

SIMULATION OF AIR QUALITY AND OPERATING COST TO VENTILATE SWINE FARROWING FACILITIES IN THE MIDWEST U.S. DURING WINTER

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ABSTRACT. *We have developed a time-dependent simulation model to estimate in-room concentrations of multiple contaminants, i.e., ammonia (NH₃), carbon dioxide (CO₂), carbon monoxide (CO), and dust, as a function of increased ventilation with filtered recirculation for swine farrowing facilities. Energy and mass balance equations were used to simulate the indoor air quality (IAQ) and operating cost for a variety of ventilation conditions over a three-month winter period for a facility located in the Midwest U.S., using simplified and real-time production parameters, and results were compared to field data. The model was improved by minimizing the sum of squared errors (SSE) between modeled and measured NH₃ and CO₂. After optimizing NH₃ and CO₂, other IAQ results from the simulation were compared to field measurements using linear regression. For NH₃, the coefficient of determination (R²) for simulation results and field measurements improved from 0.02 with the original model to 0.37 with the new model. For CO₂, the R² for simulation results and field measurements was 0.49 with the new model. When the makeup air was matched to hallway air CO₂ concentrations (1,500 ppm), simulation results showed the smallest SSE. With the new model, the R² for other contaminants were 0.34 for inhalable dust, 0.36 for respirable dust, and 0.26 for CO. Operation of the air cleaner decreased inhalable dust by 35% and respirable dust concentrations by 33%, while having no effect on NH₃ and CO₂, in agreement with field data, and increasing operating cost by \$860 (58%) for the three-month period.*

Keywords. *Ammonia, Carbon dioxide, Inhalable dust, Livestock, Simulink, Respirable dust.*

Modern swine barns are generally enclosed structures for producing a high density of swine. Feed, swine, and swine manure contribute to elevated concentrations of hazardous airborne dust and gases in these structures. Swine barn dust suspended in the air is composed of feed, feces, mold, pollen grains, insect parts, and mineral ash (Donham et al., 1986). Various gases, including ammonia (NH₃), methane (CH₄), and hydrogen sulfide (H₂S), are released by the digestion of swine manure stored in the pit below the floor. Carbon dioxide (CO₂) is generated by the respiration of swine (Donham, 1988; Chang et al., 2001) and can be generated by in-room heaters. Inhalation of these dusts and gases has been associated with adverse health outcomes in swine workers (Donham et al., 1986, 1989; Larsson et al., 1994; Donham et al., 1995; Iversen et al., 2000; Kirkhorn

and Garry, 2000; Donham et al., 2002; Charavaryamath et al., 2005; Hong et al., 2012) and may also depress the health status of swine (Stombaugh et al., 1969; Drummond et al., 1980; Donham, 1991; Diekman et al., 1993; Pedersen et al., 2000). The recommended maximum concentration of total dust is 2.4 mg m⁻³ for workers (Donham et al., 1989) and 3.7 mg m⁻³ for swine (Donham, 1991). For both workers and swine, the recommended maximum concentrations of respirable dust particles and CO₂ are 0.23 mg m⁻³ and 1,540 ppm, respectively (Donham et al., 1989). The recommended maximum concentration of NH₃ is 7 ppm for workers (Donham et al., 1989) and 11 ppm for swine (Donham, 1991).

Mechanical ventilation through pit fans and radial fans at one end of the room is commonly used to maintain temperatures slightly above ambient temperature during summer and maintain low gas and dust levels by removing these impurities from the room during winter. Room or pit air is mechanically exhausted, which pulls clean outside air into the barn. However, in upper Midwest facilities during winter, room exhaust fans are typically unused because the replacement air must be heated, resulting in increased heating cost (Peters et al., 2012). Lower ventilation rates lead to higher dust concentrations in winter compared to summer (O'Shaughnessy et al., 2009; Takai et al., 1998). Most producers run pit fans in the winter to exhaust air above the under-floor manure pits. Reeve et al. (2013) found that the use of pit fans in winter reduced dust, NH₃, and H₂S concentrations in a farrowing facility, but not necessarily below the concentrations recommended to protect worker health.

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Numerous researchers have used computer simulations or modeling to study the effects of mechanical ventilation in livestock facilities on indoor air quality (IAQ) and energy consumption (Pedersen et al., 1998; Soldatos et al., 2005; Cortus et al., 2010b). Anthony et al. (2014) evaluated the effectiveness of treating and recirculating air using multiple flow rates, percent dilution with outside air, and contaminant control devices for a representative swine farrowing room using a simulation model developed by Park et al. (2013). Anthony et al. (2014) reported that indoor dust concentrations were reduced (41% for respirable and 33% for inhalable) with the system in operation, while gas concentrations (NH_3 , CO , and CO_2) were unchanged. Their study provides evidence that incorporating standard ventilation controls can serve to reduce dust concentrations without increasing concentrations of gaseous contaminants. Park et al. (2013) developed and evaluated a mass and energy balance model to examine the relationship between IAQ, winter ventilation, air pollution control equipment, and heating needs. Outdoor temperatures were simulated based on seasonal estimates, and production levels (sow and piglet counts) and manure pit volume were held constant. However, NH_3 concentrations estimated by the model were substantially lower than those observed in swine farrowing rooms. Simplifications used in the original model included the use of a fixed NH_3 generation rate based on emission data measured in an Iowa farrowing barn (Cortus et al., 2010a), which did not consider swine number and manure volume in the pit. Cortus et al. (2009) developed a more complex equation to estimate NH_3 generation rate using temperature, total ammonia nitrogen (TAN) concentration, and pH to calculate the NH_3 generation rate from the slurry surface. However, real-time measurement of these variables is difficult. The simulation model of Park et al. (2013) may be improved by empirically incorporating room NH_3 measurements into the model.

Room concentrations of CO_2 estimated by the simulation model of Park et al. (2013) were also lower than those observed in a swine farrowing room (Reeve et al., 2013). Park et al. (2013) used an input value of 400 ppm for the CO_2 concentration of the makeup (outdoor) air entering the room, which was typical of ambient concentrations. However, the CO_2 concentration in rooms adjoining the simulated farrowing room, particularly the adjoining hallway that separated the study room and two other production areas (farrowing and nursery), may have contributed substantially to the makeup air in the test room. Therefore, an examination of appropriate makeup CO_2 concentration is required to improve the input value in the simulation model.

Room concentration measurements from field testing of a farrowing barn while deploying the optimized ventilation system are available to allow validation of the simulation model (Anthony et al., 2014). Outdoor temperatures, pit wall temperatures, and barn occupancy data were collected with room contaminant concentration data, providing data to validate the simulation model estimates. Understanding whether the time-dependent production factors or simplified seasonal temperature and production capacity numbers are sufficient to generate realistic concentration estimates in the barn over a winter can be examined using this robust data set.

The overall objective of this study was to improve the mass balance model to simulate the IAQ. The specific objectives were to: (1) enhance the existing model using an empirical model to simulate NH_3 emissions from slurry, (2) determine the extent to which realistic data on outdoor temperature and animal population within the barn are necessary to accurately estimate room concentrations (NH_3 , CO_2 , CO , and dust), and (3) validate the improved model using the field measurements.

MATERIALS AND METHODS

SIMULATED SWINE FARROWING FACILITY

A generalizable model was developed, but parameters were assigned to represent the building and operation of a specific swine farrowing facility (Mansfield Swine Education Center at Kirkwood Community College, Cedar Rapids, Iowa). Previous studies (Reeve et al., 2013; Park et al., 2013) fully described this facility (e.g., dimensions and airflow rates) with key parameters provided briefly here. Four radial wall fans positioned on the north and south room walls were not in operation during the winter. Two pit fans removed air above the manure pits, exhausting air out the west side of the building (Q_{tp} , $0.82 \text{ m}^3 \text{ s}^{-1}$) (fig. 1a). During the field testing described by Anthony et al. (2015), a single, gas-fired heater maintained the room temperature by cycling on when the room temperature dropped below 20.0°C and off when the temperature exceeded 22.2°C . Electric heat lamps were positioned in the crates when piglets were present. A combination of metal and plastic slats separated the swine crates from the manure pits below. The two pit fans exhausted air to outside of the building with an air exchange (Q_{ae}) through the slatted floor and over the manure pit located under the four rows of crates (19 total crates). An air cleaner (model 140 shaker dust collector, United Air Specialists, Inc., Blue Ash, Ohio) was installed outside the building. The air cleaner exhausted barn air at a flowrate of $0.47 \text{ m}^3 \text{ s}^{-1}$ ($1000 \text{ ft}^3 \text{ min}^{-1}$) with an internal fan, filtered the dust, and returned 100% of the treated air to the building. The air cleaner was operated with a standard fabric filter (14-pocket polyester sateen filter, United Air Specialists, Inc., Blue Ash, Ohio), as used by Peters et al. (2015).

The simulated room volume was divided into two sections (a habitable portion and the manure pit), as shown in figure 1b, with key parameters described in table 1. The habitable section was modeled as a rectangular box, dimensioned to match the test facility, as shown in figure 1a. The manure pit headspace was modeled as four equal-size rectangular boxes. In the original model, the total pit headspace volume was fixed at 67.5 m^3 (Park et al., 2013; Anthony et al., 2014). In the new model, the total pit headspace volume decreased over time as the pit was filled with swine manure over the three-month period. The pit headspace volume (V_p) was calculated as follows:

$$V_p(t) = V_{p,max} - \dot{G}_{slurry} \times m_{swine} \times t \quad (1)$$

$$(V_{p,min} \leq V_p \leq V_{p,max})$$

where $V_{p,max}$ (67.5 m^3) and $V_{p,min}$ (20.8 m^3) are the maximum

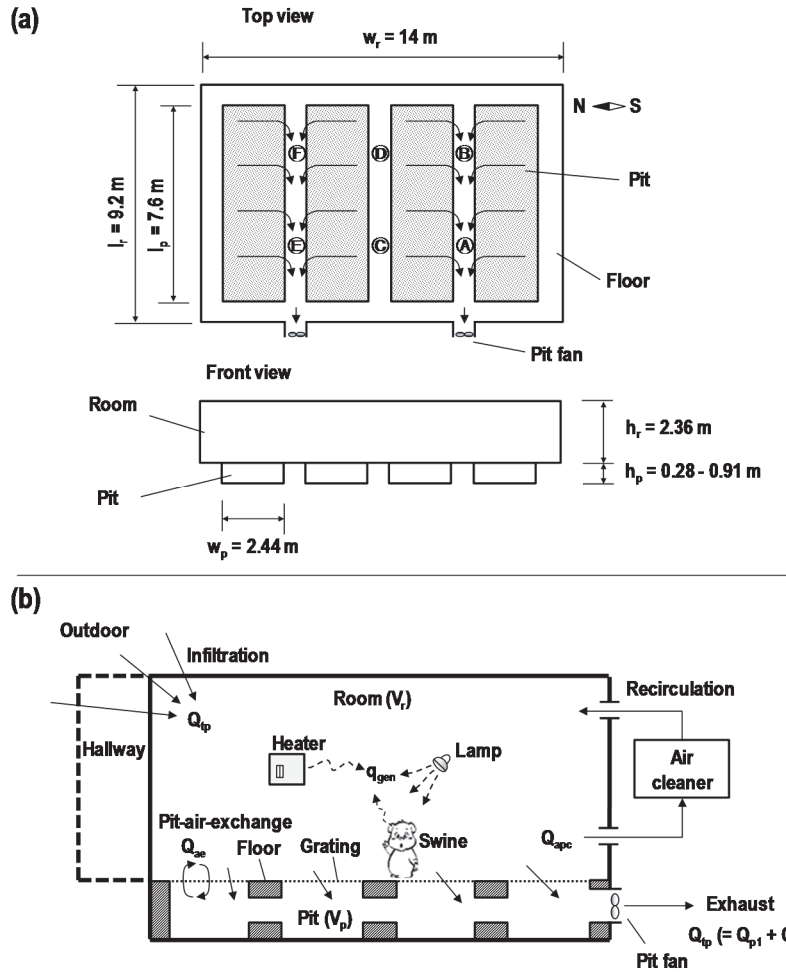


Figure 1. Schematic of modeled swine farrowing facility, identifying (a) dimensions and six sampling positions (A to F) for field measurements and (b) airflow pathways.

Table 1. Physical and operating parameters of the test site used as model input.

Parameter	Value	Note
Room volume (V_r)	304 m ³	= 9.2 m × 14 m × 2.36 m
Pit headspace volume (V_p)	67.5 m ³ max.; 20.8 m ³ min.	= 4 × 2.44 m × 7.6 m × h_p ; 0.28 m ≤ h_p ≤ 0.91 m
Total airflow rate of pit fans (Q_{fp})	0.82 m ³ s ⁻¹	= 2 × 872 ft ³ min ⁻¹
Airflow rate of pit air exchange (Q_{ae})	0.08 m ³ s ⁻¹	Assumed 10% of total airflow rate of pit fans (Cortus et al., 2010b)
Total flow rate of air cleaner (Q_{ac})	0.47 m ³ s ⁻¹	= 1000 ft ³ min ⁻¹ , per manufacturer
Slurry generation rate (\dot{G}_{slurry})	$8.38 \times 10^{-10} \text{ m}^3 \text{ kg}^{-1} \text{ s}^{-1}$	= 1.16 ft ³ AU ⁻¹ day ⁻¹ (Chastain et al., 1999)
Piglet number (n_{piglet})	0 to 120	Field measurement (table A2 in the Appendix)
Sow number (n_{sow})	0 to 19	Field measurement (table A2 in the Appendix)
Mass of one piglet (m_{piglet})	4.5 kg	Assumption (10 lb)
Mass of one sow (m_{sow})	196.4 kg	Assumption (433 lb)
Total mass of swine (m_{swine})	-	= $n_{piglet} \times m_{piglet} + n_{sow} \times m_{sow}$

and minimum volumes of the pit headspace, respectively. Initial V_p was set to $V_{p,max}$, representing an empty pit. \dot{G}_{slurry} is the slurry generation rate per unit swine mass, and m_{swine} is the total mass of sows and piglets at a given time. In the original model, the swine number was fixed and was not coupled with the pit headspace volume (Park et al., 2013; Anthony et al., 2014). The new model used real swine occupancy of the field test site (provided in table A2 in the Appendix).

In the original model, outdoor temperature (T_o) was simulated using historical seasonal average data for Cedar Rap-

ids, Iowa, modeled as a combination of two sine waves (Park et al., 2013; Anthony et al., 2014) to account for within-day and between-day temperatures. In contrast, in the current model, T_o was set to the actual temperature of Cedar Rapids, Iowa, from December 2013 to February 2014, the period of the field study for model validation (CID Airport meteorological data from NOAA's National Climatic Data Center), where at least hourly temperatures were available. Model equations for temperature and operating cost are documented in the Appendix, with electrical (\$0.0807 kWh⁻¹) and natural gas (\$0.27 m⁻³) based on industrial pricing in Iowa (December 2013 to February 2014).

The simulation started at 8:00 a.m. on December 1 and proceeded through the end of February (90 days). This project generated a time-dependent simulation model using MatLab R2014a (ver. 8.3.0.532, MathWorks, Natick, Mass.) with Simulink (ver. 8.3, MathWorks, Natick, Mass.).

MODEL EQUATIONS AND PARAMETERS FOR IAQ

Gas and dust concentrations in the room were simulated simultaneously with energy balance equations (eqs. A1 and A2 in the Appendix). Under the assumption of a well-mixed indoor space, the mass-balance equations for NH₃, CO₂, CO, and dust for the room and pit volumes were as follows (Park et al., 2013):

Room:

$$V_r \frac{dP_r}{dt} = Q_{ac}(1 - \eta_p)P_r + Q_{tp}P_o + Q_{ae}P_p + \dot{G}_{Pr} - Q_{ac}P_r - Q_{tp}P_r - Q_{ae}P_r \quad (2)$$

Pit:

$$V_p \frac{dP_p}{dt} = Q_{tp}P_r + Q_{ae}P_r + \dot{G}_{Pp} - Q_{tp}P_p - Q_{ae}P_p \quad (3)$$

where P_o is the outdoor concentration, P_r is the room concentration, and P_p is the pit concentration. Values for the input parameters for the contaminant generation rate for the room (\dot{G}_{Pr}) and for the manure pit (\dot{G}_{Pp}) and the removal efficiency of the air cleaner (η_p) are given in table 2. In the original model, the NH₃ generation rate was independent of swine number and manure volume in the pit: NH₃ had been fixed at 1.11 mg s⁻¹ (Park et al., 2013). In the current model, the NH₃ generation rate was replaced with a dynamic value. To calculate the CO₂ concentration, simulation values for the makeup air CO₂ concentration were optimized and constrained between 400 ppm (ambient air quality outside) and 1,750 ppm (concentrations in the adjoining hallway).

Generation Rate of NH₃

Generating NH₃ is a process of mass transfer from the NH₃ solution to the free atmosphere. The generation rate of NH₃ can be calculated as follows (Aarnink and Elzing, 1998):

$$\dot{G}_{NH_3} = \frac{k \times A_p \times f \times TAN}{H} \quad (4)$$

where k is the mass transfer coefficient, A_p is the surface area of the manure, f is the un-ionized fraction of the total ammoniacal nitrogen (TAN), and H is the Henry constant. The

mass transfer coefficient can be described by the following equation (Aarnink and Elzing, 1998; Bjerg et al., 2013):

$$k = Z \times v_s^{0.8} \times T_{film}^{-1.4} \quad (5)$$

where Z is the empirically determined constant, v_s is the air speed in the vicinity of the slurry, and T_{film} is the surface temperature of the slurry. Measuring or estimating values for f , TAN, and H are difficult because they change over time. Therefore, a new empirical model was proposed to estimate the generation rate of NH₃ in the pit (\dot{G}_{NH_3p}) based on m_{swine} and T_{film} . The value of \dot{G}_{NH_3p} was calculated as follows:

$$\dot{G}_{NH_3p} = Z_{NH_3} \times \frac{m_{swine}}{500 \text{ kg}} \times \dot{G}_{NH_3,EM} \times T_{film}^{-1.4} \quad (6)$$

where Z_{NH_3} (K^{1.4} m^{0.2} s^{-0.2}) is the coefficient determined by fitting modeled concentration estimates to field measurements. Z_{NH_3} includes ventilation air speed (v_s) and slurry conditions (f , TAN, and pH). $\dot{G}_{NH_3,EM}$ is the NH₃ emission rate per 500 kg swine, and a standard 1.5 mg s⁻¹ per 500 kg swine was used (Heber et al., 2000).

Generation Rate of CO₂ and Makeup Air CO₂ Concentration

Sources of CO₂ inside the room included the gas-fired heater exhaust, swine exhalation, and digestion of slurry in the pit. The CO₂ generation rate of the heater was assumed to be zero when off and 906 mg s⁻¹ when on, which was calculated as follows (Park et al., 2013):

$$\dot{G}_{CO_2r,heater} = \dot{Q}_{heater} \times ER_{CO_2} \quad (7)$$

where \dot{Q}_{heater} is the natural gas consumption of the gas heater in use at the test site (0.000472 m³ s⁻¹), and ER_{CO_2} is the CO₂ emission rate from natural gas combustion (1,920,000 mg m⁻³) (EPA, 1998). The generation rate of CO₂ by swine respiration was equal to 0.000201 m³ h⁻¹ W⁻¹ (Blanes and Pedersen, 2005). The total generation rate of CO₂ by swine respiration was calculated as:

$$\dot{G}_{CO_2r,swine} = \frac{0.000201 \times \rho_{CO_2} \times 1000000}{3600} \times \dot{q}_{swine} \quad (8)$$

where ρ_{CO_2} is the density of CO₂ (1.84 kg m⁻³), and \dot{q}_{swine} is the total heat generation rate from swine (table A1). For converting units, 1,000,000 mg kg⁻¹ and 3,600 s h⁻¹ were used. The generation rate of CO₂ from slurry in the pit was

Table 2. Input parameters for IAQ calculation.

Parameter	Initial Room Concentration (P_{ini})	Outdoor Concentration (P_o) ^[a]	Generation Rate in the Room (\dot{G}_{Pr})	Generation Rate in the Pit (\dot{G}_{Pp})	Removal Efficiency of Air Cleaner (η_p)
NH ₃	6.27 ppm	0 ppm	0 mg s ⁻¹	Equation 6	0
CO ₂	1,750 ppm	400, 750, 1,250, 1,500, and 1,750 ppm	Heater: equation 7 Swine: equation 8	37.5% of CO ₂ from swine	0
CO	0 ppm	0 ppm	Heater: equation 9	0 mg s ⁻¹	0
Inhalable dust	2.19 mg m ⁻³	0 mg m ⁻³	Equation 10 and figure A1	0 mg s ⁻¹	0.95
Respirable dust	0.23 mg m ⁻³	0 mg m ⁻³	Equation 10 and figure A1	0 mg s ⁻¹	0.85

^[a] Makeup air for CO₂

assumed as 37.5% of swine exhalation (Ni et al., 1999). The north, south, and west walls of the simulated room face an outdoor area. However, there is a hallway outside the east wall. The hallway CO₂ concentration was observed to be larger than 1,500 ppm. Makeup air pulled by the pit fans may include air from the adjoining hallway. Therefore, the model was run with different CO₂ concentrations assumed in the makeup air: 400 ppm as used in the original model (Park et al., 2013), 750 ppm to reflect in-room concentrations published by Chang et al. (2001), and 1,250, 1,500, and 1,750 ppm to reflect field concentrations measured in the adjacent hallway that served as a source of some of the makeup air exhausted by fans at the field test site. The impact of these different CO₂ input concentrations was assessed.

Generation Rates of CO and Dust

Generation rates of CO and dust used in the current model were the same as in the original model (Park et al., 2013). However, inputs for the swine number and outdoor temperature were changed to match field conditions over the 2013-2014 study period. CO was generated by a gas-fired heater in the room because combustion gases were not vented out of the room. Simulation of the heater operation (on vs. off) was triggered by the need to maintain production temperatures in the room (20.0°C to 22.2°C) as a function of heat loss from outside temperature changes. When the heater was on, the CO generation rate was 0.6 mg s⁻¹, which was calculated using:

$$\dot{G}_{CO} = \dot{Q}_{heater} \times ER_{CO} \quad (9)$$

The CO emission rate from natural gas combustion (ER_{CO}) used 640 mg m⁻³ (EPA, 1998).

The dust generation rate in the room (\dot{G}_{dust}) was modeled for both inhalable dust (1 to 100 µm in diameter) and respirable dust (1 to 10 µm in diameter, with 50% cut point of 4 µm), as defined by ACGIH (2016). Generation rates were calculated as follows:

$$\dot{G}_{dust} = \dot{G}_{mean} \times \frac{m_{swine}}{500 \text{ kg}} \quad (10)$$

where \dot{G}_{mean} is the overall mean dust generation rate per 500 kg of swine, and m_{swine} is the total mass of swine (kg). The \dot{G}_{mean} values for inhalable and respirable dust were 0.1575 and 0.0164 mg s⁻¹, respectively (Takai et al., 1998). To account for the fact that dust generation depends on feeding time, the dust generation rate was assumed to increase during feeding, as shown in figure A1 in the Appendix. The first and second feedings were modeled to occur at 9:00 a.m. for 30 min and at 4:00 p.m. for 30 min, respectively. Dust concentrations at the first and second feeding were modeled as four and two times higher than the mean daily concentration, respectively, based on previous field studies (O'Shaughnessy et al., 2009).

FIELD MEASUREMENTS

In our previous study (Anthony et al., 2015), experimental data were collected in the swine farrowing facility.

Field sampling was conducted on 18 days from 13 December 2013 to 27 February 2014. The air cleaner was off for seven of the sampling days (13 to 19 December, 22 to 27 January, and 26 to 27 February) and on for 11 of the sampling days (20 December to 21 January, and 28 January to 25 February). Twenty-four hour (from 8:00 a.m. to 8 a.m.) monitoring was conducted throughout the study period at six fixed positions, indicated as A through F (1.5 m height) in figure 1a. All devices were pre- and post-calibrated in the laboratory for each sampling event. IAQ concentration data included NH₃, CO, CO₂, and dust (inhalable and respirable). NH₃ and CO were measured with a multi-gas survey monitor (VRae, Rae Systems, San Jose, Cal.), and CO₂ was measured with a single-gas monitor (ToxiRae, Rae Systems). Inhalable dust was sampled with an IOM sampler (225-70A, SKC, Inc., Eighty-Four, Pa.) at 2 L min⁻¹ with 25 mm PVC filters (5 µm pore, 25 mm, SKC), and respirable dust was sampled using a cyclone (GK2.69, BGI, Mesa Labs, Inc., Butler, N.J.) with PVC filters (5 µm pore, 37 mm, SKC) at a sampling flow rate of 4.2 L min⁻¹. A micro-balance (MT5, Mettler-Toledo, Columbus, Ohio) was used to pre- and post-weigh the PVC filters to obtain room dust concentrations. Data averaged over all six positions and 24 h were used to validate the simulation model.

SIMULATION SCENARIOS AND DATA ANALYSIS

Park et al. (2013) previously reported the sensitivity between inputs and outputs using a simulation model for various conditions of outdoor temperature, wall insulation, pit air exchange ratio, pit fans, recirculation ratio, and filtration efficiency. The pit air exchange ratio (r_{ae}), T_o , wall insulation (U_{rw}), and recirculation ratio (r_r) affected heater operating cost. Outdoor temperature (T_o) and r_r were more sensitive to heater operation than U_{rw} and r_{ae} . The pit air exchange ratio (r_{ae}) and r_r affected room NH₃, CO, and CO₂ concentrations, although NH₃ was the most sensitive to r_{ae} because the only NH₃ source was in the pit. However, CO and CO₂ concentration were more sensitive to r_r than to r_{ae} because significant CO and CO₂ sources were in the room and not only the manure pit. Additional tests for sensitivity were not conducted because they were unchanged with the same mass balance equation used in both models. Constant values for r_{ae} (0.1) and r_r (1) were used, based on this previous analysis.

Simulations were performed for the ventilation with pit fans and recirculation with a dust filtration system. Simulation parameters are given in tables 1 and 2. The operating condition of air cleaner in the model was represented as "1" in the field test (table A3). Results from the simulations were compared to the field measurements. Linear regression was used to identify the slope and intercept between calculated and field measured data. Coefficients of determination (R^2) were calculated using Microsoft Excel 2010.

Simulations were performed to optimize NH₃ and CO₂ modeling first. To improve the NH₃ model with empirical generation rate, simulations were performed preferentially with five values (15,000, 20,000, 21,000, 21,500, and 25,000) for Z_{NH_3} . Additional simulations were performed with five replacement air CO₂ values (400, 750, 1,250, 1,500, and 1,750 ppm) to find an appropriate value for CO₂

concentration in the makeup air. The simulation with the minimum sum of squared error (SSE) was used to identify the optimum NH₃ generation rate factor and makeup air CO₂ concentration. The SSE was calculated using NH₃ and CO₂ concentrations from simulation (P_s) and measurement (P_m) as follows:

$$\text{SSE} = \sum (P_m - P_s)^2 \quad (11)$$

The values for Z_{NH_3} and makeup CO₂ concentration that showed small SSE were chosen for further simulations.

Once the model was optimized to minimize the SSE for NH₃ and CO₂, three time-dependent conditions were also assumed for the operation of the air cleaner. First, the air cleaner was modeled in the off condition for the entire winter, and then the system was simulated with the ventilation system on. Finally, the air cleaner was represented with an on-off operating cycle in field tests for the entire winter. This allowed an assessment of operating cost and air quality differences between using the air cleaner or not. For these simulations, real swine occupancy and real outdoor temperature data were used. All other parameters were fixed, as shown in tables 1 and A1.

RESULTS AND DISCUSSION

IMPROVING SIMULATION OF NH₃ AND CO₂

Optimization of NH₃ with Z_{NH_3}

Room NH₃ concentrations from the original model (Park et al., 2013), the new model, and field measurements are shown in figure 2a. In the original model (Park et al., 2013), the NH₃ concentration was 0.17 ppm and did not change because r_{ae} (0.1), pit fan operation, and r_r (1) of the air cleaner were fixed. The resulting R² was 0.02, indicating that estimates of NH₃ made with the original model poorly reflected reality. The linear regression between measured NH₃ concentration and the new model simulation had a modest R² (0.37), which is a substantial improvement over the original model (table 3). The difference in Z_{NH_3} did not change the R² value. The improvement is attributed to the difference in Z_{NH_3} affecting the NH₃ concentration only. When Z_{NH_3} was 15,000, the NH₃ concentration from the new model ranged from 0.03 to 10.59 ppm. When Z_{NH_3} increased to 25,000, the NH₃ concentration ranged from 0.06 to 17.65 ppm. When Z_{NH_3} was 21,000, SSE was 438, which was smaller than 602 at $Z_{\text{NH}_3} = 15,000$ or 524 at $Z_{\text{NH}_3} = 25,000$. The value of Z_{NH_3} was fixed at 21,000 for further simulations.

With Z_{NH_3} was fixed at 21,000, the 24 h averaged NH₃ concentration from the simulations ranged from 0.05 to 15 ppm, while field NH₃ concentrations ranged from 0.07 to 28 ppm (table A4). Anthony et al. (2015) reported that both the number of sows and T_o were significant factors for estimating 24 h NH₃ concentrations. When the swine number was zero at -13.4°C (December 31), the simulated and measured NH₃ concentrations were 0.05 and 0.07 ppm, respectively. However, when the sow number was 17 at -13.0°C (26 February), the simulated and measured NH₃ concentrations increased to 13.44 and 7.25 ppm, respectively. Outdoor temperature could affect T_g because the pit wall is an exposed structure above the ground. However, T_g in equation

6 was assumed to be a fixed value. Outdoor temperature might be considered in Z_{NH_3} instead of T_g . If measuring and simulating time-dependent T_g are available, simulation can be improved.

Optimization of CO₂ with Makeup Air CO₂ Concentration

Estimates of CO₂ concentrations from simulations for each formulation of makeup CO₂ concentration are compared to field measurement in figure 2b. Room CO₂ concentrations were underestimated when the makeup air CO₂ was assumed to be 400 and 750 ppm. When 1,500 ppm was used as the assumed makeup air CO₂, the trend line (dotted line) crossed 1:1 line (gray line). The CO₂ concentration in the hallway adjacent to the farrowing room in the field studies ranged from 1,010 to 3,330 ppm (mean of 2,140 ppm), indicating that if makeup air into the farrowing room came from the hallway, it would be closer to the 1,500 ppm that was simulated. When the makeup air CO₂ concentration was 1,500 ppm, SSE was 1.1×10^6 . When the outdoor concentration simulations used 1,750 or 450 ppm, SSE was increased 2.5 and 18 times, respectively. The makeup air CO₂ concentration was set to 1,500 ppm for further simulations.

Estimates of farrowing room CO₂ concentrations ranged from 2,103 to 2,755 ppm (using makeup air CO₂ of 1,500 ppm; table A4). Room CO₂ concentrations exceeded comfort levels established by ASHRAE Standard 62-1999 (1,000 ppm; ASHRAE, 1999) and were generally higher than industry recommendations (1,540 ppm) (Donham et al., 1989). Anthony et al. (2015) reported that piglet number and T_o are significant factors for estimation of room CO₂ concentration. Similar to their results, simulated CO₂ concentrations increased when T_o decreased or swine number increased. Room CO₂ concentration was highest during the coldest day because the heater, which produces CO₂, must operate more frequently and for a longer time to maintain an acceptable room air temperature. The swine number is the other important factor for CO₂ simulation. When the swine number was zero, as it was on 31 December in the field study, the room CO₂ concentration was the lowest even though T_o was colder than on other days. The presumption that makeup air entering into a farrowing room is independent of concentrations throughout the rest of the building is perhaps a poor assumption, as it was in the current work.

Room Concentrations of CO and Dust

Simulated room concentrations of CO and dust, along with comparative field measure results, are shown in figures 2c to 2e (data given in table A4). As shown in figure 2c, CO concentration was underestimated in the model, ranging from only 0.25 to 0.31 ppm, while the field-measured CO concentrations ranged from 0.94 to 3.29 ppm. One reason for the difference could be the detection limit of the field instrument. The nominal range of the CO sensor was 0 to 500 ppm with a resolution of 1 ppm. The measured CO concentrations were close to the lower detection limit of the sensor.

The linear regression between measured and simulated inhalable and respirable dust concentrations had modest R² values (0.34 and 0.40, respectively). In the simulated and field-measured results, both inhalable and respirable dusts

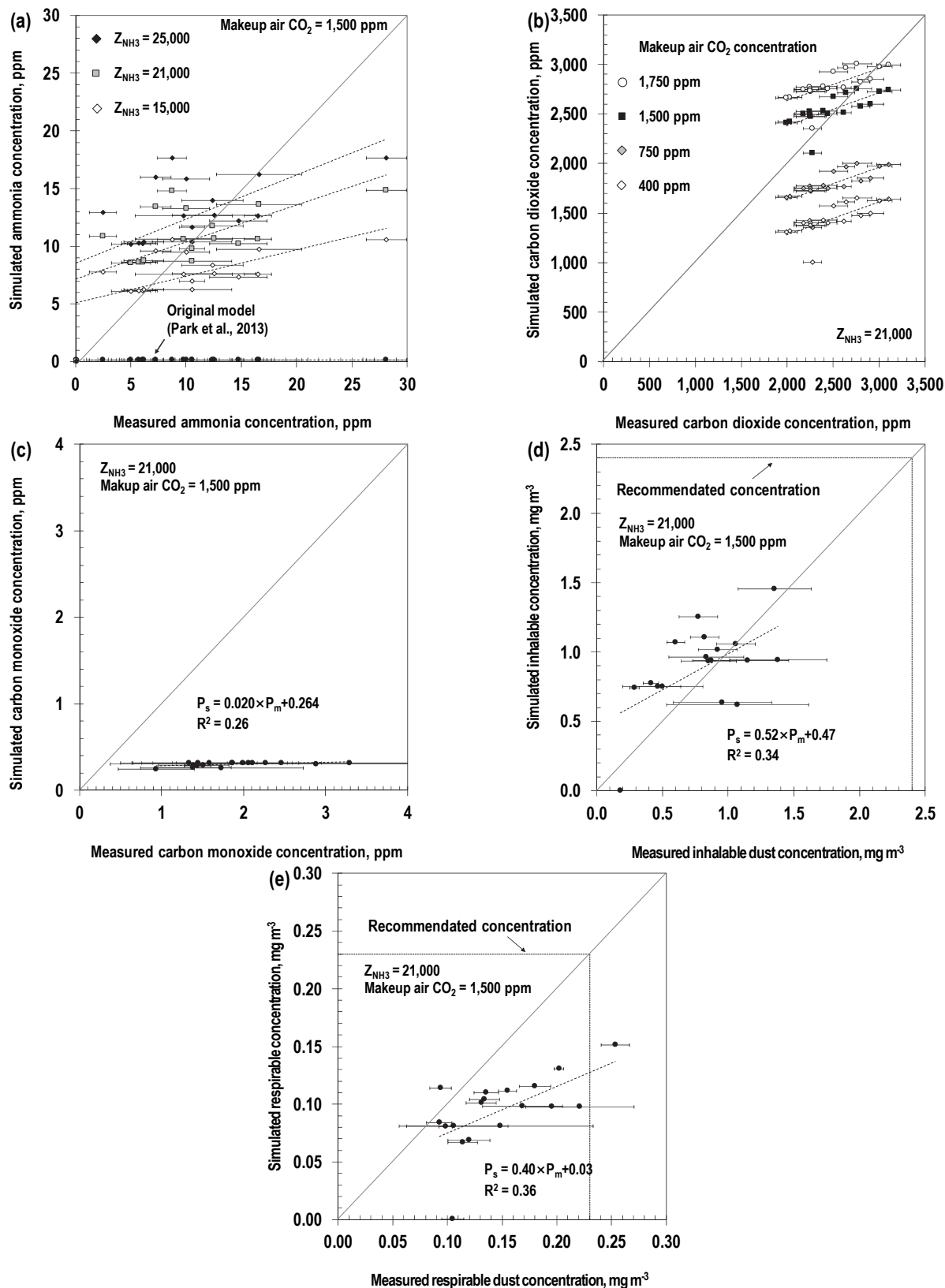


Figure 2. Linear regression between measured and calculated results for (a) ammonia, (b) carbon dioxide, (c) carbon monoxide, (d) inhalable dust, and (e) respirable dust (solid line = 1 on 1 line, dashed line = trend line, P_s = simulated concentration, and P_m = measured concentration).

decreased when the air cleaner was operated. Compared to days when the air cleaner was not in operation and when field measurements were available, inhalable dust concen-

trations decreased by 32% in simulated results and by 31% in field-measured results with the air cleaner on. Similarly, respirable dust concentration decreased by 31% in simulated

Table 3. Linear regression and SSE results for NH₃ and CO₂.

Parameter	Variable		Linear Regression ^[a]			SSE
			A	B	R ²	
NH ₃	Z _{NH3}	15,000	0.23	5.12	0.37	602
		20,000	0.31	6.82	0.37	441
		21,000	0.32	7.16	0.37	438
		21,500	0.33	7.33	0.37	440
		25,000	0.38	8.56	0.37	524
CO ₂	Makeup air	400	0.33	628	0.49	2.0 × 10 ⁷
	CO ₂ conc. (ppm)	750	0.33	978	0.49	9.6 × 10 ⁶
		1,250	0.33	1,478	0.49	1.7 × 10 ⁶
		1,500	0.33	1,728	0.49	1.1 × 10⁶
		1,750	0.33	1,978	0.49	2.8 × 10 ⁶

^[a] $P_s = A \times P_m + B$, where P_s = simulated concentration, and P_m = measured concentration.

results and by 41% in field-measured results. Swine number also affected dust concentration because dust generation rate was based on the total mass of swine. Maximum values of dust concentration from the model and field measurements occurred on 26 February, when there were large numbers of swine without the air cleaner on.

AIR CLEANER OPERATION

Simulations with the air cleaner off for the entire winter and with the air cleaner on during the entire winter were conducted, and the results are documented in table 4. The on-off cycle for the field testing data was required to examine the air cleaner performance as part of other research. However, a producer would likely run such an air cleaning system for an entire season (i.e., permanently on) or would not have an air cleaning system, which is equivalent to the off condition. The air cleaner operation only reduced dust concentrations, by design, because this air cleaner does not remove contaminant gases (NH₃ and CO₂). When the air cleaner was not operated during the winter season, the inhalable and respirable dust concentrations had three-month means of 0.97 and 0.10 mg m⁻³, respectively. When the air cleaner was running during the entire winter, simulated three-month mean inhalable and respirable dust concentrations decreased by 35% and 30%, respectively. The room dust concentration decreased linearly with increasing filtration efficiency (Park et al., 2013). Inhalable dust was removed more than respirable dust because the air cleaner collection efficiency was 95% for inhalable dust and 85% for respirable dust.

Three-month operating costs (electrical and heating) are also given in table 4. When the air cleaner was turned off during winter, the total operating cost was \$1,496. When the air cleaner was turned on, the total operating cost increased by 58% to \$2,359, primarily due to additional power consumption to operate the air cleaner's fan.

Table 4. Simulation results for air cleaner operation, three-month means (Z_{NH3} = 21,000; makeup air CO₂ concentration = 1,500 ppm).

Contaminant	Operating Time (days)		
	0 on, 90 off	59 on, 31 off ^[a]	90 on, 0 off
NH ₃ (ppm)	9.0	9.0	9.0
CO ₂ (ppm)	2,475	2,475	2,475
CO (ppm)	0.30	0.30	0.30
Inhalable dust (mg m ⁻³)	0.98	0.73	0.63
Respirable dust (mg m ⁻³)	0.10	0.08	0.07
Total operating cost	\$1,496	\$2,062	\$2,359
Operating cost for air cleaner	\$0	\$566	\$863

^[a] Matches field test condition.

LIMITATION AND RECOMMENDATION

One limitation of this study is that the simulation was conducted for only one configuration of a farrowing room. Simulation with Z_{NH3} of 21,000 and makeup air CO₂ concentration of 1,500 ppm showed the best results for the room configuration in this study. Z_{NH3} covers various factors such as ventilation and slurry properties; these values may be different for other swine barns. Differences also exist between this simulated room and high-production facility rooms, including room dimensions, crate layout, and pit volume (total and headspace above manure). In addition, other production facilities may house more swine per square foot than this study location and have larger piglet production targets (e.g., 11 piglets per sow), which would yield higher generation rates for multiple contaminants. These differences can affect Z_{NH3} and other parameters in the current model. For buildings with vented heaters or with lower CO₂ contamination in the makeup air, results may differ, but the simulation model was set up to account for these differences. Further validation is needed for different building designs and production densities to evaluate the robustness of the model validated here.

The model could be used to optimize ventilation systems for livestock facilities to provide good air quality at the lowest cost, even though the model has limitations. The simulation model provides a useful tool for examining the costs and benefits of installing common ventilation technology in confined animal feeding operations and, ultimately, making sound management decisions.

CONCLUSION

We improved our original mass and energy balance model to improve room NH₃ and CO₂ concentrations, matching the test conditions in a field study of a swine farrowing barn. An empirical model was used to improve NH₃ simulation, with a substantial but imperfect increase in agreement between the simulation and field measurements (R² improved from 0.02 to 0.37 with the new formulation). To improve CO₂ concentration estimates, higher concentrations in the makeup air were required, which was justified by assuming that makeup air entered the farrowing room from other rooms in the building, as verified by field data. After optimization for NH₃ and CO₂, air cleaner operation was evaluated. The air cleaner operation only changed the dust concentrations and did not remove NH₃ or CO₂. While the operating cost increased by \$863 over three months due to air cleaner operation, the concentrations of inhalable and respirable dusts decreased by 35% and 33%, respectively. The new model was able to simulate a variety of conditions, making it a potential tool for future simulations of IAQ and operating cost in swine farrowing rooms.

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NOMENCLATURE

A_p = surface area of manure (m^2)
 A_{pw} = overlap area of pit and wall (m^2)
 A_{rf} = overlap area of room and floor (m^2)
 A_{rw} = overlap area of room and wall (m^2)
 c_p = specific heat at constant pressure of air ($J\ kg^{-1}\ K^{-1}$)
 ER_{CO} = CO emission rate from natural gas combustion ($mg\ m^{-3}$)
 ER_{CO2} = CO₂ emission rate from natural gas combustion ($mg\ m^{-3}$)
 f = un-ionized fraction of TAN (dimensionless)
 \dot{G}_{CO_r} = generation rate of CO by heater ($mg\ s^{-1}$)
 $\dot{G}_{CO2r,heater}$ = generation rate of CO₂ by heater ($mg\ s^{-1}$)
 $\dot{G}_{CO2r,swine}$ = generation rate of CO₂ by swine ($mg\ s^{-1}$)
 \dot{G}_{dustr} = dust generation rate in the room ($mg\ s^{-1}$)
 \dot{G}_{mean} = overall mean dust generation rate per 500 kg

swine ($mg\ s^{-1}$)
 \dot{G}_{NH_3} = generation rate of NH₃ ($mol\ s^{-1}$)
 $\dot{G}_{NH_3,EM}$ = generation rate of NH₃ per 500 kg swine ($mg\ s^{-1}$)
 \dot{G}_{NH_3p} = generation rate of NH₃ in the pit ($mg\ s^{-1}$)
 \dot{G}_{Pp} = contaminant generation rate for pit ($mg\ s^{-1}$)
 \dot{G}_{Pr} = contaminant generation rate for room ($mg\ s^{-1}$)
 \dot{G}_{slurry} = slurry generation rate ($m^3\ kg^{-1}\ s^{-1}$)
 H = Henry's constant (dimensionless)
 k = mass transfer coefficient ($m\ s^{-1}$)
 m_{piglet} = mass of one piglet (kg)
 m_{sow} = mass of one sow (kg)
 m_{swine} = total mass of swine (kg)
 n_{piglet} = number of piglets (dimensionless)
 n_{sow} = number of sows (dimensionless)
 P_{elect} = electricity cost ($\$ kWh^{-1}$)
 P_{gas} = natural gas cost ($\$ m^{-3}$)
 P_m = measured concentration of NH₃ or CO₂ (ppm)
 P_o = outdoor concentration of contaminant (ppm or $mg\ m^{-3}$)
 P_p = pit concentration of contaminant (ppm or $mg\ m^{-3}$)
 P_r = room concentration of contaminant (ppm or $mg\ m^{-3}$)
 P_s = simulated concentration of NH₃ or CO₂ (ppm)
 Q_{ac} = total flow rate of air cleaner ($m^3\ s^{-1}$)
 Q_{ae} = airflow rate of pit air exchange ($m^3\ s^{-1}$)
 Q_{tp} = total airflow rate of two pit fans ($m^3\ s^{-1}$)
 \dot{Q}_{heater} = natural gas consumption of one gas heater ($m^3\ s^{-1}$)
 \dot{q}_{ac} = power consumption of air cleaner (W)
 \dot{q}_{gen} = total heat generation rate in the room (W)
 \dot{q}_{heater} = heat generation rate from one gas heater (W)
 \dot{q}_{lamp} = total heat generation rate from 20 lamps (W)
 \dot{q}_{piglet} = heat generation rate from one piglet (W)
 \dot{q}_{sow} = heat generation rate from one sow (W)
 \dot{q}_{swine} = total heat generation rate from swine (W)
 \dot{q}_{tp} = total power consumption of two pit fans (W)
 r_{ae} = pit air exchange ratio (dimensionless)
 r_r = recirculation ratio (dimensionless)
 S_{heater} = switch function of heater (dimensionless)
 T_f = floor temperature (K)
 T_{film} = manure film temperature (K)
 T_g = pit wall temperature (K)
 T_o = outdoor temperature (K)
 T_p = pit headspace temperature (K)
 T_r = room temperature (K)
 TAN = total ammoniacal nitrogen ($mol\ m^{-3}$)
 TC = total operating cost ($\$$)
 U_{pw} = heat transfer coefficient of pit and ground ($W\ m^{-2}\ K^{-1}$)
 U_{rw} = heat transfer coefficient between room and wall ($W\ m^{-2}\ K^{-1}$)
 U_{rf} = heat transfer coefficient of floor and room ($W\ m^{-2}\ K^{-1}$)
 V_p = pit headspace volume (m^3)
 $V_{p,max}$ = maximum volume of pit headspace (m^3)

$V_{p,min}$ = minimum volume of pit headspace (m^3)
 V_r = room volume (m^3)
 v_s = air speed in the vicinity of the slurry (m s^{-1})
 Z = empirically determined constant
 Z_{NH_3} = coefficient determined by fitting modeled concentration estimates to field measurements ($\text{K}^{1.4} \text{m}^{0.2} \text{s}^{-0.2}$)
 η_P = removal efficiency of the air cleaner (dimensionless)
 ρ_a = air density (kg m^{-3})
 ρ_{CO_2} = density of CO_2 (kg m^{-3})

APPENDIX

MODEL EQUATIONS AND PARAMETERS FOR TEMPERATURE AND OPERATING COST

Equations A1 and A2 describe the energy balances for the room volume and the pit volume (Park et al., 2013).

Room:

$$\begin{aligned} \rho_a V_r c_a \frac{dT_r}{dt} = & \rho_a Q_{ac} c_a T_r + \rho_a Q_{tp} c_a T_p \\ & + \rho_a Q_{ae} c_a T_p + \dot{q}_{gen} - \rho_a Q_{ac} c_a T_r \\ & - \rho_a Q_{tp} c_a T_r - \rho_a Q_{ae} c_a T_r \\ & - U_{rw} A_{rw} (T_r - T_o) - U_{rf} A_{rf} (T_r - T_f) \end{aligned} \quad (\text{A1})$$

Pit:

$$\begin{aligned} \rho_a V_p c_a \frac{dT_p}{dt} = & \rho_a Q_{tp} c_a T_r + \rho_a Q_{ae} c_a T_r \\ & - \rho_a Q_{tp} c_a T_p - \rho_a Q_{ae} c_a T_p \\ & - U_{pw} A_{pw} (T_p - T_g) \end{aligned} \quad (\text{A2})$$

In our previous model, outdoor temperature (T_o) was simulated using historical seasonal average data for Cedar Rapids, Iowa, modeled as a combination of two sine waves (Park et al., 2013; Anthony et al., 2014) to account for within-day and between-day temperatures. However, for the current model, T_o was set to the actual temperature of Cedar Rapids, Iowa, from December 2013 to February 2014, the period of the field study for model validation (CID Airport meteorological data from NOAA's National Climatic Data Center).

The total operating cost was computed using equation A3, which included continuous operation of the heat lamps, the cost of running the heater to maintain the room temperature within the optimum production range, and the cost of running contaminant control equipment during each test case using power requirements from the device manufacturers (Park et al., 2013):

$$\begin{aligned} TC = & P_{elect} \int_0^{3\text{months}} (\dot{q}_{lamp} + \dot{q}_{tp} + \dot{q}_{ac}) dt \\ & + P_{gas} \int_0^{3\text{months}} (S_{heater} \dot{Q}_{heater}) dt \end{aligned} \quad (\text{A3})$$

where P_{elect} is the electricity cost, P_{gas} is the natural gas cost, \dot{q}_{tp} is the power consumption of the pit fans, and \dot{q}_{ac} is the power consumption of the air cleaner. The switch function of heater operation (S_{heater}) was computed by the model as 0 (off) or 1 (on) at any moment in time as determined by the need for the heater to activate to warm the room, based on computed room temperatures. Tables 1 and A1 detail each parameter used in these equations.

Table A1. Input parameters for energy and cost equations.

Parameter	Value	Note
Air density (ρ_a)	1.2041 kg m^{-3}	Assumed dry air at 20.0°C and 101.325 kPa
Specific heat at constant pressure of air (C_p)	1,006.1 $\text{J kg}^{-1} \text{K}^{-1}$	Assumed dry air at 20.0°C and 101.325 kPa
Outdoor temperature (T_o)	-	Data measured at Eastern Iowa Airport during Dec. 2013 to Feb. 2014
Room temperature (T_r)	-	Computed in equation 2; initial value = 293 K
Pit headspace temperature (T_p)	-	Computed in equation 2; initial value = 293 K
Pit wall temperature (T_g) and floor temperature (T_f)	290 K (17°C)	Assumed $T_g = T_f$; field measurement data
Manure film temperature (T_{film})	-	$= (T_g + T_p) / 2$
Heat generation rate from one piglet (\dot{q}_{piglet})	24.5 W	$= m_{piglet} \times (4.3 \times m_{piglet}^{0.15})$ (Brown-Brandl et al., 2004)
Heat generation rate from one sow (\dot{q}_{sow})	372.6 W	$= m_{sow} \times (14.11 \times m_{sow}^{-0.38})$ (Brown-Brandl et al., 2004)
Total heat generation rate from swine (\dot{q}_{swine})	-	$= n_{piglet} \times \dot{q}_{piglet} + n_{sow} \times \dot{q}_{sow}$
Total heat generation rate from 20 lamps (\dot{q}_{lamp})	2,500 W	$= 20 \times 125 \text{ W}$, per manufacturer
Heat generation rate from one gas heater (\dot{q}_{heater})	17,585 W	$= 60,000 \text{ BTU h}^{-1}$ per manufacturer; on: $T_r \leq 20.0^\circ\text{C}$; off: $T_r \geq 22.2^\circ\text{C}$
Total heat generation rate in the room (\dot{q}_{gen})	-	$= \dot{q}_{heater} + \dot{q}_{swine} + \dot{q}_{lamp}$
Heat transfer coefficient between room and wall (U_{rw})	0.286 $\text{W m}^{-2} \text{K}^{-1}$	U-value of ceiling was assumed to be same as U_{rw} (Zhang et al., 1993)
Heat transfer coefficient of floor and room (U_{rf})	0.568 $\text{W m}^{-2} \text{K}^{-1}$	Zhang and Barber (1993)
Heat transfer coefficient of pit and ground (U_{pw})	0.568 $\text{W m}^{-2} \text{K}^{-1}$	Assumed $U_{rf} = U_{pw}$ (Zhang and Barber, 1993)
Overlap area of room and wall (A_{rw})	238.3 m^2	$= l_r \times w_r + 2 \times (l_r + w_r) \times h_r$
Overlap area of room and floor (A_{rf})	54.6 m^2	$= l_r \times w_r - (4 \times l_p \times w_p)$
Overlap area of pit and wall (A_{pw})	96.6 to 147.3 m^2	$= 4 \times [(l_p \times w_p) + 2 \times (l_p + w_p) \times h_p]$
Electricity cost (P_{elect})	\$0.0807 kWh^{-1}	Average industrial price in Iowa during Dec. 2013 to Feb. 2014
Natural gas cost (P_{gas})	\$0.27 m^{-3}	Average industrial price in Iowa during Dec. 2013 to Feb. 2014
Natural gas consumption of one gas heater (\dot{Q}_{heater})	0.000472 $\text{m}^3 \text{s}^{-1}$	Per manufacturer
Total power consumption of two pit fans (\dot{q}_{tp})	690 W	$= 2 \times 345 \text{ W}$, per manufacturer
Power consumption of air cleaner (\dot{q}_{ac})	4,950 W	Per manufacturer
Switch function of heater (S_{heater})	0, 1	0 = off; 1 = on

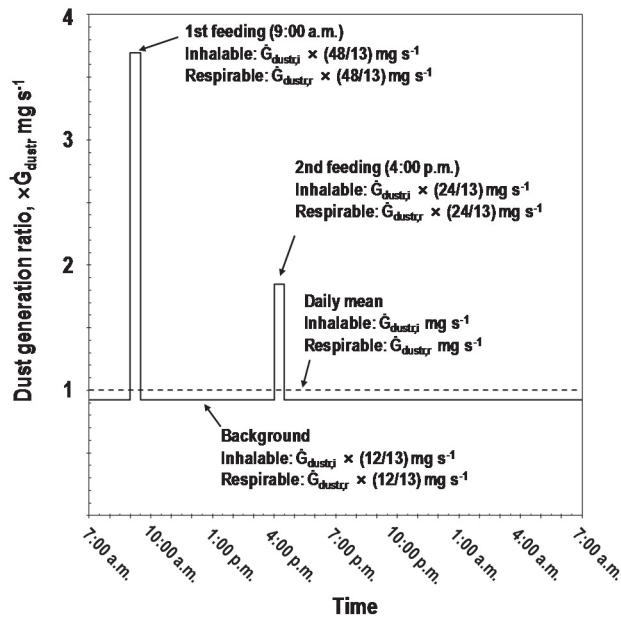


Figure A1. Time-dependent dust generation rates.

Table A2. Swine numbers.

Date and Time	Simulation Time		Sows	Piglets
	Days	Seconds		
1 Dec. 2013, 8:00 h	0.00	0	0	0
10 Dec. 2013, 0:00 h	8.67	748,800	11	0
13 Dec. 2013, 8:00 h	12.00	1,036,800	11	64
14 Dec. 2013, 8:00 h	13.00	1,123,200	11	63
16 Dec. 2013, 8:00 h	15.00	1,296,000	11	64
17 Dec. 2013, 8:00 h	16.00	1,382,400	11	64
18 Dec. 2013, 8:00 h	17.00	1,468,800	11	68
19 Dec. 2013, 8:00 h	18.00	1,555,200	11	69
21 Dec. 2013, 8:00 h	20.00	1,728,000	11	77
22 Dec. 2013, 8:00 h	21.00	1,814,400	11	79
26 Dec. 2013, 8:00 h	25.00	2,160,000	11	74
27 Dec. 2013, 8:00 h	26.00	2,246,400	11	74
31 Dec. 2013, 8:00 h	30.00	2,592,000	0	0
10 Jan. 2014, 8:00 h	40.00	3,456,000	16	0
11 Jan. 2014, 8:00 h	41.00	3,542,400	16	0
17 Jan. 2014, 8:00 h	47.00	4,060,800	15	11
18 Jan. 2014, 8:00 h	48.00	4,147,200	15	8
20 Jan. 2014, 8:00 h	50.00	4,320,000	15	18
22 Jan. 2014, 8:00 h	52.00	4,492,800	16	30
23 Jan. 2014, 8:00 h	53.00	4,579,200	16	57
24 Jan. 2014, 8:00 h	54.00	4,665,600	13	54
25 Jan. 2014, 8:00 h	55.00	4,752,000	13	76
26 Jan. 2014, 8:00 h	56.00	4,838,400	13	75
27 Jan. 2014, 8:00 h	57.00	4,924,800	13	85
28 Jan. 2014, 8:00 h	58.00	5,011,200	13	91
29 Jan. 2014, 8:00 h	59.00	5,097,600	13	99
3 Feb. 2014, 8:00 h	64.00	5,529,600	17	120
4 Feb. 2014, 8:00 h	65.00	5,616,000	17	119
10 Feb. 2014, 8:00 h	71.00	6,134,400	19	82
11 Feb. 2014, 8:00 h	72.00	6,220,800	19	86
17 Feb. 2014, 8:00 h	78.00	6,739,200	19	117
18 Feb. 2014, 8:00 h	79.00	6,825,600	19	117
24 Feb. 2014, 8:00 h	85.00	7,344,000	17	96
25 Feb. 2014, 8:00 h	86.00	7,430,400	17	95
26 Feb. 2014, 8:00 h	87.00	7,516,800	17	102
27 Feb. 2014, 8:00 h	88.00	7,603,200	17	100

FEEDING FUNCTION

The overall mean inhalable and respirable dust generation rates were 567 and 59 mg h^{-1} per 500 kg swine mass, respectively (Takai et al., 1998). The mean generation rates of inhalable ($\dot{G}_{dust,i}$) and respirable ($\dot{G}_{dust,r}$) dust were calculated from total swine mass. To account for the fact that dust generation depends on feeding time, dust generation rate was assumed to increase during feeding, as shown in figure A1. The first feeding was prescribed at 9:00 a.m., and the second feeding occurred at 4:00 p.m.. Dust concentrations at the first and second feedings were modeled as four and two times higher than background concentration, respectively, based on previous research (O'Shaughnessy et al., 2009).

INPUT CONDITIONS AND SIMULATION RESULTS

The following tables identify the field conditions at the farrowing barn at the test site that was used to validate the model. These tables detail how the model was adjusted throughout the 90-day simulation period to represent actual conditions at the test location. Table A2 identifies the animals housed in the room (changed over time). Table A3 identifies when the ventilation system was turned on or off (to assess system performance at the test location). Table A4 identifies room-averaged concentrations, measured in the field and modeled, over the testing and simulation periods.

SEE NEXT PAGE FOR TABLE A4.

Table A3. Air cleaner operation.

Date and Time	Simulation Time		Air Cleaner (0 = off, 1 = on)
	Days	Seconds	
1 Dec. 2013, 8:00 h	0.00	0	0
21 Dec. 2013, 9:00 h	20.04	1,731,600	1
22 Dec. 2013, 9:00 h	21.04	1,818,000	1
26 Dec. 2013, 9:00 h	25.04	2,163,600	1
27 Dec. 2013, 9:00 h	26.04	2,250,000	1
31 Dec. 2013, 9:00 h	30.04	2,595,600	1
1 Jan. 2014, 9:00 h	31.04	2,682,000	1
11 Jan. 2014, 9:00 h	41.04	3,546,000	1
12 Jan. 2014, 9:00 h	42.04	3,632,400	1
17 Jan. 2014, 9:00 h	47.04	4,064,400	1
18 Jan. 2014, 9:00 h	48.04	4,150,800	1
20 Jan. 2014, 9:00 h	50.04	4,323,600	1
21 Jan. 2014, 9:00 h	51.04	4,410,000	0
28 Jan. 2014, 9:00 h	58.04	5,014,800	1
29 Jan. 2014, 9:00 h	59.04	5,101,200	1
3 Feb. 2014, 9:00 h	64.04	5,533,200	1
4 Feb. 2014, 9:00 h	65.04	5,619,600	1
10 Feb. 2014, 9:00 h	71.04	6,138,000	1
11 Feb. 2014, 9:00 h	72.04	6,224,400	1
17 Feb. 2014, 9:00 h	78.04	6,742,800	1
18 Feb. 2014, 9:00 h	79.04	6,829,200	1
24 Feb. 2014, 9:00 h	85.04	7,347,600	1
25 Feb. 2014, 9:00 h	86.04	7,434,000	0

Table A4. Model simulation and field measurement results for IAQ ($Z_{NH_3} = 21,000$, makeup air CO_2 concentration = 1,500 ppm).

Date	Air Cleaner	24 h Mean T_o (°C)	Sow/Piglet Count	NH ₃ (ppm)		CO ₂ (ppm)		CO (ppm)		Inhalable Dust (mg m ⁻³)		Respirable Dust (mg m ⁻³)	
				Model	Field	Model	Field	Model	Field	Model	Field	Model	Field
13 Dec. 2013	Off	-5.3	11/64	8.57	5.02 ±1.78	2500	2177 ±92	0.31	1.58 ±1.08	0.94	0.87 ±0.14	0.10	0.20 ±0.03
16 Dec. 2013	Off	-7.9	11/63	8.62	5.73 ±1.52	2501	2440 ±101	0.31	2.06 ±0.89	0.94	1.15 ±0.31	0.10	0.22 ±0.05
18 Dec. 2013	Off	-0.6	11/68	8.59	6.05 ±1.34	2418	2029 ±130	0.27	1.38 ±0.46	0.94	1.38 ±0.37	0.10	0.17 ±0.04
21 Dec. 2013	On	-3.2	11/77	8.73	6.19 ±1.79	2523	2241 ±145	0.31	1.86 ±0.73	0.64	0.96 ±0.37	0.07	0.12 ±0.02
26 Dec. 2013	On	-4.8	11/74	8.73	10.53 ±3.59	2498	2251 ±181	0.30	2.89 ±2.51	0.62	1.07 ±0.54	0.07	0.11 ±0.01
31 Dec. 2013	On	-13.4	0/0	0.05	0.07 ±0.16	2103	2276 ±100	0.31	3.29 ±2.54	0.00	0.18 ±0.01	0.00	0.10 ±0.01
10 Jan. 2014	On	0.1	16/0	10.87	2.50 ±1.24	2410	1990 ±116	0.26	1.73 ±0.99	0.78	0.41 ±0.06	0.08	0.09 ±0.01
17 Jan. 2014	On	-11.6	15/11	10.63	9.79 ±4.40	2515	2614 ±79	0.31	2.46 ±1.57	0.74	0.29 ±0.04	0.08	0.10 ±0.01
20 Jan. 2014	On	-7.0	15/18	10.65	12.53 ±3.78	2478	2250 ±176	0.29	1.39 ±0.41	0.75	0.50 ±0.31	0.08	0.15 ±0.09
22 Jan. 2014	Off	-16.5	16/30	11.73	12.38 ±2.78	2574	2801 ±99	0.31	2.00 ±0.13	1.25	0.78 ±0.15	0.13	0.20 ±0.00
24 Jan. 2014	Off	-2.7	13/54	9.77	10.54 ±1.11	2473	2254 ±160	0.28	1.44 ±0.16	1.07	0.60 ±0.07	0.11	0.16 ±0.01
26 Jan. 2014	Off	-10.3	13/75	10.24	14.73 ±1.40	2529	2394 ±145	0.29	1.51 ±0.22	1.11	0.82 ±0.11	0.12	0.18 ±0.01
28 Jan. 2014	On	-17.1	13/91	10.62	16.48 ±1.23	2600	2907 ±141	0.31	2.11 ±0.32	0.75	0.47 ±0.17	0.08	0.11 ±0.05
3 Feb. 2014	On	-10.9	17/120	13.64	16.58 ±3.84	2755	2759 ±158	0.31	1.87 ±0.30	0.96	0.83 ±0.28	0.10	0.13 ±0.01
10 Feb. 2014	On	-23.1	19/82	14.83	28.09 ±1.83	2743	3103 ±130	0.31	2.27 ±0.26	1.02	0.92 ±0.15	0.11	0.14 ±0.01
17 Feb. 2014	On	-3.1	19/117	14.82	8.73 ±1.30	2675	2505 ±156	0.25	0.94 ±0.46	1.06	1.06 ±0.15	0.11	0.09 ±0.01
24 Feb. 2014	On	-11.0	17/96	13.29	10.01 ±2.14	2715	2639 ±91	0.31	1.34 ±0.68	0.94	0.85 ±0.21	0.10	0.13 ±0.01
26 Feb. 2014	Off	-13.0	17/102	13.44	7.25 ±1.40	2725	3007 ±136	0.31	1.45 ±0.81	1.45	1.36 ±0.28	0.15	0.25 ±0.01