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REVIEW ARTICLE

Occupational exposure limits for manufactured nanomaterials, a systematic review

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ABSTRACT

Background: The toxicological properties of manufactured nanomaterials (MNM)s can be different from their bulk-material and uncertainty remains about the adverse health effects they may have on humans. Proposals for OELs have been put forward which can be useful for risk management and workers' protection. We performed a systematic review of proposals for OELs for MNMs to better understand the extent of such proposals, as well as their derivation methods.

Methods: We searched PubMed and Embase with an extensive search string and also assessed the references in the included studies. Two authors extracted the data independently.

Results: We identified 20 studies that proposed in total 56 OEL values. Of these, two proposed a generic level for all MNMs, 14 proposed a generic OEL for a category of MNMs and 40 proposed an OEL for a specific nanomaterial. For specific fibers, four studies proposed a similar value but for carbon nanotubes (CNTs) the values differed with a factor ranging from 30 to 50 and for metals with a factor from 100 to 300. The studies did not provide explanations for this variation. We found that exposure to MNMs measured at selected workplaces may exceed even the highest proposed OEL. This indicates that the application and use of OELs may be useful for exposure reduction.

Conclusion: OELs can provide a valuable reference point for exposure reduction measures in workplaces. There is a need for more and better supported OELs based on a more systematic approach to OEL derivation.

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Background

Nanotechnology is an expanding field, with new manufactured nanomaterials (MNM)s and products containing these materials appearing on the market every year. There is also a growing diversity of industries in which MNMs are used such as construction, health care, energy, automobile and aerospace, chemical products, electronics and communication, cleaning and maintenance, textile, and military. This means that a growing number of workers worldwide are potentially exposed to MNMs (Kaluza et al., 2009). Systematic reviews of exposure studies confirm that workplace exposure to nanoparticles occurs and that control measures can be improved to reduce exposure (Debia et al., 2016; Ding et al., 2017).

Nanoparticles can be classified into three categories. There are naturally occurring nanoparticles resulting from the nucleation of low-volatile gas-phase compounds followed by growth into small particles such as via volcanic eruptions and forest fires, via erosion. Then, there are also incidental nanoparticles generated (as by-products) of heating and combustion processes, machining and other high-energy processes, also called process-generated nanoparticles or combustion-derived nanoparticles (Donaldson et al., 2005). Finally there are man-made manufactured nanoparticles intentionally produced by industry, as defined by the European Commission (EC, 2011), such as carbon nanotubes. Since different sources of nanoparticles require different approaches, in this article we focus only on the third category, i.e. MNMs in the workplace.

Many MNMs are still given the name of their bigger chemical bulk material, but due to their extremely small size (≤ 100 nm),

their physical and chemical properties can be different from those of the "mother" material of the same structure and composition (Kulinowski & Bruce, 2011). The International Organization for Standardization (ISO) defined bulk material as "a material of the same chemical composition as nano-objects and their agglomerates and aggregates (NOAAs), at a scale greater than the nanoscale."

The toxicity of MNMs largely depends on numerous physico-chemical properties, including size, shape, composition, surface characteristics, charge and solubility. While workers may be exposed to MNMs via inhalation, ingestion or dermal absorption, the inhalation pathway is the most likely to result in larger systemic doses (Oberdorster et al., 2005). Once inhaled, the mechanisms, pattern and efficiency of particle deposition in the respiratory tract remains a function of its aerodynamic diameter, shape and density. Particles with a diameter from 1 to 100 nm show a much higher fraction of deposition in the pulmonary region of the lung compared to larger particles. Of inhaled particles with various diameters, only those in the nanorange are known to systematically translocate from the lungs into the circulatory system through the air-blood tissue barrier and subsequently accumulate in secondary organs and tissues of the body (Geiser & Kreyling, 2010). Due to inherent ethical concerns, most evidence comes from rodent studies, with the most reliable data using quantitative particle biokinetics assessments, which balance the total nanoparticle fractions as measured in the rodent body and total excretion collected between application and autopsy (Geiser & Kreyling, 2010). In a study by Semmler-Behnke et al. with nanosized ¹⁹²Iridium, it was confirmed that nanoparticles

are predominantly retained long-term within interstitial spaces of the alveolar region of the rat lung, with limited translocation toward the circulation (Semmler-Behnke et al., 2007). A series of studies of particle inhalation in rodents has shown that nanoparticle translocation into the circulation and to secondary organs remains highly dependent on the nanoparticle physicochemical properties, including size, material, surface charge and surface modifications (Kreyling et al., 2002, 2009; Semmler et al., 2004). There is currently no evidence of the nanoparticle translocation to the circulation and to secondary organs beyond 1% of the mass-based dose (Kreyling et al., 2014; Mills et al., 2006; Wiebert et al., 2006). However, this figure is based on extrapolation from animal studies, resulting in the lack of precise information for inhaled MNM bio-kinetics and long-term results in the human model. Nevertheless, while acute effects from MNM translocation to secondary organs are not likely to be considerable, it is possible that chronically exposed populations may face greater risks from cumulative, low-dose translocation processes, for example from biopersistent MNMs.

Occupational Exposure Limits (OEL) for chemical substances have long been in use for controlling workplace exposures. In 1887, Germany was the first country to publish selected limit values that were considered occupational exposure limits, but it was only in 1977 that the term had been fully adopted by the International Labor Organization (ILO) and later, in 1981, that the World Health Organization (WHO) started to use the same term: occupational exposure limits (Schulte et al., 2010).

ISO defines OELs as a "maximum concentration of airborne contaminants deemed to be acceptable, as defined by the authority having jurisdiction" (ISO 16972:2010).

Even though there is no generally accepted uniform definition for an OEL, there is at least agreement that they constitute a level of usually airborne exposure to an agent beyond which unacceptable health risks might occur. In this general sense, we will also use the term OEL in this article. OELs are commonly established based on the actual state of the scientific knowledge and intended for protecting against adverse health effects for workers. Health-based OELs are usually based on the estimation of a no effect level and therefore they represent an exposure level below which no adverse health effects are expected (Stouten et al., 2008). However, for genotoxic and carcinogenic substances, that have no threshold below which there is no detectable effect, some countries, for example the Netherlands and Germany, have developed what they call risk-based OELs (B A u A, 2013; Ding et al., 2014; S E R, 2007). These risk concepts define a *tolerable risk* level and an *acceptable risk* level. Germany defines a tolerable risk level as a calculated additional cancer risk of 4:1000, meaning that statistically 4 out of 1000 persons exposed to the substance during their working life may develop cancer, and they define an acceptable risk level as a calculated cancer risk level of 4:10 000 (until 2013) and 4:100 000 (at the latest in 2018). In the Netherlands, the levels are respectively 1:10 000 and 1:1 000 000 (new cancers per year). This means that these countries acknowledge that no safe level can be defined, that even the lowest exposures may induce an adverse effect and accept that a certain number of workers may develop cancer each year as a result of the exposure. On the other hand, countries also derive OELs that include technical and economic feasibility considerations for regulatory purposes and these are thus not entirely health or risk based and these are sometimes called administrative OELs. The naming of OELs is quite inconsistent between different national and international bodies.

Within the REACH framework the EU defines derived no-effect levels (DNEL) for substances with a detectable threshold for

health-based effects (European Chemical Agency, 2012). For genotoxic carcinogenic substances, without a threshold effect, the EU defines derived minimal effect levels (DMEL), which is a semi-quantitative value. In the US, the American Conference of Governmental Industrial Hygienists (ACGIH) defines threshold limit values (TLV) for health-based OELs. In their local context, these values may have a distinct and complementary meaning. The values may have been established based on what was considered to be technically feasible, or may have been calculated with the goal of preventing adverse health effects, or for limiting the potential number of health effects. These factors make it complex to compare the values, although efforts have been made to harmonize procedures for deriving OELs (Deveau et al., 2015).

Regardless of the difficulty of unambiguously defining OELs, they form an important tool for occupational risk management within a health context. They provide a rationale for risk assessment and control measures. Based on long-term analysis of exposures at the workplace, Creely et al. argued that regulation, including the establishment of OELs, has led to a decrease in workplace exposure to a number of hazardous chemicals (Creely et al., 2007).

Moreover, according to common principles in behavioral theory, formulating a goal that has to be achieved is a strong driver for desirable behavior (Locke & Latham, 2002).

Practice shows that OELs for chemical substances in general must be regarded as being provisional, requiring regular updating to comply with growing knowledge of the hazards. Therefore, insufficient scientific evidence should not be a barrier to accept provisional OELs but on the contrary ask for the operationalization of the existing knowledge and if necessary for the application of precautionary measures. In fact, international experts advocate the development of provisional OELs for MNMs (Gordon et al., 2014; van Broekhuizen & Dorbeck-Jung, 2013).

Currently, specific regulatory OELs for MNMs have not been established by the EU or by any national authority and it is expected that it may take a long time before OELs have been derived for all highly diverse frequently used MNMs. This is mainly due to the still existing large gaps in knowledge on particle toxicology, the high diversity of the newly developed, and used, MNMs, the uncertainties about their hazardous nature and the ongoing discussions on the metrics to be used for the nano-OELs, be it mass-based or particle number based. Alternatively, generic precautionary particle number-based nano-reference values (NVR) for groups of nanomaterials have been proposed in some countries (I F A, 2009; S E R, 2012). Here the adjective "reference" is used to emphasize that these values are not health-based and indicates that these values should be for risk management: as an incentive to take control measures if the NVR is exceeded (S E R, 2012; van Broekhuizen & Dorbeck-Jung, 2013).

For a few specific nanomaterials the industry and research have advised an OEL or a DNEL. NIOSH (2011) proposed an OEL for nano-TiO₂ based on toxicological data and used the US threshold limit value (TLV) for coarse TiO₂ (of 1.5 mg/m³) as a reference. Bayer (Pauluhn, 2010), Nanocyl (Luizi, 2009) and NIOSH (National Institute for Occupational Safety and Health, 2013) proposed OELs for multiwall carbon nanotubes (MWCNTs). DNELs were calculated in an experimental study by Stone et al. applying the DNEL methodology with the prescribed assessment factors to MWCNTs, fullerenes, silver (Ag) and titanium dioxide (TiO₂) (Stone, 2009).

Currently, the World Health Organization is preparing a guideline for protecting workers from potential risks of MNMs. One of the questions is: which OEL/reference value should specific nanomaterials or groups of materials be assigned to? So far, there has been limited information on the development and use of OELs for

MNMs (Gordon et al., 2014; Schulte et al., 2010). To address this problem, we conducted a systematic review of existing OELs for MNMs and analyzed how these values were derived.

Objective

To develop an exhaustive list of OELs that have been proposed for MNMs, and to describe differences and similarities in the approaches by which they were derived.

Methods

Inclusion criteria

We based our inclusion criteria on the PICO approach, which is an acronym that specifies that eligible studies must comply with criteria for one or more of the following elements: participants (P), intervention/exposure (I/E), control (C), outcome (O), and study design (S) (Guyatt et al., 2011; Morgan et al., 2016). These criteria were defined as follows.

Study design

We included all the proposals using an exposure limit approach or that proposed a quantitative exposure limit value for an MNM or a group of MNMs for protecting workers exposed to manufactured nanomaterials from adverse health effects. To be included, the studies also had to indicate the process by which the authors derived the OELs.

Participants

The OEL is a tool intended to protect workers potentially exposed to MNMs.

Intervention/exposure

The OEL should be formulated as a concrete exposure value for a MNM or group of MNMs and should address the MNMs' potential for adverse health effects and it should indicate how the exposure should be measured and expressed.

We considered that the control (C) and outcome (O) elements were not applicable in our specific situation where we are not looking for effects of controlled studies but where we want to list a specific set of OEL proposals

Search methods for inclusion of studies

Electronic searches

We conducted a systematic literature search in PubMed and Embase until 15 February 2016, which was not limited to the English language. The search string contained specific search words for MNMs, such as nanomaterial and synonyms, occupational exposure limit and synonyms, and OELs. We combined both search strings with AND (see Supplementary Appendix 1 for the full search strategy).

Searching other sources

We checked the reference lists of all included studies to find additional proposals. We also asked experts involved in the development of the WHO Guidelines on protecting workers from potential risks of manufactured nanomaterials (draft, WHO 2016) or one of the systematic reviews of the WHO guideline to report any proposed OELs for MNMs that they knew of.

Analysis

We grouped the OELs per MNM or group of MNMs and analyzed per OEL which process was used to derive the OEL value. Next,

we categorized the derivation processes according to Gordon 2014 (Gordon et al., 2014), which we slightly adapted as:

- Traditional quantitative risk assessment (QRA) defined as a stepped approach that starts with assessing toxicological data for substance and selecting a dose – usually a no-observed-adverse-effect-level or benchmark dose to use as a point of departure to calculate a human equivalent concentration and by applying various uncertainty and modifying factors finally arriving at an OEL.
- Bridging or read across defined as applying hazard information of one material (nano- or bulk material) to predict the hazards of another material (Oomen et al., 2015; Patlewicz et al., 2013). Even though the method has been advocated for bulk materials to save time, money and animals, there is no consensus on how to do this (Patlewicz et al., 2013).
- Using environmental exposure limits for particulate matter (World Health Organisation, 2005)
- Grouping defined as an approach that groups MNMs based on a common aspect of the material (Oomen et al., 2015). Even though grouping should be based on similar principles as read-across, we believe that it is important to distinguish grouping from read-across for one material because of its practical consequences.

Data collection

Two authors (RM, JV) independently extracted the following data per proposal into an Excel sheet: MNM, value(s), measurement metric(s), approach (how were the OELs derived), year of development, country, category of development, key study.

Risk of bias assessment

We did not try to assess the risk of bias in the development process since there are no generally accepted methods to derive OELs.

Results

Results of the searches

Our systematic searches resulted in 498 references. The search in MEDLINE/PubMed resulted in 259 references and the search in Embase in 239, altogether 498 references. In addition, we located 23 potential references from other sources. After removing the duplicates this resulted in 397 references that we screened for inclusion based on title and abstract. This resulted in 73 references that we checked for inclusion based on full text assessment. After the exclusion of those ($n=49$) that did not meet our inclusion criteria, we included 24 articles. To prevent double counting of studies, we aggregated articles that described the same values into one study. For example, we aggregated van Broekhuizen (2011, 2012, 2013) and the German Institute of Occupational Safety and Health (IFA) (2016) into one study van Broekhuizen (2012) because IFA (2016) explicitly referred to van Broekhuizen as the source of their values (Institut für Arbeitsschutz der Deutschen Gesetzlichen Unfallversicherung, 2016; van Broekhuizen & Dorbeck-Jung, 2013; van Broekhuizen & Reijnders, 2011; van Broekhuizen et al., 2012). This finally resulted in the inclusion of 20 studies. Some groups updated their proposals for OELs over time. In this case, we only included the most recent reported value. Many studies included more than one proposal and this resulted in 56 proposals for OELs. See the flow diagram (Figure 1).

We have attached also a comprehensive list with all the full text articles that have been screened and included (Supplementary Appendix 2).

Description of included studies

See Table 1 for a description of included studies.

Nanomaterials addressed

Studies with a general approach. We found two studies that took a generic approach and proposed an OEL for all MNMs. In one study, the OEL was based on environmental exposure limits for particulate matter (PM₁₀) (Guidotti, 2010). In the other study, the OEL was based on the number of times the potential MNM exposure concentration exceeded the local background level (McGarry et al., 2013).

Studies with a categorical approach. We found six studies that used a categorical approach when they derived an OEL for a group of nanomaterials (British Standards Institution, 2007; German Hazardous Substances Committee, A. & BAuA, 2013; Kuempel et al., 2012; Pauluhn, 2011; Stockmann-Juvala et al., 2014; van Broekhuizen et al., 2012). Groups were: fibers, granular biopersistent particles (GBP), MNMs with bulk material classified as CMAR-chemicals (carcinogenic, mutagenic, asthmagenic, reproductive risk), MNMs that are soluble, and MNMs that are non-biopersistent.

Studies with a MNM specific approach. Most studies evaluated specific MNMs. There were seven that evaluated TiO₂ (Aschberger et al., 2011; Kuempel et al., 2006; National Institute for Occupational Safety and Health, 2011; Ogura et al., 2011; Stockmann-Juvala et al., 2014; Swidwinska-Gajewska & Czerczak, 2014; Warheit, 2013), six that evaluated carbon nanotubes (Aschberger et al., 2011; Luizi, 2009; Nakanishi et al., 2015; National Institute for Occupational Safety and Health, 2013;

Pauluhn, 2010; Stone, 2009), three evaluated fullerene (Aschberger et al., 2010; Shinohara et al., 2011; Stone, 2009), three evaluated nanosilver (Aschberger et al., 2011; Stone, 2009; Swidwinska-Gajewska & Czerczak, 2015), and one study evaluated amorphous SiO₂, low-toxicity dust, nanocellulose and nanoclays (Stockmann-Juvala et al., 2014).

Routes of exposure

All proposals addressed chronic inhalation exposure of the workers. One study also evaluated dermal and oral exposure to carbon nanotubes and fullerene (Stone, 2009). Stone 2009 and Aschberger et al., 2010 also derived OEL values for short-term (15 min) inhalation exposure of the workers (Aschberger et al., 2010; Stone, 2009).

There were 15 studies which used traditional quantitative risk assessment (QRA) (Aschberger et al., 2010, 2011; Kuempel et al., 2006; Luizi, 2009; Nakanishi et al., 2015; National Institute for Occupational Safety and Health, 2011; National Institute for Occupational Safety and Health, 2013; Ogura et al., 2011; Pauluhn, 2010; Shinohara et al., 2011; Stockmann-Juvala et al., 2014; Stone, 2009; Swidwinska-Gajewska & Czerczak, 2014, 2015; Warheit, 2013). There were all together six studies that used bridging or read across from short-term *in vivo* studies as follows. Three studies adjusted OELs that exist for the larger counterpart bulk material (British Standards Institution, 2007; van Broekhuizen et al., 2012; Institut für Arbeitsschutz der Deutschen Gesetzlichen Unfallversicherung, 2016); four studies used a bridging and a grouping approach (British Standards Institution, 2007; German Hazardous Substances Committee, A. & BAuA, 2013; Stockmann-Juvala et al., 2014; van Broekhuizen et al., 2012); three studies used only a bridging approach (British Standards Institution, 2007; Stockmann-Juvala et al., 2014; van Broekhuizen et al., 2012); two used only a grouping approach (German Hazardous Substances Committee, A. & BAuA, 2013; van Broekhuizen et al., 2012). Then there were two studies that used environmental exposure limits for particulate matter (Guidotti, 2010; McGarry et al., 2013), and

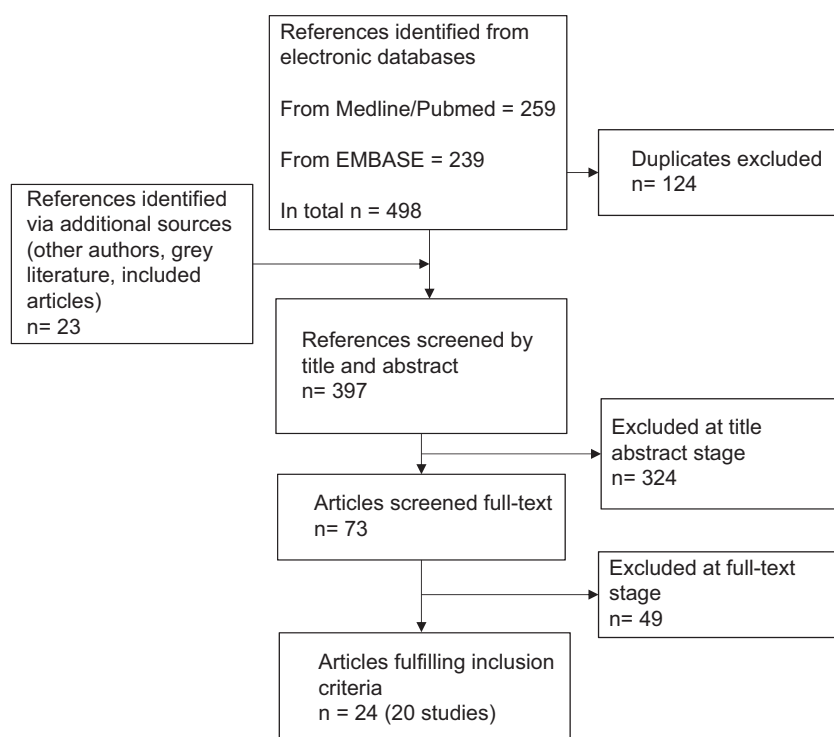


Figure 1. Flow of references, full-text assessment and inclusion of studies.

Table 1. Description of studies proposing OELs included in the review (N = 20).

Study reference	Professional Group/Institution	Funded by	Country	Nanomaterial(s)
German Hazardous Substances Committee, A. & BAuA (2013)	German Hazardous Substances Committee, German Federal Institute for Occupational Safety and Health (BAuA)	National Institute	Germany	Granular biopersistent particles and non-entangled fibrous nanomaterials
Aschberger et al. (2011)	ENRHES project 2009	European project	EU	Carbon nanotubes (multi-walled), fullerenes, nanosilver and nano titanium dioxide
BSI (2007)	British Standards Institution	National Institute	UK	Fibrous nanomaterials, CMAR, insoluble and soluble nanomaterials
Guidotti (2010)	Archives of Environmental and Occupational Health, journal	Independent	Canada	Environmental fine particulate matter ≤ 2500 nm
Kuempel et al. (2006)	National Institute for Occupational Safety and Health (NIOSH)	National Institute	USA	Titanium dioxide (ultrafine) and carbon black
Luizi (2009)	Nanocyl	Company	Belgium	Carbon nanotubes
McGarry et al. (2013)	International Laboratory for Air Quality and Health, Queensland University of Technology	University	Australia	Nanomaterials
Nakanishi et al. (2015)	New Energy and Industrial Technology Development Organization (NEDO)	National Institute	Japan	Carbon nanotube group: single-, double- and multi-walled nanotubes
NIOSH (2011)	National Institute for Occupational Safety and Health (NIOSH)	National Institute	USA	Titanium dioxide (ultrafine)
NIOSH (2013)	National Institute for Occupational Safety and Health (NIOSH)	National Institute	USA	Carbon nanotubes and carbon nanofibers
Ogura et al. (2011)	New Energy and Industrial Technology Development Organization (NEDO)	National Institute	Japan	Titanium dioxide
Pauluhn (2010)	Institute of Toxicology, Bayer Schering Pharmaceuticals	Company	Germany	Carbon nanotubes (multi-walled)
Pauluhn (2011)	Institute of Toxicology, Bayer Schering Pharmaceutical	Company	Germany	Inhaled poorly soluble particles
Shinohara et al. (2011)	New Energy and Industrial Technology Development Organization (NEDO)	National Institute	Japan	Carbon fullerenes (C_{60})
Stockmann-Juvala et al. (2014)	Scaffold SPD-7	European project	EU	Silicon dioxide (amorphous silica), titanium dioxide, carbon nanofibers, nanocellulose, nanoclays and low-toxicity dusts
Stone (2009)	ENRHES project 2009	European project	EU	Carbon nanotubes, fullerenes, metals and metal oxides
Swidwinska-Gajewska & Czerzak (2014)	Nofer Institute of Occupational Medicine, Lodz	National Institute	Poland	Titanium dioxide
Swidwinska-Gajewska & Czerzak (2015)	Nofer Institute of Occupational Medicine, Lodz	National Institute	Poland	Nanosilver
van Broekhuizen et al. (2012)	IVAM, University of Amsterdam	University	The Netherlands	Metals and metal oxides, biopersistent granular nanomaterial
Warheit (2013)	DuPont	Company	USA	Titanium dioxide (nanoscale)

one study that used both a categorical QRA and a grouping approach based on common aspects of MNMs (Pauluhn, 2011).

None of the studies was based on read across from *in-vitro* studies.

Geographical location and research groups

The included proposals were performed by a limited number of research groups. There were three studies funded by the EU: the ENHRES program (Engineered Nanoparticles: Review of Health and Environmental Safety) (Aschberger et al., 2011; Stone, 2009) and the Scaffold program (Stockmann-Juvala et al., 2014). There were ten studies conducted by national occupational health or technological research institutes. One study from the United Kingdom by the British Standards Institution (British Standards Institution, 2007), three studies from NIOSH in the United States (Kuempel et al., 2006; National Institute for Occupational Safety and Health, 2011; National Institute for Occupational Safety and Health, 2013), one from BAuA in Germany (German Hazardous Substances Committee, A. & BAuA, 2013), three from NEDO in Japan (Nakanishi et al., 2015; Ogura et al., 2011; Shinohara et al., 2011), and two from Poland (Swidwinska-Gajewska & Czerczak, 2014, 2015). There were two studies from universities, one from the Netherlands (van Broekhuizen et al., 2012) and the second one from Australia (McGarry et al., 2013). There were four studies by the chemical companies: Bayer (Pauluhn, 2010, 2011), BASF/Nanocyl (Luizi, 2009), and DuPont (Warheit, 2013). There was also one proposal by an individual editor of a journal in Canada (Guidotti, 2010).

Terminology used

Five research groups used the term occupational exposure limit (OEL) (Nakanishi et al., 2015; German Hazardous Substances Committee, A. & BAuA, 2013; Pauluhn, 2010; Stockmann-Juvala et al., 2014; Warheit, 2013). For one group this term had a regulatory meaning (German Hazardous Substances Committee, A. & BAuA, 2013) but not for the rest (Nakanishi et al., 2015; Pauluhn, 2010; Stockmann-Juvala et al., 2014; Warheit, 2013).

The four proposals by the Japanese research groups also used the term OEL but with a suffix indicating subchronic exposure spanning over 15 years called OEL period-limited or OEL PL (Nakanishi et al., 2015; Ogura et al., 2011; Shinohara et al., 2011).

There were two related groups which used the same data from the ENRHES 2009 project, but used different terms for the OELs. Both Stone et al. and Aschberger et al. applied the methodology as described in the appendix of the European Chemical Agency, REACH project, for deriving DNELs (European Chemical Agency, 2012; Tynkkynen et al., 2015), Aschberger and colleagues used an indicative no effect level (INEL), while Stone et al. used a derived no effect level (DNEL) (Aschberger et al., 2010, 2011; Stone, 2009). The reason for this was that the authors want to highlight that the derived INEL values should not be considered as having the same regulatory status as the DNEL values.

Van Broekhuizen proposed a nano-reference value (NRV) with a provisional status, not a regulatory one (van Broekhuizen et al., 2012).

Two studies by NIOSH used a recommended exposure limit (REL) which is a term used by NIOSH to describe an OEL, as the Occupational Health and Safety Administration (OSHA) uses permissible exposure limits (PELs) which are mandatory according to regulation (National Institute for Occupational Safety and Health, 2011, 2013).

The British Standards Institution (BSI) used the term benchmark exposure level (BEL), which indicates fairly well that this is not a

health-based recommendation but a tool to help in assessing the need for control measures (British Standards Institution, 2007).

One study used a no effect concentration in air, which is an unusual term that was directly based on the findings of an animal exposure study carried out by the same research group (Luizi, 2009).

One study used particle control values (PCVs), which they defined as a concentration that exceeds three times the local background particle concentration in the air. For this concentration value, emission or exposure controls may need to be implemented or modified, or further assessment of the controls be undertaken (McGarry et al., 2013).

Two Polish studies used a maximum admissible concentration-time weighted average (MAC-TWA) (Swidwinska-Gajewska & Czerczak, 2014, 2015), which is defined as the time-weighted average concentration for a conventional 8-h workday and a work week, to which workers may be exposed during their whole working life, without any adverse effects on their health.

One study used benchmark occupational exposure level (Guidotti, 2010). This proposal was derived using an environmental approach, and the author suggested this term so that it should not be confused with an OEL.

Kuempel et al. used a benchmark dose approach and determined and extrapolated the values belonging to the lower limit of the 95% confidence interval of the dose that caused a 0.1% excess risk of lung cancer in rats (BMDL) (Kuempel et al., 2006). They did not use the term OEL and discussed the derived OEL only as a human equivalent exposure estimate.

Exposure metrics used

The majority of OELs are only expressed as mass concentration ($\mu\text{g}/\text{m}^3$). There are, however, some exceptions. There is one proposal expressed in particle concentration (either fibers/ cm^3 or particle/ml) for each of the following: MNM (McGarry et al., 2013), fibers (British Standards Institution, 2007; van Broekhuizen et al., 2012; Stockmann-Juvala et al., 2014), GBP for metals and metal oxides (van Broekhuizen et al., 2012), and nanocellulose (Stockmann-Juvala et al., 2014). And there are two proposals expressed in particle- and mass concentrations for GBP insoluble nanomaterials (British Standards Institution, 2007).

Four out of 56 OEL-proposals contain a value both for mass and particle number concentration (British Standards Institution, 2007; German Hazardous Substances Committee, A. & BAuA, 2013). Only one study had proposals for mass-, particle- and surface concentration (nm^2/cm^3) for nanosilver (Stone, 2009).

For readability and clarity, we transformed all inhalation mass concentration values that were expressed as mg/m^3 into $\mu\text{g}/\text{m}^3$.

Proposed OELs reported in studies

See Table 2 for OEL values reported in the included studies.

OELs with a general approach

McGarry proposes a particle concentration of three times the local background particle concentration (LBPC) level that indicates particle emission from the process at hand. This would also take into account 'natural' variation of the background level. The authors propose that control measures may need to be implemented if this level is exceeded for more than a total of 30 minutes during a workday, and/or if a single short-term measurement exceeds five times the LBPC. Guidotti proposes as the benchmark occupational exposure level value to simply use the value of $30\mu\text{g}/\text{m}^3$ that is set for particulate matter (PM_{10}) in ambient air and as agreed upon

Table 2. Proposed occupational exposure limits for manufactured nanomaterials.

Category	Study Reference	Nanomaterials and specifications	OEL Name	Mass concentration $\mu\text{g}/\text{m}^3$	Particle concentration (particle/ml, fibers/ cm^3)	Surface concentration (nm^2/cm^3)	Approach
Inhalation: general approach							
MNM	Guidotti (2010)	Fine particulate matter $\leq 2500 \text{ nm}$	BOEL	30			Environmental
MNM	McGarry et al. (2013)	Airborne particles from nanotechnology processes	PCVs		3 times LBPC for over 30 minutes		Environmental
Inhalation: categorical approach							
CMAR Fibers	BSI (2007) AGS (2013)	CMAR nanomaterials Non-entangled fibrous NM	BEL Acceptance level (default), respirable fraction	$0.1 \times \text{bulk WEL}$	0.01		Bridging Bridging/Grouping
Fibers	BSI (2007)	Fibrous nanomaterials	BEL		0.01		Bridging/Grouping
Fibers	Stockmann-Juvala et al. (2014)	Carbon nanofibers, CNFs	OEL		0.01		Bridging/Grouping
Fibers	van Broekhuizen et al. (2012)	Carbon nanotubes, CNTs, NRV insoluble NM with high aspect ratio $>3:1$	NRV		0.01		Bridging/Grouping
GBP	AGS (2013)	In operations with NM: nanosized GBP with no specific toxicity	OEL respirable fraction, default	500			Grouping
GBP	AGS (2013)	No specific operations with NM: granular bio-persistent particles	OEL respirable fraction	1250			Grouping
GBP	BSI (2007)	Insoluble nanomaterials	BEL	$0.066 \times \text{bulk WEL}$	20 000		Bridging
GBP	Pauluhn (2011)	Inhaled poorly soluble particles	DNEL	$0.5 \mu\text{L PM}_{\text{respirable}}/\text{m}^3 \times \text{agglomerate density}$			Categorical QRA/Grouping
GBP	van Broekhuizen et al. (2012)	Metals and metal oxides, biopersistent granular NM $>6000 \text{ kg}/\text{m}^3$	NRV		20 000		Grouping
GBP	van Broekhuizen et al. (2012)	Metals and metal oxides, biopersistent granular NM $<6000 \text{ kg}/\text{m}^3$	NRV		40 000		Grouping
Low-toxicity dust	Stockmann-Juvala et al. (2014)		OEL	$300 \text{ (respirable fraction)}, 4000 \text{ (inhalable fraction)}$ applicable OEL, WEL			Bridging/Grouping
Non bio-persistent	van Broekhuizen et al. (2012)	Non-biopersistent granular NM 1–100 nm	NRV				Bridging
Soluble Carbon	BSI (2007) Inhalation: specific approach Aschberger et al. (2011)	Soluble nanomaterials	BEL	$0.5 \times \text{bulk WEL}$			Bridging
Carbon	Aschberger et al. (2011)	Multi-walled carbon nanotubes, MWCNT 10 nm		1			QRA
Carbon	Aschberger et al. (2011)	Multi-walled carbon nanotubes, MWCNT 140 nm		2			QRA
Carbon	Luizi (2009)	Carbon nanotubes, CNTs	No effect concentration in air	2.5			QRA
Carbon	Nakanishi et al. (2015)	Carbon nanotube group, SWCNT, DWCNT, MWCNT	OEL 15 years	30			QRA
Carbon	NIOSH (2013)	All carbon nanotubes and nanofibers	REL respirable elemental carbon	<1			QRA

(continued)

Table 2. Continued

Category	Study Reference	Nanomaterials and specifications	OEL Name	Mass concentration $\mu\text{g}/\text{m}^3$	Particle concentration (particle/ml, fibers/ cm^3)	Surface concentration (nm^2/cm^3)	Approach
Carbon	Pauluhn (2010)	Multi-walled carbon nanotubes, MWCNT	inhalable fraction	50			QRA
Carbon	Stone (2009)	MWCNT					
Carbon	Kuempel et al. (2006)	Carbon black, CB ultrafine	DNEL chronic inhalation, systemic immune effect	0.67			QRA
Carbon	Kuempel et al. (2006)	Carbon black, CB ultrafine	BMDL 45 years (lung dosimetry, model 1)	120			QRA
Carbon	Aschberger et al. (2011)	Fullerenes, C_{60}	Carbon black, CB ultrafine	240			QRA
Carbon	Shinohara et al. (2011)	Fullerenes, C_{60}	INEL	7.4			QRA
Nanocellulose	Stockmann-Juvala et al. (2014)	Fullerenes, C_{60}	OEL (PL) 15 years	390			QRA
Nanoclays	Stockmann-Juvala et al. (2014)	Nanocellulose	OEL		0.01		Bridging
		Nanoclays	OEL	300 (respirable fraction), 4000 (inhalable fraction)			Bridging/Grouping
Nanosilver	Aschberger et al. (2011)	Nano Ag	INEL lung function	0.33			QRA
Nanosilver	Aschberger et al. (2011)	Nano Ag	INEL lung other effects	0.67			QRA
Nanosilver	Stone (2009)	Nano Ag	DNEL lung exposure, extrapolating factor 10	0.098	1200	2.2×10^6	QRA
Nanosilver	Stone (2009)	Nano Ag	DNEL lung exposure, extrapolating factor 3	0.33	4000	7.2×10^6	QRA
Nanosilver	Stone (2009)	Nano Ag	DNEL liver effect	0.67	7000	1.2×10^7	QRA
Nanosilver	Swidwiska-Gajewska & Czerzszak (2015)	Nano Ag	MAC-TWA inhalable fraction	10			QRA
Silicon dioxide	Stockmann-Juvala et al. (2014)	Amorphous silica, SiO_2	OEL respirable fraction	300			QRA
Titanium dioxide	Aschberger et al. (2011)	TiO_2	INEL	17			QRA
Titanium dioxide	Kuempel et al. (2006)	TiO_2 ultrafine	BMDL 45 years (lung dosimetry, model 1)	73			QRA
Titanium dioxide	Kuempel et al. (2006)	TiO_2 ultrafine	BMDL 45 years (lung dosimetry, model 2)	140			QRA
Titanium dioxide	NIOSH (2011)	TiO_2 ultrafine	REL (up to 10 h/day, 40 h/week)	300			QRA
Titanium dioxide	Ogura et al. (2011)	TiO_2	OEL (PL) 15 years	610			QRA
Titanium dioxide	Stockmann-Juvala et al. (2014)	TiO_2	OEL respirable fraction	100			QRA
Titanium dioxide	Swidwiska-Gajewska & Czerzszak (2014)	TiO_2	MAC	300			QRA
Titanium dioxide	Warheit (2013)	High surface reactivity anatase-rutile nanoscale TiO_2	OEL	1000			QRA
Titanium dioxide	Warheit (2013)	Low surface reactivity nanoscale TiO_2	OEL	2000			QRA
Dermal Carbon	Stone (2009)	MWCNT	DNEL dermal chronic exposure, assessment factor 3	0.414 mg/person bodyweight			QRA
Carbon	Stone (2009)	MWCNT	DNEL dermal chronic exposure	1.241 mg/person bodyweight			QRA
Oral Carbon	Stone (2009)	Fullerite, mixture of C_{60} + C_{70}	DNEL oral acute exposure	40 mg/kg body weight			QRA
Carbon	Stone (2009)	Water soluble C_{60} , polyalkylsulfonated	DNEL oral chronic exposure	0.17 mg/kg body weight			QRA
Acute MNM	McGarry et al. (2013)	Airborne particles from nanotechnology processes	PCVs, single short-term measurement		5 times the local particle reference value		Environmental
Carbon	Stone (2009)	MWCNT	DNEL acute inhalation, systemic immune effect	4.02			QRA
Carbon	Aschberger et al. (2010)	Fullerenes, C_{60}	INEL short term, inhalable fraction	44.4			QRA
Carbon	Stone (2009)	MWCNT	DNEL acute inhalation, pulmonary effect	201			QRA

(continued)

Table 2. Continued

Category	Study Reference	Nanomaterials and specifications	OEL Name	Mass concentration $\mu\text{g}/\text{m}^3$	Particle concentration (particle/ml, fibers/ cm^3)	Surface concentration (nm^2/cm^2)	Approach
Carbon	Stone (2009)	MWCNT	DNEL dermal acute exposure	7448 $\mu\text{g}/\text{person bodyweight}$		QRA	
Carbon	Stone (2009)	MWCNT	DNEL dermal acute exposure, assessment factor 3	2483 $\mu\text{g}/\text{person bodyweight}$		QRA	

AGS = German Hazardous Substances Committee.

BEL = benchmark exposure level.

BMDL = benchmark dose lower (95% confidence limit of the benchmark dose, BMD).

BOEL = benchmark occupational exposure level.

BSI = British Standards Institution.

CMAR = carcinogenic, mutagenic, asthmagenic or a reproductive toxin.

CNT = carbon nanotube.

DNEL = derived no-effect level.

DWCNT = double-walled carbon nanotube.

GBP = granular biopersistent particles.

INEL = indicative no effect level.

LBPC = local background particle concentration.

MAC = maximum admissible concentration.

MAC-TWA = maximum admissible concentration time weighted average.

MNM = manufactured nanomaterial.

MWCNT = multi-walled carbon nanotube.

NIOSH = National Institute for Occupational Safety and Health (United States).

NM = nanomaterial.

NRV = nano reference value.

OEL = occupational exposure limit.

OEL (PL) = occupational exposure limit period-limited.

PCVs = particle control values.

REL = recommended exposure limit.

QRA = traditional quantitative risk assessment.

SWCNT = single-walled carbon nanotube.

TWA = time weighted average exposure over the 8-hour working day.

WEL = workplace exposure limit.

for the general population in Canada (Guidotti, 2010). He argued that there are many similarities between PM_x and MNMs and that if these values are deemed fit to protect the general population, this probably also protects workers.

OELs for fibers

For fibers, all four included proposals used the same value, a level ten times lower than the asbestos OEL of 0.1 fibers/ml, because of the use of a safety margin of a factor 10. This particular value was chosen because of the assumed physico-chemical similarities with asbestos (British Standards Institution, 2007; German Hazardous Substances Committee, A. & BAuA, 2013; Stockmann-Juvala et al., 2014; van Broekhuizen et al., 2012). Moreover, Stockmann-Juvala mentioned that this limit is "based on the precautionary principle" (Stockmann-Juvala et al., 2014). Similarly based on what is tolerated for asbestos exposure, the German authority considers a level that is ten times lower an acceptable level (Bundesanstalt für Arbeitsschutz und Arbeitsmedizin, 2008; German Hazardous Substances Committee, A. & BAuA, 2013).

OELs for granular biopersistent nanoparticles (GBP)

For GBPs, van Broekhuizen proposes two groups based on the classification recommended by the German Institute of Occupational Safety and Health (IFA): one category with a density higher than 6000 kg/m³ and the second one having a lower density than this value (van Broekhuizen et al., 2012). The starting point for the number-based reference values is the calculation of the number of nanoparticles with a diameter of 100 nm that constitute a mass of 0.1 mg/m³. Values are defined as corrected for the background concentrations.

OELs for non-biopersistent nanoparticles

Van Broekhuizen proposes the same OEL as for the bulk material in the case that the chemical is soluble or not biopersistent (van Broekhuizen et al., 2012).

OELs for specific MNMs

Carbonaceous material. For carbon nanotubes and nanofibers, the proposed OELs differ considerably. The lowest proposed value is 0.67 µg/m³ (Stone, 2009), which is smaller than 1 µg/m³ recommended by NIOSH 2013 (National Institute for Occupational Safety and Health, 2013). Nakanishi proposes a value that is at least 30 times larger but that would protect only for 15 years (Nakanishi et al., 2015), while the NIOSH value is calculated based on 45-year working lifetime.

Also for fullerenes, the values differ by a factor of 50 with the same difference that the highest value protects only for 15 years (Shinohara et al., 2011). The value proposed by Aschberger et al. 2011 is significantly lower (Aschberger et al., 2011).

With 120 and 240 µg/m³, the OEL values for carbon black (Kuempel et al., 2006) are much higher than for carbon nanotubes for which the highest value is 50 µg/m³ (Pauluhn, 2010).

Metals and metaloxides. As for nanosilver, the differences are considerable with 0.098 µg/m³ based on a large extrapolation factor and effects on the lungs, and 0.67 µg/m³ based on more systemic effects (Aschberger et al., 2010, 2011; Stone, 2009). However, the Polish group proposed 100 to 15 times higher value of 10 µg/m³ which was already considerably lower than the current value of 50 µg/m³ (Swidwinska-Gajewska & Czerczak, 2015). The authors provide no clear justification for such high values.

Also for titanium dioxide there is considerable variation. Aschberger et al., 2010 proposed 17 µg/m³, which is the lowest

compared to the other groups. The highest limit was proposed by Warheit (2013) with 5000 µg/m³ which is almost 300-fold higher (Warheit, 2013). Three studies proposed the same value of 300 µg/m³ (National Institute for Occupational Safety and Health, 2011; Stockmann-Juvala et al., 2014; Swidwinska-Gajewska & Czerczak, 2014) where we assumed that two of those values were simply taken over from NIOSH, but this was not clearly stated in the papers.

For acute exposure to nanocarbon, only Aschberger and Stone derived values. For inhalation of fullerenes C₆₀ the limits were identical, 44.4 µg/m³ (Aschberger et al., 2011; Stone, 2009). For acute dermal exposure Stone set two limits: 0.414 mg/person bodyweight and 1.241 mg/person bodyweight based on different assumptions in the derivation (Stone, 2009).

Other materials. For other MNMs there are only single values available that are not proposed by other groups such as for low-toxicity dust, nanoclays, nanocellulose (Stockmann-Juvala et al., 2014), CMAR and soluble nanomaterials (British Standards Institution, 2007), and non-biopersistent nanomaterials (van Broekhuizen et al., 2012).

Discussion

In total, we found 56 proposals for OELs for MNMs in 20 papers. Of these two proposed a level for all MNMs, 14 proposed OELs for a category of MNMs and 40 proposed OELs for a specific material. For fibers, four studies proposed a similar value but for CNTs the values differed with a factor of 30 to 50 and for metals with a factor of 100 to 300. We could not explain these differences.

When we compare the exposure levels that have been reported in workplace exposure studies to the OEL values that we have reported here, it seems that there is ample room for a reduction of exposure in workplaces to comply with the proposed OELs. Debia et al. reported occupational exposure to carbon nanofibers (CNFs) in potential exposure situations with values ranging from not detected to 193 fibers/cm³ for studies that measured particle number concentrations and from not detected to 1000 µg/m³ for studies that measured mass concentration (Debia et al., 2016). Most of these values exceed the proposed OELs discussed in this paper. For CNTs, there were only two exposure situations that exceeded the highest proposed OEL of 50 µg/m³ but the lowest OEL of 0.67 µg/m³ was exceeded in almost all situations that reported mass concentrations. For TiO₂ on the other hand, all but one exposure situation was below the NIOSH recommended value of 300 µg/m³. For nanosilver, only two out of ten exposure situations were below the proposed OEL of 0.33 µg/m³ based on inhalation exposure. Because these were workplaces that admitted researchers to take measurements, it is conceivable that in many other workplaces exposures will be higher. Applying and using the OELs presented here will be a helpful indication that control measures should be taken.

The strength of our study is that we performed a systematic search to identify developed OELs, assessed them, and listed them in a systematic way. We did not exclude studies based on language or on publication status. We believe that we have compiled a comprehensive list of all available values. The proposed values can be used as reference or benchmark values for comparison with workplace measurement and for risk management.

One of the limitations of our review is that we could not compare in detail the different methodologies used to derive the OELs. However, for those that used quantitative risk assessment, differences in the proposed levels can be explained by the animal studies used, how no-observed-adverse-effect level (NOAEL) were

identified in the studies and which adjustment factors were used to extrapolate the results to a human exposure during an entire working life. Another limitation is that the studies did not always report sufficient information on the type of MNM studied. For example, TiO₂ may have different crystal structures with different toxicological properties. From the studies it was not always clear which form of TiO₂ was considered.

Some studies indicated that the values were meant as a time-weighted average (TWA) over an 8-h working day and a 40-h workweek. Some advised as well how to calculate a value for a TWA-15 min for short-term exposure, also referred to as STEL (Short-Term Exposure Limit). By definition all OELs are 8 h-TWA, unless otherwise stated. There is a similar situation with the interpretation of the OELs where some authors mentioned that the proposed values should prevent adverse effects over the time span of a 45-year working life but others only proposed this for a period of 15 years. Another limitation is that OELs assume that the MNMs are measured as primary nanoparticles. However, workplace exposure studies indicate that most MNMs are present in micro-sized agglomerates, which may also be the case for the rodent studies (Debia et al., 2016). It is unclear how this would be taken into account.

Progress in the nanotechnology field is continuously growing. In his 2006 article, Maynard presented five challenges regarding nanotechnology research that would span over the following two decades. Among the challenges the author proposed for the next decade the development of “instruments to assess exposure to manufactured nanomaterials” including at the workplace (Maynard et al., 2006). Our review is timely in this fashion, but still more research is needed regarding OELs.

Implications for practice

The OELs listed here can be used as reference or benchmark values for comparison with workplace exposure towards a better understanding of the need for control measures. For some MNM categories such as fibers, one concrete OEL was proposed by four different studies (British Standards Institution, 2007; German Hazardous Substances Committee, A. & BAuA, 2013; Stockmann-Juvala et al., 2014; van Broekhuizen et al., 2012). For other categories or specific MNMs there is a range of values proposed making it difficult to recommend one value over another. However, given current workplace exposure reports and when using the highest OEL values, this should be an incentive to lower exposures in the workplace.

Implications for research

There is a need to develop a coordinated approach among researchers and relevant stakeholders towards the harmonization of OEL derivation for nanomaterials. This will improve transparency and communication towards stakeholders. Communication will also be improved with a common terminology used by all the parties involved from academia to professionals and workers. Moreover, the recent and emerging need for nanomaterial exposure limits provides a unique opportunity for organizations worldwide to finally find consensus about the naming of OELs.

Currently, there is variation in the selection and analysis of animal studies used to underpin quantitative risk assessment. Using systematic reviews of animal studies, including systematic risk of bias assessment (Hooijmans & Ritskes-Hoitinga, 2013; Hooijmans et al., 2014), would lead to more uniform conclusions. Finally, agreement about interspecies and intraspecies adjustment factors

would be needed to come to more similar conclusions and exposure values.

Regular updating of this list will be necessary to keep up with scientific progress in both the field of (nano) particle toxicological research and in the field of OEL derivation.

Declaration of interest

The findings and conclusions in this report are those of the authors and do not necessarily represent the views of host/employing organizations.

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