, followed by digestion of the modified HSA by endoproteinase Glu-C or pepsin.

Endoproteinase Glu-C digestion was done by mixing 7.5 μ L of 250 μ g/mL Glu-C solution in 25 mM ammonium bicarbonate buffer with the modified HSA solution and then incubated at 37 °C for 17 h.

Prior to pepsin digestion, the pH of the modified HSA solution was adjusted to 2–3 by addition of 20 μ L of 100 mM HCl. HSA was digested with 3 μ L of 250 μ g/mL pepsin in 10 mM HCl for 17 h at 37 °C.

After digestion (with Glu-C or pepsin), the sample was concentrated to about 20 μ L using a centrifugal vacuum concentrator (Eppendorf Vacufuge plus). The sample was then desalted with a ZipTip-C18 pipette tip, according to the manufacturer's instructions.

A blank control sample of HSA was prepared in the same manner as described above using aqueous acetonitrile solution as a test solution for each batch of samples. In this study all assays were performed in triplicate.

2.3. MALDI-TOF/TOF-MS analysis

The desalted HSA digest samples were analyzed with MALDI-TOF/TOF-MS and HCCA was used as the MALDI matrix. HCCA was dissolved at a concentration of 10 mg/mL in a solvent mixture of 50% acetonitrile, 47.5% water and 2.5%TFA. A 1 μ L aliquot of the HSA digest sample was applied onto a stainless-steel target plate, air-dried, and overlaid with 1 μ L HCCA solution and allowed to air-dry at room temperature. MALDI-TOF-MS experiments were carried out with an UltraflexTM III MALDI-TOF/TOF mass spectrometer (Bruker, Billerica, MA) equipped with Smartbeam laser system. Mass spectra of the peptides were acquired in reflector positive mode. The following parameters were set for peptide detection: ion source 1, 25.08 kV; ion source 2, 21.72 kV; lens, 9.28 kV; reflector 1, 26.42 kV; reflector 2, 13.867 kV. For MS/MS experiments, LIFT mode

was used under the following parameters: ion source 1, 8.01 kV; ion source 2, 7.21 kV; lens, 3.61 kV; reflector 1, 29.50 kV; reflector 2, 13.91 kV; LIFT 1, 19.08 kV; and LIFT 2, 2.71 kV. The measurements were performed in a mass range from m/z 300–4000 Da. Four thousand laser shots were accumulated in 200 shot increments and the sum of all shots represented the final mass spectrum. External mass calibration was carried out with the peptide calibration standard II from Bruker. FlexControl software (v.3.4) was used for controlling the system and data were processed using FlexAnalysis (v.3.4) and BioTools (v.3.2), with a maximum mass tolerance of 100 ppm.

2.4. LC-TOF-MS analysis

The HSA digest samples were also analyzed with LC-TOF-MS to confirm the results from MALDI-TOF-MS analysis. The remaining sample solution was concentrated to dryness and then reconstituted in 50 μ L 5% aqueous acetonitrile containing 0.1% formic acid. The LC-TOF-MS analysis was performed on an Agilent 1200 LC system coupled to an Agilent 6520A quadrupole time-of-flight mass spectrometer (LC-Q-TOF-MS) system (Agilent Technologies, Mississauga, ON, Canada) with an electrospray ionization (ESI) source. A 10 μ L aliquot of sample was injected into the LC system. LC separation was performed on a Kinetex XB-C18 column (100 mm \times 2.0 mm, 1.7 μ m particle size) (Phenomenex Co., CA, USA) at 40 °C. The LC mobile phase consisted of A: 0.1% formic acid in water and B: 0.1% formic acid in acetonitrile. The mobile phase flow rate was 0.3 mL/min. The LC gradient started at 5% B, increasing to 30% B in 15 min, increasing to 40% in 24 min, increasing to 95% in 25 min and was held for 5 min. Thereafter, the mobile phase composition was returned to initial conditions and the column was allowed to equilibrate for 15 min prior to the next injection.

The mass spectrometer was operated in positive ion mode. The capillary voltage was 3500 V. The fragmentor and skimmer voltages were 170 V and 70 V, respectively. Nitrogen was used as drying and nebulizing gases. The gas temperature was 350 °C. Dry gas flow rate was 5 L/min. Nebulizer pressure was 20 psi. Full-scan data acquisition was performed by scanning from m/z 70 to m/z 3200. The TOF-MS was tuned and calibrated with mixture calibration solution from Agilent. The resolution of MS (TOF) was $>20,000$ at m/z 622.028413. Internal mass calibration was carried out with reference masses of 121.050873 and 922.009798. Data were acquired with Mass Hunter 6.0.

3. Results and discussion

3.1. Enzymatic cleavage of human serum albumin

In the *in vitro* assay, pure HSA was used as a model protein to assay for the possible formation of OPE-HSA adducts. HSA has a molecular mass of 66.5 kDa and is comprised of 585 amino acids in a single peptide chain (Wa et al., 2006). Fatty acid-free HSA was used because fatty acids and the OPE might bind to the same albumin domain and therefore fatty acids could hinder the binding of OPE. The relatively high molecular weight of HSA creates a challenge to accurately measure adducts at the protein level. Therefore, in such assays, HSA and other large proteins are usually digested by a proteolytic enzyme to give characteristic peptides that can be accurately analyzed with a mass spectrometer.

Trypsin digestion is commonly used in amino acid sequence analysis among many protein digestion methods. However, trypsin digestion (cleavage occurs at arginine and lysine at the C-terminus) is known to generate a product containing Y_{411} that is not detectable by MALDI-TOF-MS (Li et al., 2007; John et al., 2010). In

addition, some common enzymes used for protein digestion, such as pronase, are not suitable for this purpose because their products of protein digestion contain a large number of peptides at low masses (i.e., <500 Da), which cannot be identified by MALDI-TOF-MS due to overlapping signals from the matrix (Wa et al., 2006).

Therefore, in the present study Glu-C or pepsin was used to investigate HSA phosphorylation by OPEs. Pepsin (preferential cleavage occurs at F, W, or Y at the N-terminus) and Glu-C (preferential cleavage of the glutamyl bonds) are widely used peptidases (Thompson et al., 2010). Our experiment showed that using DTT and IAA to respectively reduce and alkylate the disulfide bonds of HSA could largely improve the result repeatability and increase the responses of peptides of interest on MALDI-TOF-MS, especially for Glu-C digestion. Our result was in agreement with some earlier reports (Wa et al., 2006), though in some reported assays the process was omitted (Li et al., 2007; John et al., 2010).

3.2. Determination of HSA phosphorylation by OPE FRs and OPE plasticizers using MALDI-TOF-MS

MALDI-TOF-MS is a quick and reliable method for identification

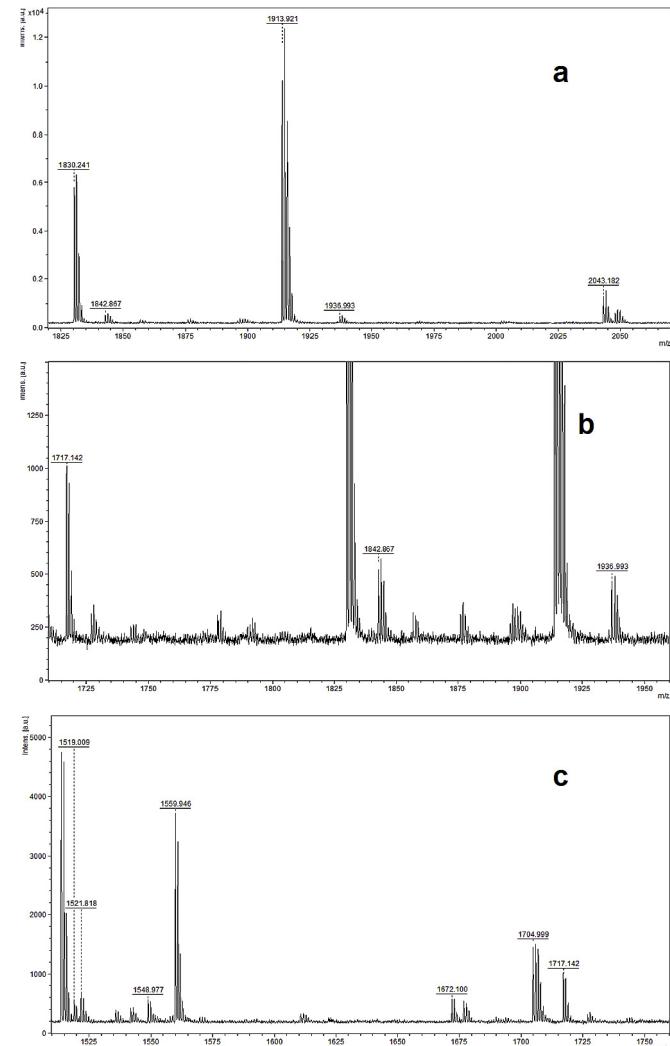


Fig. 1. MALDI-TOF mass spectra of pepsin-digested human serum albumin incubated with profenofos. Peptide adduct ions with a $\Delta m/z$ of 166 relative to the unmodified peptide were observed at (a) m/z 1996.3 originating from m/z 1830.2, (b) m/z 1883.2 originating from m/z 1717.1 and (c) m/z 1685.0 originating from m/z 1519.0.

of proteins and peptides and requires no chromatographic separation (Nomura, 2015). A major advantage of MALDI is that it can give a cleaner ion profile of peptide mixtures directly because of the absence of multiple charging. This relatively simple mass spectrum makes chemical identification easy and simple (Cristadoro et al., 2008; Thompson et al., 2010; Harvey, 2015).

Fig. 1(a, b, c) shows the MALDI-TOF MS spectra of phosphorylated target peptides in a sample of profenofos-treated HSA, digested with pepsin (positive control sample). In comparison with the blank control sample (spectrum not shown), there were three peaks with a mass increase of 166.021 Da and each of the three corresponded to a profenofos-HSA adduction at Tyr₄₁₁, with a mass tolerance of 100 ppm. Fig. 1a shows peaks corresponding to the peptide L₄₀₈VRY₄₁₁TKKVPQVSTPL₄₂₃ without (mono-isotopic ion at m/z 1830.222) and with (m/z 1996.259) addition of OP(OC₂H₅)(SC₃H₇) from profenofos. Similarly, Fig. 1b shows a mass spectrum of the peptide V₄₀₉RY₄₁₁TKKVPQVSTPL₄₂₃ without (m/z 1717.128) and with (m/z 1883.160) addition of profenofos. Another peptide, corresponding to the non-specific cleavage of HSA, L₄₀₈VRY₄₁₁TKKVPQVST₄₂₀, was observed in the mass spectrum without (m/z 1518.998) and with (m/z 1685.043) addition of profenofos (Fig. 1c). This additional ion gave more information about OPE adduction to HSA and has not been reported in the literature. This additional ion most likely resulted from the relatively long digestion incubation time. These results were confirmed by MALDI-TOF/TOF MS/MS and the mass increase (166.021 Da) of the relevant fragments ions clearly showed that profenofos covalently reacted with the Tyr₄₁₁ residue on HSA (Fig. 2), which was consistent with earlier studies (Li et al., 2007; John et al., 2010).

Eleven OPE FRs and OPE plasticizers (Table 1) were examined for OPE-HSA adduct formation along with the blank control sample and positive profenofos control sample. However, in contrast to the OP insecticide profenofos, no phosphorylated peptide products were found. Fig. 3(a, b, c) shows the MALDI-TOF-MS spectra obtained for pepsin-digested HSA after treatment with TPHP. If TPHP-HSA adducts had been formed, after pepsin digestion, their phosphorylated peptide ion would have been detected with MALDI-TOF-MS with ions of m/z 2062.11, 1949.03 and 1751.00 corresponding to the unphosphorylated peptide ions at m/z 1830.09, 1717.0 and 1519.0 (calculated m/z). For all other target OPEs, the MALDI-TOF-MS mass spectral results also showed no detectable phosphorylated peptides found for pepsin digested HSA after treatment with these compounds.

It was reported that relative to pepsin, Glu-C appeared to be a more suitable enzyme for catalyzing albumin cleavage and allowed

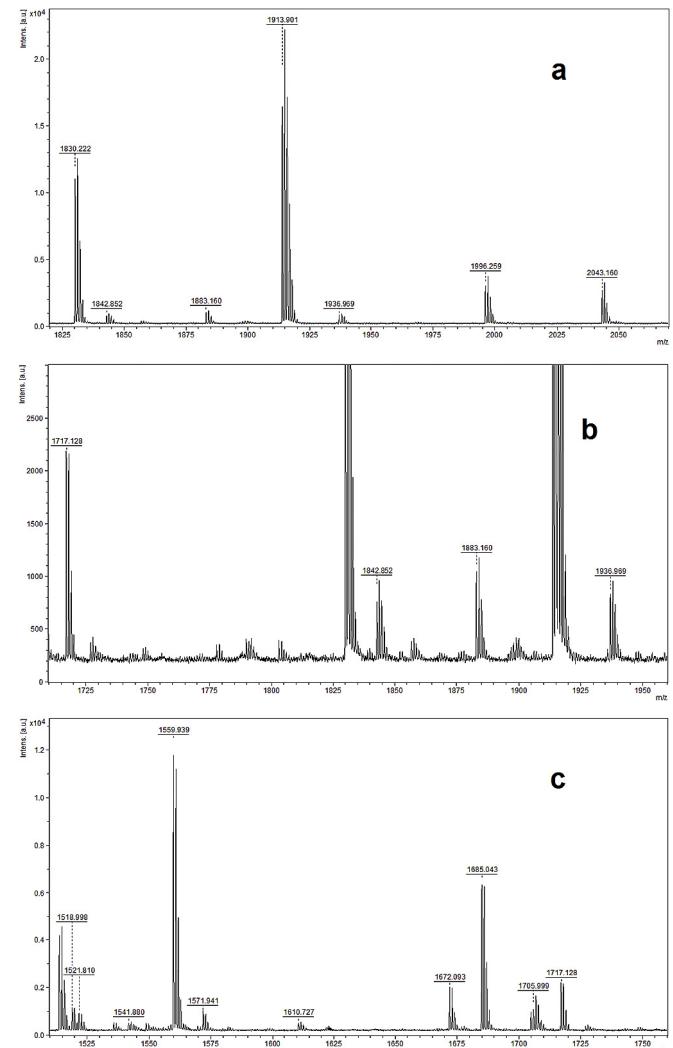


Fig. 3. MALDI-TOF mass spectra for pepsin-digested human serum albumin incubated with triphenyl phosphate (TPHP) showing the absence of adduct peptides at (a) m/z 2062.11 (b) m/z 1949.03 and (c) m/z 1751.00.

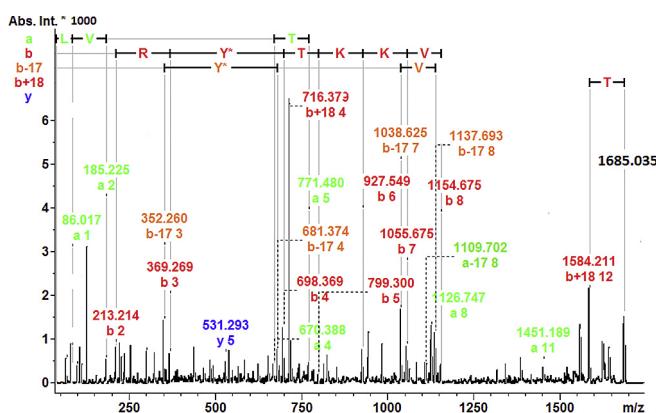


Fig. 2. MALDI-TOF/TOF MS/MS spectrum of the human serum albumin peptide L₄₀₈VRY₄₁₁TKKVPQVST₄₂₀, with adduction of profenofos at Tyr₄₁₁ (Y₄₁₁-OP(OC₂H₅)(SC₃H₇)).

for the detection of more adduct sites, specifically Y₁₄₈, Y₁₅₀ (I₁₄₂ARRHPY₁₄₈FY₁₅₀APE₁₅₃, at m/z 1519.8), Y₁₆₁ (L₁₅₄LFFAKRY₁₆₁-KAAFE₁₆₇, at m/z 1704.9) and Y₄₁₁ (Y₄₀₁KFQNALLVRY₄₁₁TKKVPQVSTPL₄₂₅, at m/z 2922.6) (John et al., 2010). Therefore, Glu-C digestion was also used to investigate possible phosphorylation of HSA by the target OPE FRs and OPE plasticizers under study. For the profenofos positive control sample, a mass increase ($\Delta m/z$ 166.021) indicated the profenofos adducted peptides Y₄₀₁KFQNALLVRY₄₁₁TKKVPQVSTPL₄₂₅ (m/z 2922.672 and 3088.702, in Fig. 4a) and I₁₄₂ARRHPY₁₄₈FY₁₅₀APE₁₅₃ (m/z 1519.770 and 1685.799, in Fig. 4b). However, consistent with the pepsin results already described, none of the OPE FRs and OPE plasticizers tested showed such phosphorylated peptides. Fig. 5a and b shows MALDI-TOF-MS spectra of HSA derived peptides treated with TPHP after Glu-C digestion. If there were TPHP-HSA adducts formed, their phosphorylated peptide ions would be detected by MALDI-TOF-MS analysis at m/z 3154.675 and 1751.810, corresponding to the unphosphorylated peptide ions at m/z 2922.646 and 1519.781 (calculated m/z).

We suspected the high concentration of acetonitrile might have affected OPE-HSA adduct formation. In a set of OPE-HSA adduct assay experiments, individual OPE compound solutions were

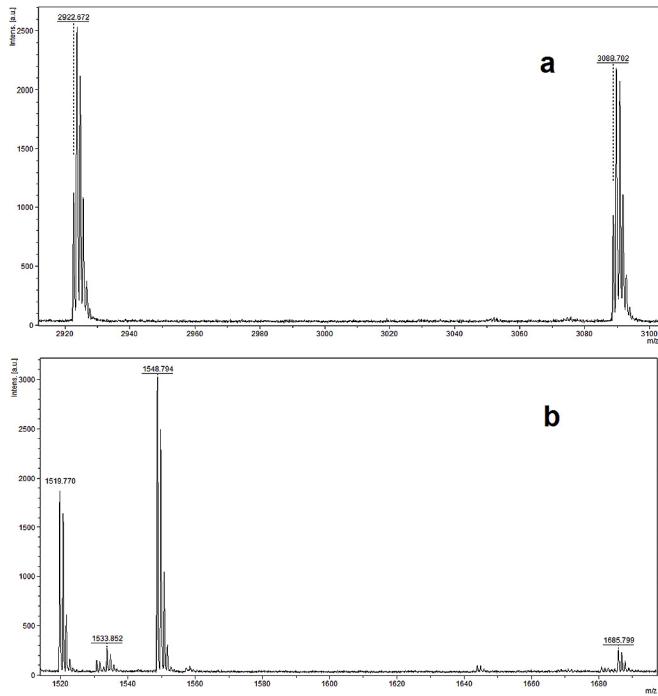


Fig. 4. MALDI-TOF mass spectra of human serum albumin incubated with profenofos followed by digestion with Glu-C, which show the adduct peptide ions with a $\Delta m/z$ of 166 at (a) m/z 3088.702 originating from m/z 2922.672 and (b) m/z 1685.788 originating from m/z 1519.770.

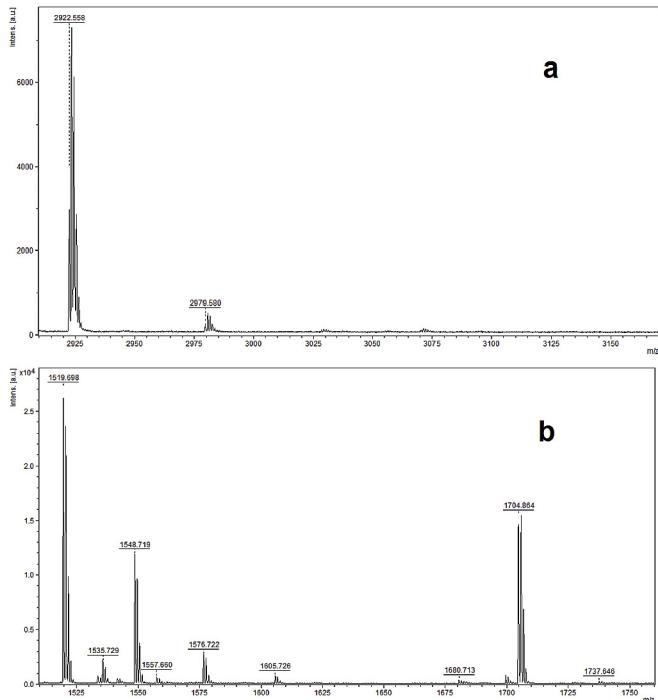


Fig. 5. MALDI-TOF mass spectra of human serum albumin incubated with triphenyl phosphate (TPHP) followed by digestion with Glu-C, which show the absence of the adduct peptide ions (a) at m/z 3154.675 and (b) at m/z 1751.810.

prepared with a concentration below its saturation limit in 0.1% aqueous acetonitrile, or at 2 mg/mL in pure water for OPEs that have high water solubility, such as TEP and TCEP, and then the

adduction assay was performed in the same way as described above. The OPE-HSA adduct assay results also showed that no phosphorylated peptides were detectable by MALDI-TOF MS, whereas in digestion samples of HSA incubated with profenofos (0.1% aqueous acetonitrile) the phosphorylated HSA-derived peptide ions were detectable. Therefore, the acetonitrile concentration was not a factor that affected adduct formation in the assay.

3.3. Determination of HSA phosphorylation by OPE FRs and OPE plasticizers using LC-Q-TOF-MS

To further confirm the MALDI-TOF-MS results, all OPE-HSA adduct assay samples were reanalyzed with LC-Q-TOF-MS. LC-Q-TOF-MS can detect multiply charged peptide ions with high mass resolution, high accuracy and relative high sensitivity. It has the advantage of being able to identify large-molecular-weight peptides and provides a versatile method for the detection of protein adducts (Thompson et al., 2010; Harvey, 2015).

Consistent with the MALDI-TOF-MS results, LC-Q-TOF-MS analysis revealed detection of multiply charged ions of phosphorylated peptides as a result of pepsin digestion of HSA samples incubated with the positive control profenofos. Fig. 6a shows the EIC of a phosphorylated peptide ion with multiple charge states at m/z 998.554 ($[M+2H]^{2+}$), m/z 666.040 ($[M+3H]^{3+}$), and m/z 499.783 ($[M+4H]^{4+}$). The molecular weight of the phosphorylated peptide (at retention time of 13.5 min) was calculated to be 1995.098 Da, corresponding to the profenofos adducted peptide $L_{408}VRY_{411}TKKVPQVSTPTL_{423}$ (theoretical mass 1995.100 Da). Fig. 6b shows the EIC of a phosphorylated peptide ion at m/z 942.017 ($[M+2H]^{2+}$), m/z 628.346 ($[M+3H]^{3+}$), and m/z 471.511 ($[M+4H]^{4+}$) in the same sample, which corresponded to the phosphorylated peptide $V_{409}RY_{411}TKKVPQVSTPTL_{423}$, with a calculated mass of 1882.017 Da (theoretical mass 1882.016 Da). Phosphorylated peptide ions at m/z 842.966 ($[M+2H]^{2+}$), m/z 562.313 ($[M+3H]^{3+}$) and m/z 421.986 ($[M+4H]^{4+}$) were detected in the same sample (Fig. 6c) and corresponded to the phosphorylated peptide $L_{408}VRY_{411}TKKVPQVST_{420}$ with a calculated mass of 1683.917 Da (theoretical mass 1683.916 Da). As with the pepsin digested sample, in the Glu-C digestion sample incubated with profenofos, the LC-Q-TOF-MS spectrum showed multiply charged ions of phosphorylated peptides, which had calculated molecular weights of 3087.653 Da (theoretical mass 3087.659 Da) and 1684.801 Da (theoretical mass 1684.794 Da).

In the present study, the signal-to-noise (S/N) ratios of analyte peaks on LC-QTOF-MS were much improved because peptides were separated by LC. For one of the profenofos-adducted peptide ions (at m/z 421.986), the S/N ratio was 294 (Fig. 6c). However, even at this high sensitivity, no peaks were detected in the corresponding EICs of HSA treated with the 11 OPE FRs and OPE plasticizers.

3.4. Explanations for no detectable HSA peptides phosphorylated by OPE FRs and OP plasticizers

Both LC-Q-TOF-MS and MALDI-TOF-MS analyses clearly showed the absence of any phosphorylation of HSA by the 11 OPE FRs and OPE plasticizers under study, despite the fact that these compounds have similar molecular structures with OP insecticides, such as profenofos and paraoxon-ethyl, which yielded detectable adducts of HSA (John et al., 2010). John et al. (2010) studied a series of toxic OP compounds including pesticides and nerve agents. Although some OP compounds have very similar molecular structures, their behaviours differed largely towards albumin phosphorylation (John et al., 2010). It is well known that HSA is phosphorylated preferentially at the nucleophilic Tyr_{411} located on the protein surface (John et al., 2008). Cleavage of the P–O bond is priority in the

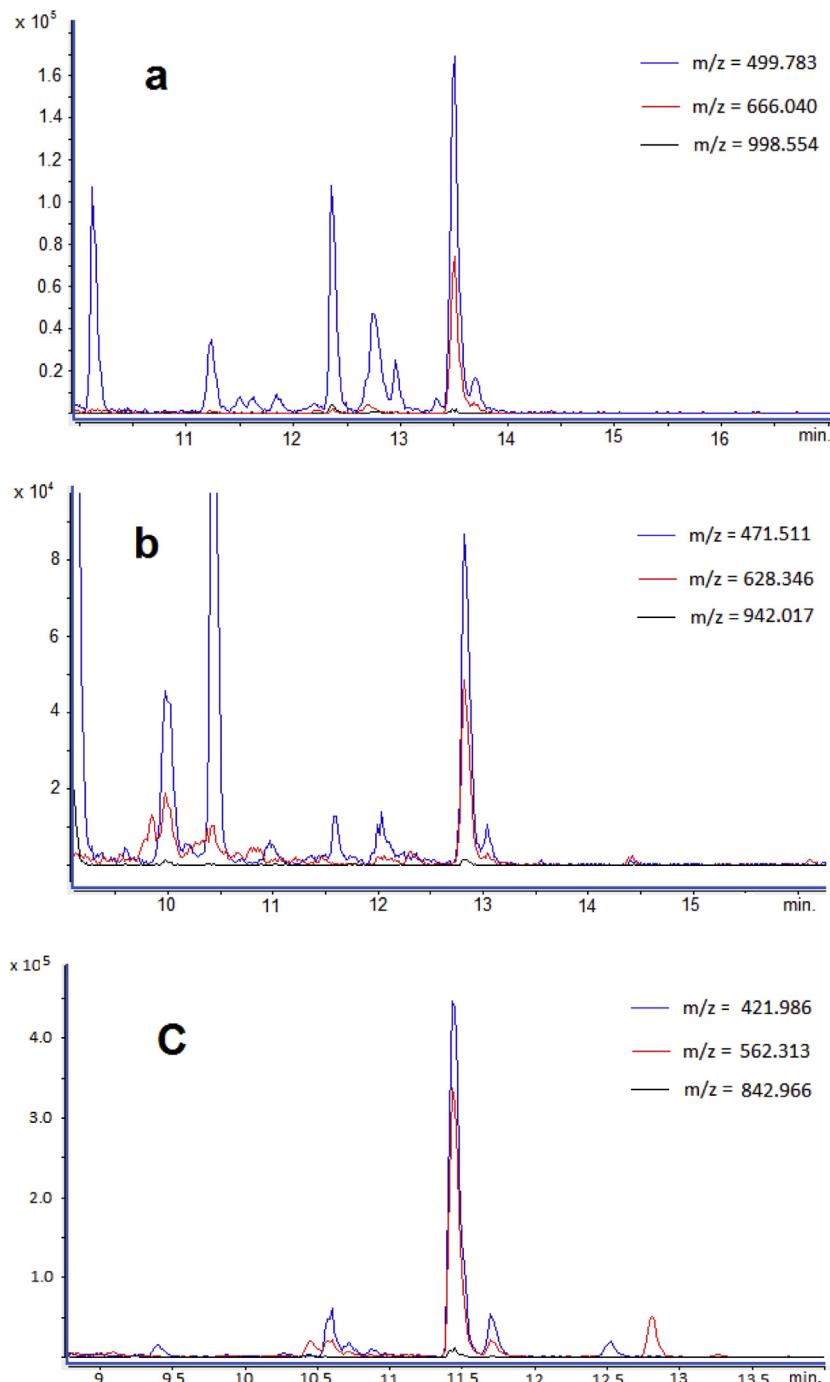


Fig. 6. Extraction ion chromatograms of phosphorylated peptide ions (a) multiply charged at m/z 998.554 ($[M+2H]^{2+}$), m/z 666.040 ($[M+3H]^{3+}$) and m/z 499.783 ($[M+4H]^{4+}$) corresponding a molecular weight of 1995.098 Da (1995.100 Da, theoretical value), (b) multiply charged at m/z 942.017 ($[M+2H]^{2+}$), m/z 628.346 ($[M+3H]^{3+}$) and m/z 471.511 ($[M+4H]^{4+}$) corresponding a molecular weight of 1882.017 Da (1882.016 Da, theoretical value) and (c) multiply charged at m/z 842.966 ($[M+2H]^{2+}$), m/z 562.313 ($[M+3H]^{3+}$) and m/z 421.986 ($[M+4H]^{4+}$) corresponding a molecular weight of 1683.917 Da (1683.916 Da, theoretical value).

process. Our results and those described in the literature indicated that HSA phosphorylation resulted from cleavage of the aryl phosphoester bond in profenofos and attachment of $OP(OC_2H_5)(SC_3H_7)$. If phosphorylation had instead occurred through breaking of the ethyl phosphoester bond, leading to attachment of $C_6H_3ClBrO-PO(SC_3H_7)$, the OPE-HSA adduct should have had a $\Delta m/z$ of 325.893. However, no such ion was detected by MALDI-TOF-MS and LC-ESI(+)TOF-MS. John et al. (2010) studied albumin phosphorylation by some OP pesticides and G- and V-type

nerve agents. They did not find albumin adducts resulting from cleavage of the alcohol chain from the target compounds. Comparing OP insecticides (e.g., profenofos, paraoxon-ethyl and chlorpyrifos-oxon) with the aryl OPE plasticizers TPHP and TMPP, these OP insecticides bear one aryl group with electron-withdrawing substituents at the *ortho*- or/and *para*-positions. Such good leaving-groups (more acidic conjugated acid) can result in easy cleavage of the aryl phosphoester bond to form adducts at Tyr₄₁₁ in HSA. However, although the OPE plasticizers TPHP and

TMPP have an aryl group, they do not have electron-withdrawing substituents, and therefore, TPHP and TMPP showed no such property.

This difference between OP insecticides and OPE FRs and OPE plasticizers can also be found in their hydrolysis behavior. Most OPE FRs and OPE plasticizers are relatively stable against hydrolysis at neutral pH. In a very recent paper by [Su et al. \(2016\)](#), over a period of 35 days at 20 °C at pH 7, 9 out of the 16 OPEs were stable and were not significantly hydrolyzed. For the other 7 OPEs, including TPHP and TMPP, their hydrolysis half-life ranged from 23 d to 112 d at pH 7. It was reported that hydrolysis half-lives at pH 7 for TEP and TPHP are in a range of 1.2–5.5 years ([Mabey and Mill, 1978](#)). However, profenofos hydrolysis half-lives were approximately 24–62 d at pH 7 and 15 d at pH 8 ([Zamy et al., 2004](#)). The results of the present study suggest that OPE molecular weight is not an important factor in HSA phosphorylation because there was no detectable phosphorylation for the low molecular weight TEP.

With little information about covalent adducts of OPs available, these explanations are far from adequately addressing this complex phosphorylation process. Covalent adduct formation might not be a simple process of cleavage and attachment. The complex albumin protein structure may play an important role in the process.

4. Conclusions

Although HSA adduct formation is a potential biomarker of exposure to OPEs, in the present study we did not detect any HSA adduct formation with any of the 11 target OPE FRs and OPE plasticizers. The results suggest that these OPEs have very different protein binding affinities in comparison to OP insecticides, or they may possess a specific form of currently unknown toxicity ([Abou-Donia, 2003; Casida and Quistad, 2004; Schang et al., 2016](#)). For example, *ortho*-TMPP (also known as tri-*ortho*-cresyl phosphate (ToCP)) is neurotoxic, and exposure to ToCP has been associated with the alleged Aerotoxic syndrome ([Abou-Donia et al., 2013; de Ree et al., 2014](#)). Most environmentally relevant OPE FRs and OPE plasticizers have shown low persistence and bioaccumulation potential in biota as they are rapidly metabolized ([Greaves et al. 2016c](#)). Therefore, their metabolisms and potent molecular toxicity may be very different from common persistent organic pollutants. A large amount and variety of OPE FRs and OP plasticizers have been produced and used, with more OPEs likely to be produced in future. Further studies are necessary to elucidate the behavior of OPEs, including their toxicology and effects in exposed organisms, to facilitate assessments of risk with respect to human and environmental health.

Conflict of interest statement

The authors declare that there are no conflicts of interest regarding the submitted article.

Acknowledgments

This work was supported in part by grant G12 MD007601 from the National Institutes of Health Research Centers in Minority Institutions Program, NIEHS R01ES002718, and a subcontract with the Western Center for Agricultural Health and Safety (NIOSH grant 2U54OH007550).

References

Abou-Donia, M.B., 2003. Organophosphorus ester-induced chronic neurotoxicity. *Arch. Environ. Health* 58 (8), 484–497.

Abou-Donia, M.B., Abou-Donia, M.M., ElMasry, E.M., Monro, J.A., Mulder, M.F.A., 2013. Autoantibodies to nervous system specific proteins are elevated in sera of flight crew members: biomarkers for nervous system injury. *J. Toxicol. Environ. Health A* 76 (6), 363–380.

Brommer, S., Harrad, S., 2015. Sources and human exposure implications of concentrations of organophosphate flame retardants in dust from UK cars, classrooms, living rooms, and offices. *Environ. Int.* 83, 202–207.

Camarasa, J.G., Serrabaldrich, E., 1992. Allergic contact-dermatitis from triphenyl phosphate. *Contact Dermat.* 26 (4), 264–265.

Casida, J.E., Quistad, G.B., 2004. Organophosphate toxicology: Safety aspects of nonacetylcholinesterase secondary targets. *Chem. Res. Toxicol.* 17 (8), 983–998.

Cristadoro, A., Rader, H.J., Mullen, K., 2008. Quantitative analyses of fullerene and polycyclic aromatic hydrocarbon mixtures via solvent-free matrix-assisted laser desorption/ionization mass spectrometry. *Rapid. Commun. Mass Spectrom.* 22 (16), 2463–2470.

Crow, B.S., Pantazides, B.G., Quinones-Gonzalez, J., Garton, J.W., Carter, M.D., Perez, J.W., Watson, C.M., Tomcik, D.J., Crenshaw, M.D., Brewer, B.N., Riches, J.R., Stubbs, S.J., Read, R.W., Evans, R.A., Thomas, J.D., Blake, T.A., Johnson, R.C., 2014. Simultaneous measurement of tabun, sarin, soman, cyclosarin, VR, VX, and VM adducts to tyrosine in blood products by isotope dilution UHPLC-MS/MS. *Anal. Chem.* 86 (20), 10397–10405.

de Ree, H., van den Berg, M., Brand, T., Mulder, G.J., Simons, R., van Zanten, B.V., Westerink, R.H.S., 2014. Health risk assessment of exposure to tricresyl phosphates (TCPs) in aircraft: a commentary. *Neurotoxicology* 45, 209–215.

EPA, 2015. EPA Document# 740-R1-5001. https://www.epa.gov/sites/production/files/2015-09/documents/cpe_fr_cluster_problem_formulation.pdf.

Greaves, A.K., Letcher, R.J., Chen, D., Gauthier, L.T., McGoldrick, D., Backus, S., 2016a. Retrospective trends of organophosphate flame retardants in herring gull eggs and in relation to the aquatic food web for the Laurentian Great Lakes of North America. *Environ. Res.* 150, 255–263.

Greaves, A.K., Letcher, R.J., 2016b. Organophosphate esters: a review on wildlife distribution, fate and toxicology. *Bull. Environ. Contam. Toxicol.* <http://dx.doi.org/10.1007/s00128-016-1898-0> (in press).

Greaves, A.K., Su, G., Letcher, R.J., 2016c. Environmentally relevant organophosphate triester flame retardants in herring gulls: *In vitro* biotransformation and kinetics and diester metabolite formation using a hepatic microsomal assay. *Toxicol. Appl. Pharmacol.* 308, 59–65.

Hallanger, I.G., Sagerup, K., Evensen, A., Kovacs, K.M., Leonards, P., Fuglei, E., Routti, H., Aars, J., Strom, H., Lydersen, C., Gabrielsen, G.W., 2015. Organophosphorus flame retardants in biota from Svalbard, Norway. *Mar. Poll. Bull.* 101 (1), 442–447.

Hartmann, P.C., Burgi, D., Giger, W., 2004. Organophosphate flame retardants and plasticizers in indoor air. *Chemosphere* 57 (8), 781–787.

Harvey, D.J., 2015. Analysis of carbohydrates and glycoconjugates by matrix-assisted laser desorption/ionization mass spectrometry: an update for 2009–2010. *Mass Spectrom. Rev.* 34 (3), 268–422.

He, R.W., Li, Y.Z., Xiang, P., Li, C., Zhou, C.Y., Zhang, S.J., Cui, X.Y., Ma, L.Q., 2016. Organophosphorus flame retardants and phthalate esters in indoor dust from different microenvironments: bioaccessibility and risk assessment. *Chemosphere* 150, 528–535.

John, H., Worek, F., Thiermann, H., 2008. LC-MS-based procedures for monitoring of toxic organophosphorus compounds and verification of pesticide and nerve agent poisoning. *Anal. Bioanal. Chem.* 391 (1), 97–116.

John, H., Breyer, F., Thumfart, J.O., Hochstetter, H., Thiermann, H., 2010. Matrix-assisted laser desorption/ionization time-of-flight mass spectrometry (MALDI-TOF MS) for detection and identification of albumin phosphorylation by organophosphorus pesticides and G- and V-type nerve agents. *Anal. Bioanal. Chem.* 398 (6), 2677–2691.

Li, B., Schopfer, L.M., Hinrichs, S.H., Masson, P., Lockridge, O., 2007. Matrix-assisted laser desorption/ionization time-of-flight mass spectrometry assay for organophosphorus toxicants bound to human albumin at Tyr411. *Anal. Biochem.* 361 (2), 263–272.

Mabey, W., Mill, T., 1978. Critical-review of hydrolysis of organic compounds in water under environmental conditions. *J. Phys. Chem. Ref. Data* 7 (2), 383–415.

Marklund, A., Andersson, B., Haglund, P., 2005. Organophosphorus flame retardants and plasticizers in Swedish sewage treatment plants. *Environ. Sci. Technol.* 39 (19), 7423–7429.

Marklund, A., Andersson, B., Haglund, P., 2003. Screening of organophosphorus compounds and their distribution in various indoor environments. *Chemosphere* 53 (9), 1137–1146.

Marsillach, J., Costa, L.G., Furlong, C.E., 2013. Protein adducts as biomarkers of exposure to organophosphorus compounds. *Toxicology* 307, 46–54.

McDonough, J.H., Shih, T.M., 1997. Neuropharmacological mechanisms of nerve agent-induced seizure and neuropathology. *Neurosci. Biobehav. Rev.* 21 (5), 559–657.

McGoldrick, D.J., Letcher, R.J., Barresi, E., Keir, M.J., Small, J., Clark, M.G., Sverko, E., Backus, S.M., 2014. Organophosphate flame retardants and organosiloxanes in predatory freshwater fish from locations across Canada. *Environ. Pollut.* 193, 254–261.

Meeker, J.D., Stapleton, H.M., 2010. House dust concentrations of organophosphate flame retardants in relation to hormone levels and semen quality parameters. *Environ. Health Persp.* 118 (3), 318–323.

Mercier, F., Gorenne, P., Thomas, O., Le Bot, B., 2011. Organic contamination of settled house dust, A review for exposure assessment purposes. *Environ. Sci. Technol.* 45 (16), 6716–6727.

Morgan, A.B., Gilman, J.W., 2013. An Overview of flame retardancy of polymeric

materials: application, technology, and future directions. *Fire Mater* 37 (4), 259–279.

Nomura, F., 2015. Proteome-based bacterial identification using matrix-assisted laser desorption ionization–time of flight mass spectrometry (MALDI-TOF MS): a revolutionary shift in clinical diagnostic microbiology. *Biochim. Biophys. Acta* 1854 (6), 528–537.

Peeples, E.S., Schopfer, L.M., Duyzen, E.G., Spaulding, R., Voelker, T., Thompson, C.M., Lockridge, O., 2005. Albumin, a new biomarker of organophosphorus toxicant exposure, identified by mass spectrometry. *Toxicol. Sci.* 83 (2), 303–312.

Reemtsma, T., Quintana, J.B., Rodil, R., Garcia-Lopez, M., Rodriguez, I., 2008. Organophosphorus flame retardants and plasticizers in water and air I. Occurrence and fate. *Trends Anal. Chem.* 27 (9), 727–737.

Schang, G., Robaire, B., Hales, B.F., 2016. Organophosphate flame retardants act as endocrine-disrupting chemicals in MA-10 mouse tumor Leydig cells. *Toxicol. Sci.* 150 (2), 499–509.

Schopfer, L.M., Furlong, C.E., Lockridge, O., 2010. Development of diagnostics in the search for an explanation of aerotoxic syndrome. *Anal. Biochem.* 404 (1), 64–74.

Schreder, E.D., Uding, N., La Guardia, M.J., 2016. Inhalation a significant exposure route for chlorinated organophosphate flame retardants. *Chemosphere* 150, 499–504.

Su, G., Letcher, R.J., Yu, H., 2016. Organophosphate flame retardant and plasticizer chemicals in aqueous solutions: hydrolytic pH stability, kinetics and mechanisms. *Environ. Sci. Technol.* 50 (15), 8103–8111.

Thompson, C.M., Prins, J.M., George, K.M., 2010. Mass spectrometric analyses of organophosphate insecticide oxon protein adducts. *Environ. Health Perspect.* 118 (1), 11–19.

van der Veen, I., de Boer, J., 2012. Phosphorus flame retardants: properties, production, environmental occurrence, toxicity and analysis. *Chemosphere* 88 (10), 1119–1153.

Wa, C.L., Cerny, R., Hage, D.S., 2006. Obtaining high sequence coverage in matrix-assisted laser desorption time-of-flight mass spectrometry for studies of protein modification: analysis of human serum albumin as a model. *Anal. Biochem.* 349 (2), 229–241.

Wang, X., Liu, J., Yin, Y., 2010. The pollution status and research progress on organophosphate ester flame retardants. *Prog. Chem.* 22 (10), 1983–1992.

Wei, G.L., Li, D.Q., Zhuo, M.N., Liao, Y.S., Xie, Z.Y., Guo, T.L., Li, J.J., Zhang, S.Y., Liang, Z.Q., 2015. Organophosphorus flame retardants and plasticizers: sources, occurrence, toxicity and human exposure. *Environ. Poll.* 196, 29–46.

Williams, N.H., Harrison, J.M., Read, R.W., Black, R.M., 2007. Phosphorylated tyrosine in albumin as a biomarker of exposure to organophosphorus nerve agents. *Arch. Toxicol.* 81, 627–639.

Zamy, C., Mazellier, P., Legube, B., 2004. Analytical and kinetic study of the aqueous hydrolysis of four organophosphorus and two carbamate pesticides. *J. Environ. Anal. Chem.* 84, 1059–1068.