



## Surface chemistry of a pine-oil cleaner and other terpene mixtures with ozone on vinyl flooring tiles

Jason E. Ham <sup>\*</sup>, J. Raymond Wells

Exposure Assessment Branch, Health Effects Laboratory, National Institute for Occupational Safety and Health, 1095 Willowdale Road, Morgantown, WV 26505, United States

### ARTICLE INFO

#### Article history:

Received 20 August 2010  
Received in revised form 1 November 2010  
Accepted 8 December 2010  
Available online 14 January 2011

#### Keywords:

Pine oil  
Ozone  
Reaction products  
Surface chemistry

### ABSTRACT

Indoor environments are dynamic reactors where consumer products (such as cleaning agents, deodorants, and air fresheners) emit volatile organic compounds (VOCs) that can subsequently interact with indoor oxidants such as ozone ( $O_3$ ), hydroxyl radicals, and nitrate radicals. Typically, consumer products consist of mixtures of VOCs and semi-VOCs which can react in the gas-phase or on surfaces with these oxidants to generate a variety of oxygenated products. In this study, the reaction of a pine-oil cleaner (POC) with  $O_3$  (100 ppb) on a urethane-coated vinyl flooring tile was investigated at 5% and 50% relative humidity. These results were compared to previous  $\alpha$ -terpineol +  $O_3$  reactions on glass and vinyl surfaces. Additionally, other terpene and terpene alcohol mixtures were formulated to understand the emission profiles as seen in the POC data. Results showed that the  $\alpha$ -terpineol +  $O_3$  reaction products were the prominent species that were also observed in the POC/ $O_3$  surface experiments. Furthermore,  $\alpha$ -terpineol +  $O_3$  reactions generate the largest fraction of oxygenated products even in equal mixtures of other terpene alcohols. This finding suggests that the judicious choice of terpene alcohols for inclusion in product formulations may be useful in reducing oxidation product emissions.

Published by Elsevier Ltd.

### 1. Introduction

The indoor office environment atmosphere is a mixture of hundreds of volatile organic compounds such as alcohols, alkanes, alkenes, terpenes, and phthalates that all react differently with  $OH^\cdot$ ,  $O_3$  and  $NO_3^\cdot$  (Nazaroff and Weschler, 2004). These three reactants have been identified as the key drivers for the complex chemistry that can occur in the indoor environment (Nazaroff and Cass, 1986; Weschler et al., 1992, 1994; Weschler and Shields, 1996; Weschler, 2001; Sarwar et al., 2002; Nojgaard, 2010). Most of the pertinent research has been conducted on individual volatile organic chemicals (VOCs). Therefore, understanding the fate of this complex mixture of indoor office environments has been limited to theoretically combining the chemical research results of individual VOCs.

Recent studies to understand the fate of mixtures in simulated indoor environments have focused on consumer cleaning product formulations. For example, in an effort to understand glycol ether and terpenoid emissions over time for several common cleaning products and air fresheners, chamber experiments were conducted to characterize the emissions from simulated indoor activities (Singer et al., 2006). Five cleaning products either full strength or diluted were applied to surfaces or floors while VOC emissions

were collected and analyzed using sorbent tubes and thermal desorption gas chromatography/mass spectrometry. Emissions of limonene, 2-butoxyethanol, dihydromyrcenol,  $\alpha$ -terpineol, benzyl acetate and many other VOCs were detected. Limonene for instance had a 0–1 h chamber concentration range after product use of approximately 173–1120 ppb while 2-butoxyethanol's concentration range was 56–477 ppb. 2-Butoxyethanol has an established reference exposure level (REL) by the California Air Resources Board of 2.9 ppm (OEHHA, 1999). This recent work determined realistic chemical composition and concentration for common cleaning products and highlighted possible practices to reduce VOC emissions.

An examination of the primary and secondary products as a result of ozone reacting with a complex VOC mixture has been conducted (Fan et al., 2003). A 23 component VOC mixture, a 21 component VOC mixture (limonene and  $\alpha$ -pinene removed from mixture), and a limonene and  $\alpha$ -pinene only mixture were all reacted in separate experiments with 40 ppb  $O_3$ . Addition of  $O_3$  to the 21 VOC mixture yielded either none or reduced concentrations of the observed products compared to addition of  $O_3$  to the 23 VOC mixture. An additional experiment adding  $O_3$  to the limonene and  $\alpha$ -pinene mixture resulted in similar yields of reaction products as observed in the 23 VOC mixture. Observed products included particulate matter, formaldehyde, glyoxal, methyl glyoxal, acetaldehyde, propionaldehyde, and butyraldehyde. This series of experiments are important for demonstrating the influence of

<sup>\*</sup> Corresponding author. Tel.: +1 304 285 6214; fax: +1 304 285 6041.

E-mail address: [bvo2@cdc.gov](mailto:bvo2@cdc.gov) (J.E. Ham).

terpene/ozone chemistry on gas-phase oxygenated organic reaction product concentrations in indoor environments.

The gas-phase reactions of VOCs and semi-VOCs with the indoor oxidant ozone have been studied extensively (Atkinson et al., 1995; Atkinson, 1997; Aschmann et al., 2002, 2003; Ljubic and Sabljic, 2002, 2005; Atkinson and Arey, 2003; Baker et al., 2004); however, equivalent reactions with these compounds on surfaces are limited (Moise and Rudich, 2002; Voges et al., 2007; Ham and Wells, 2008, 2009; Pandrangi and Morrison, 2008; Wells et al., 2008; Gunschera et al., 2009). Several limitations of characterizing surface-O<sub>3</sub> reactions come from (i) accurately collecting emission for reaction product yields, (ii) consistent delivery of O<sub>3</sub> to a surface and (iii) controlling the concentration of a volatile species on a surface in order to determine reaction probabilities. Recently, the O<sub>3</sub> surface-phase reaction of low vapor pressure squalene showed a twenty fold enhancement of reaction rate when compared to the calculated gas-phase reaction (Fruekilde et al., 1998; Wisthaler et al., 2005; Tamas et al., 2006; Coleman et al., 2008; Wells et al., 2008). Voges et al. (2007) also performed surface experiments with the atmospheric terpene, menthenol, on a glass plate with O<sub>3</sub>. However, in these experiments, menthenol was oriented so that the OH group was bound to the glass surface by silane-functionalization chemistry. Results showed that the initial reaction probability for ozone interacting with the cyclohexene moiety of the surface-bound terpene increased 5–20 times over the corresponding gas-phase reaction. This observation of higher reaction probabilities indicates improved accessibility of O<sub>3</sub> to the unsaturated carbons of the surface-bound terpene. This result further suggests that if polar groups of VOCs and semi-VOCs in the gas phase reorient themselves as they are adsorbed onto surfaces, then reactions with indoor oxidants can be enhanced.

Several different types of pine-oil based cleaners have been consistently used in hard surface cleaning of flooring and countertops. These cleaners contain a number of terpenoid compounds including:  $\alpha$ -terpineol, limonene, terpinolene, and  $\alpha$ -pinene. The gas-phase reactions of these compounds (among others not listed) with ozone have been studied (Nunes et al., 2005; Wells, 2005; Jonsson et al., 2006, 2007; Nojgaard et al., 2006; Cain et al., 2007; Sarwar and Corsi, 2007; Vibenholt et al., 2009). Recently, the surface-phase reactions of  $\alpha$ -terpineol and dihydromyrcenol (lavender/citrus fragrance) with ozone on glass and vinyl flooring tiles have been investigated (Ham and Wells, 2008, 2009). In both of these experiments, new oxygenated organics were observed compared to gas-phase experiments indicating that the surface can modify the mechanism of formation of reaction products as suggested by Wells et al. (2008) and Voges et al. (2007). The goals of the experiments (described below) are to identify the products from the reaction of O<sub>3</sub> and a pine-oil based cleaner (POC) and other terpene/terpene alcohol mixtures on urethane-coated vinyl flooring tiles and to qualitatively assess these against single component terpene + O<sub>3</sub> surface-phase reactions.

## 2. Experimental methods

### 2.1. Materials

(1R)-(+)- $\alpha$ -pinene (98%),  $\alpha$ -terpineol (90% Tech grade), (R)-(+)- $\beta$ -citronellol (98%), dihydromyrcenol (99%), geraniol ( $\geq 99\%$ ), limonene ( $\geq 99\%$ ), and *O*-(2,3,4,5,6-pentafluoro-benzyl) hydroxylamine hydrochloride (PFBHA) (98+%), were purchased from Sigma Aldrich (St. Louis, MO). A general purpose pine-oil based cleaner packaged as a concentrated liquid was purchased at a local market. Methanol (high-performance liquid chromatography grade) was purchased from Fisher Scientific (Pittsburgh, PA). Water (DI H<sub>2</sub>O) was distilled, deionized to a resistivity of 18 M $\Omega$  cm and filtered using a

Milli-Q<sup>®</sup> filter system (Billerica, MA). Ultra-high purity nitrogen (UHP N<sub>2</sub>) was purchased from Butler Gas (McKees Rocks, PA). Vinyl floor tiles (composition: 80% limestone filler, 12% (vinyl resin – min 60% of binder), 4% plasticizers, 4% stabilizers and processing aids, with urethane finish) that are currently used in the National Institute for Occupational Safety and Health (NIOSH) facility (Morgantown, WV) were purchased from a local building supply center.

### 2.2. Surface application of prepared mixtures

Two questions have emerged regarding terpene mixture/ozone/surface reactions: (i) do multiple terpene alcohols compete for surface sites and affect gas-phase emissions? and (ii) does the presence of  $\alpha$ -terpineol in the mixture dominate emissions due to physical characteristics and chemical structure? To answer these questions, a series of terpene/terpene alcohol + water solutions were formulated (see Table 1):  $\alpha$ -terpineol/ $\alpha$ -pinene/limonene (TAL),  $\alpha$ -terpineol/ $\alpha$ -pinene/limonene/dihydromyrcenol (TALD),  $\alpha$ -terpineol/geraniol/citronellol/dihydro-myrcenol (TGCD), and  $\alpha$ -pinene/limonene/geraniol/dihydromyrcenol (ALGCD). These mixtures were applied to the urethane-coated vinyl flooring surface (as described below) with 100 ppb O<sub>3</sub> in separate experiments and the resulting chromatograms were compared to those in the POC + O<sub>3</sub> experiments.

Mixtures of terpenes/terpene alcohols were prepared according to Table 1. Three  $\mu$ L of each compound were added to 1 mL of deionized water. The POC solution was made according to the directions on the bottle (1/4 cup POC to a gallon of water), but scaled down for 2 mL solutions. All solutions were made in 4 mL amber glass vials and then 0.44 mL (0.44 g on average) was sprayed using a previously described apparatus (Flemmer et al., 2007) on a vinyl floor tile (urethane finish side up) for each of the experiments below. All experiments were conducted at room temperature, 22  $\pm$  2 °C. An aluminum plate (30.5 cm  $\times$  30.5 cm  $\times$  0.64 cm), with a 14.9 cm diameter hole in the center, was used as a template for spraying the surface. This ensured that the area sampled was directly under the Field and Laboratory Emission Cell (FLEC, Markes International, UK) and that no detectable emissions were collected due to slowly evaporating solvent under the FLEC o-ring.

Before application of any of the described mixtures, the urethane-coated vinyl flooring tiles were aerated to reduce volatile gaseous emissions. New vinyl flooring tiles were placed in a 10 L Teflon chamber filled with UHP N<sub>2</sub> and left overnight. The Teflon chamber was then emptied and allowed to refill and vent using UHP N<sub>2</sub> at 600 mL min<sup>-1</sup> for approximately 6 h. After 6 h, the chamber was filled again with UHP N<sub>2</sub> and left overnight for experiments the next day. Before beginning an experiment, the chamber was again filled and vented at 600 mL min<sup>-1</sup> for 45 min. Several background experiments with only air flowing through the FLEC showed no detectable emission products after ventilation, compared to new “non-ventilated” tiles where air only was present (Ham and Wells, 2008). Immediately following the ventilation step, the tile was sprayed and placed under the FLEC.

### 2.3. FLEC automation and control system (FACS)

The FACS has been described previously (Flemmer et al., 2007). Briefly, the FACS is composed of three stages; the air purification stage, the air humidification stage, and the reactant injection and delivery stage. The air purification stage regulates, cleans, and dries the “in-house” compressed air via two drierite columns and a carbon filter for delivery to the rest of the system. The air humidification stage mixes both dry and humid air to create and deliver the desired humidity for the duration of the experiment. The reactant ejection and delivery stage introduces the reactant, at the desired

**Table 1**

Chemical components of mixtures that were formulated then applied to vinyl flooring tiles in surface-O<sub>3</sub> experiments.

Chemical	Mixture 1 (TAL)	Mixture 2 (TALD)	Mixture 3 (TGCD)	Mixture 4 (ALGCD)
$\alpha$ -terpineol	X	X	X	
$\alpha$ -pinene	X	X		X
Limonene	X	X		X
Dihydromyrcenol		X	X	X
Geraniol			X	X
Citronello			X	X

X indicates that 3.0  $\mu$ L of "pure" chemical was injected to 1 mL of DI H<sub>2</sub>O in a 4 mL amber vial.

concentration, into the air stream and delivers the air to the FLEC for the experiment.

#### 2.4. Mixture solutions + O<sub>3</sub> on a vinyl tile at 50% or 5% RH (FACS)

For all mixtures (Table 1) and the POC solution, the FACS was initialized using the following parameters: the reactant concentration (i.e., O<sub>3</sub>) was set to 100 ppb, the relative humidity was set to 50% or 5% (depending on experiment), and the flow rate through FLEC was set to 300 mL min<sup>-1</sup>. High concentration ozone (45–50 ppm) was prepared in advance by placing a UV lamp (Jelight ozone generator, Model 600, Irvine, CA) in a 70 L Teflon chamber filled with clean dry air. The ozone/air mix was then regulated to 100 ppb for the air flow through the FLEC.

#### 2.5. Vinyl tile – Rinse experiments

A series of rinse experiments were conducted by placing a vinyl flooring tile (coated/uncoated with POC or Mixture solutions) under the FLEC and then exposing with O<sub>3</sub> (100 ppb in air) or air for 12 h at 5% or 50% RH. After exposure the vinyl tiles were rinsed with 6 mL of HPLC grade methanol into a custom Pyrex container (Ham and Wells, 2008, 2009). The rinsate was then collected and

split into two vials. The first vial contained 2.0 mL of sample, while the second vial contained the sample remaining in the Pyrex container. To the first vial, 200  $\mu$ L of PFBHA (20 mM in ACN) was added then left overnight to complete carbonyl derivatization. PFBHA derivatizes products with carbonyl substructures to facilitate detection by gas chromatography/mass spectrometry (Yu et al., 1998).

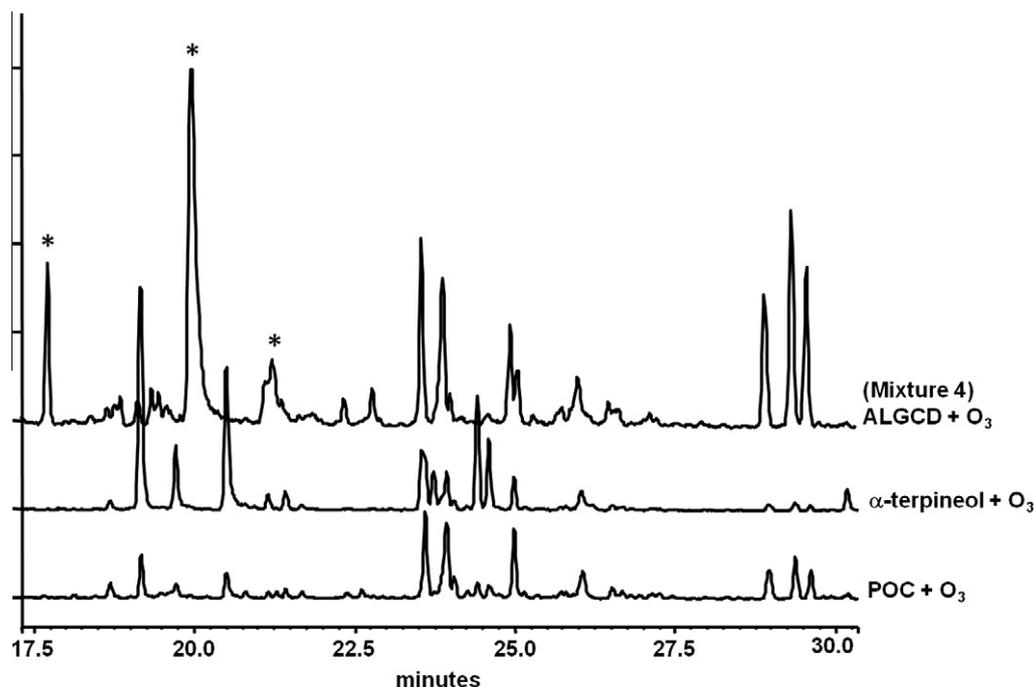
#### 2.6. Sample collection and analysis

Emissions from POC or Mixture reactions (FACS) were collected in a 100 mL glass impinger immersed in a thermostated water bath at 0  $\pm$  0.5 °C controlled by an immersion chiller (EK45, Fisher Scientific, Pittsburgh, PA) and connected to the output of the FLEC using 0.64 cm O.D. Teflon tubing. To the impinger, 14.7 mL of HPLC grade methanol was added. Output air from the FLEC at 300 mL min<sup>-1</sup> was bubbled through the impinger for 12 h.

At the end of the 12 h sampling period, the solution left in the impinger (~4 mL) was divided into two 4 mL amber vials. The first vial (labeled 1A) was filled with 2.0 mL of solution from impinger. The remaining solution (~1.5–2 mL) was placed in the second vial (vial 1B). To identify reaction products (i.e., aldehydes, ketones, and dicarbonyls), 200  $\mu$ L of PFBHA (20 mM in ACN) was added to the 1A vials. These vials were then left overnight to allow for the reaction to go to completion. No PFBHA was added to the 1B vials.

After reaction completion, the 1A vials were blown to complete dryness using UHP N<sub>2</sub> then reconstituted in 100  $\mu$ L of methanol. The 1B vials were unaltered and injected as pure samples. This was done to monitor the concentration of POC or mixture solutions over time.

All samples were then analyzed using a Varian (Palo Alto, CA) 3800/Saturn 2000 GC/MS system operated in the electron impact (EI) mode. Compound separation was achieved by a J&W Scientific (Folsom, CA) DB-5MS (0.25 mm I.D., 30-m long, 1  $\mu$ m film thickness) column and the following GC oven parameters: 60 °C for 1 min, then 20 °C min<sup>-1</sup> to 170 °C, then 3 °C min<sup>-1</sup> to 280 °C and held for 5 min. One  $\mu$ L of each sample (Vial 1A or 1B) was injected



\* Indicates peaks not observed in POC or  $\alpha$ -terpineol + O<sub>3</sub> on a vinyl flooring tile at 50% relative humidity

Fig. 1. Chromatograms obtained from POC + O<sub>3</sub>,  $\alpha$ -terpineol + O<sub>3</sub>, and ALGCD mixture + O<sub>3</sub> reactions on vinyl flooring tiles.

**Table 2**Reaction products observed due to  $\alpha$ -terpineol and POC reactions with O<sub>3</sub> at RH = 50% on vinyl flooring tiles.

Retention time (min)	Name * indicates proposed identity	Molecular weight	Structure	Cl ions (observed)
19.2	6-hydroxyhept-5-en-2-one	128		324
20.5	5-(1-hydroxy-1-methylethyl)-2-methylcyclohex-2-en-1-one	168		364
23.6	Ethanedial (glyoxal)	58		449
23.9	(5E)-3-(2-hydroxypropan-2-yl)hept-5-enal	170		364
24.4				
24.6				
25.0	2-oxopropanal (methylglyoxal)	72		463
28.9	4-oxopentanal	100		491
29.4				
29.6				

in the splitless mode, and the GC injector was returned to split mode 1 min after sample injection, with the following injector temperature parameters: 60 °C for 1 min then 180 °C min<sup>-1</sup> to 250 °C and held to the end of the chromatographic run. The Saturn 2000 ion trap mass spectrometer was tuned using perfluorotributylamine (FC-43). Full-scan EI ionization spectra were collected from *m/z* 40–650. Each sample was analyzed in duplicate. Acetonitrile was the chemical ionization reagent used for all CI spectra. When possible, commercially available samples of the identified products were derivatized and subsequently analyzed to verify matching ion spectra and chromatographic retention times.

### 3. Results and discussion

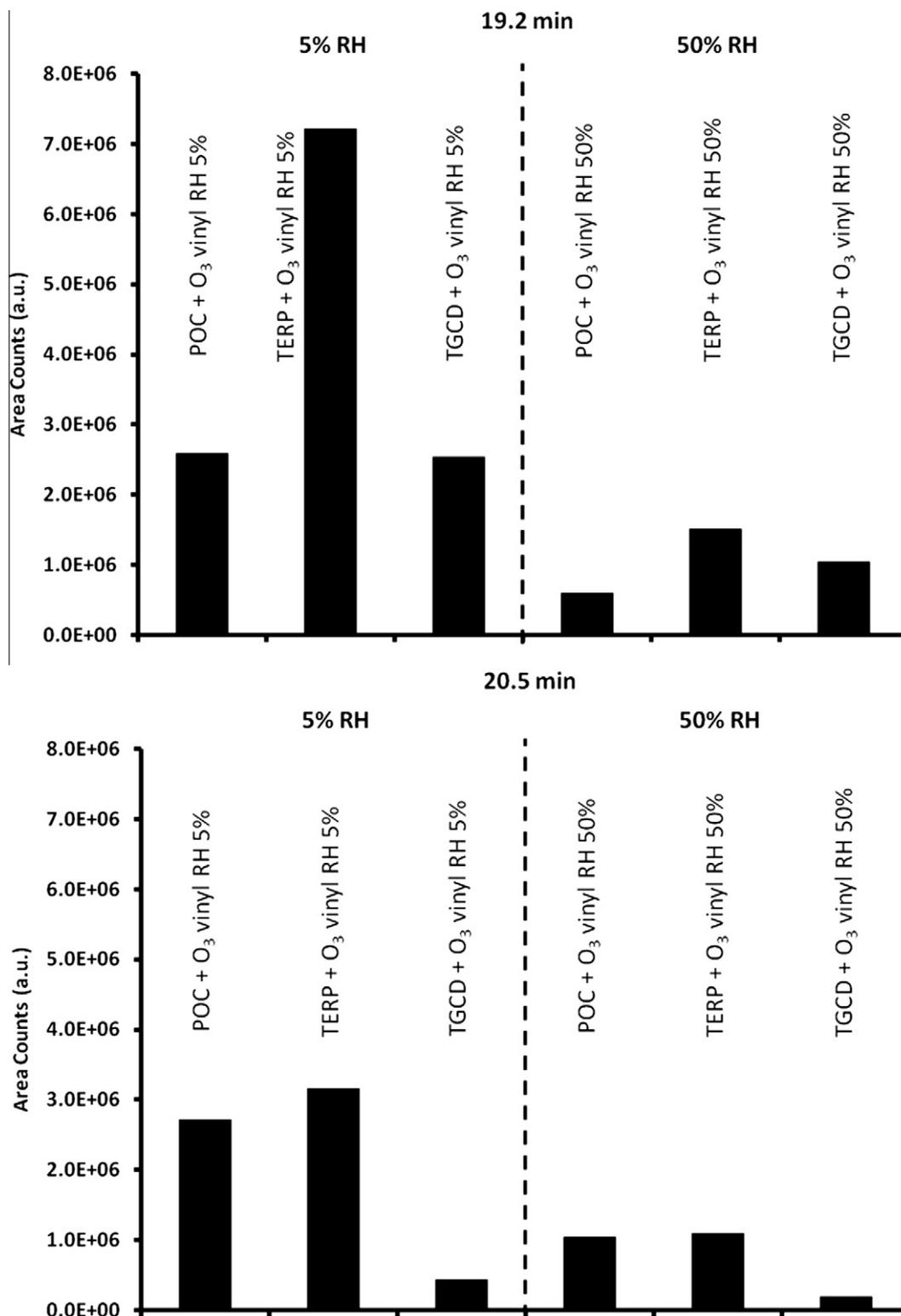
Derivatization of nonsymmetric carbonyls using PFBHA typically resulted in multiple chromatographic peaks due to geometric isomers of the oximes, Fig. 1. Identification of multiple peaks of the same oxime compound is relatively simple since the mass spectra

for each chromatographic peak of a particular oxime are almost identical. Typically, the PFBHA-derivatized oximes' (generic structure: F<sub>5</sub>C<sub>6</sub>CH<sub>2</sub>ON = C(R<sub>1</sub>)(R<sub>2</sub>)) mass spectra included an ion at *m/z* 181 ([CH<sub>2</sub>C<sub>6</sub>F<sub>5</sub>]<sup>+</sup> fragment) with a large relative intensity (>40%) and a [PFBHA oxime + 181]<sup>+</sup> ion (due to reactions in the ion trap mass spectrometer). In most cases, the *m/z* 181 ion relative intensity for the chromatographic peaks due to terpene + O<sub>3</sub> reaction product oximes was either the largest or one of the largest in the mass spectrum and was used to generate selected ion chromatograms (Yu et al., 1998).

Common POCs consist of a mixture of a number of terpenoid compounds in differing concentrations. Although the formulations can vary widely, the three major terpenes found in POCs are  $\alpha$ -terpineol, terpinolene, and d-limonene as observed in GC/MS analysis and reported in recent literature (Singer et al., 2006). When appropriately diluted in water (per instructions) and 0.4 g of this solution is applied to a vinyl flooring surface (as described above) in the presence of 100 ppb of ozone, a number of carbonyl com-

**Table 3**Additional reaction products that were observed but were not due to  $\alpha$ -terpineol reactions with O<sub>3</sub> at RH = 50% on vinyl flooring tiles.

Retention time (min)	Structure * indicates proposed identity	POC mixture	Mixture 1 (TAL)	Mixture 2 (TALD)	Mixture 3 (TGCD)	Mixture 4 (ALGCD)
9.9	 glycolaldehyde	-	-	-	X	X
17.2	 6-hydroxy-6-methyl-2-heptanone	-	-	X	X	X
17.7						
19.9	 6-hydroxy-4-methylhexanal	-	-	-	X	X
21.1	 5-methyl-4-hexenal *	-	-	-	X	X



**Fig. 2.** Averaged peak areas for POC + O<sub>3</sub>,  $\alpha$ -terpineol + O<sub>3</sub>, and TGCD mixture + O<sub>3</sub> reactions at 5% and 50% relative humidity for the (a) 19.2 and (b) 20.5 min peaks proposed identities for these retention times are shown Table 2.

pounds were subsequently emitted into the gas-phase as shown in Fig. 1. When O<sub>3</sub> was absent from the air flowing to the surface, no carbonyl oxidation products were observed.

Recently the surface-phase reactions of  $\alpha$ -terpineol with O<sub>3</sub> on both glass and vinyl flooring tiles were investigated (Ham and Wells, 2008). Results showed the formation of a number of car-

bonyl species that emulated those observed in the POC + O<sub>3</sub> experiments seen here. When chromatograms from  $\alpha$ -terpineol + O<sub>3</sub> and POC + O<sub>3</sub> were overlaid and the resulting mass spectra were analyzed, the observed peaks from POC + O<sub>3</sub> matched three of the six products (at retention times: 19.2, 20.5, and 24.4 & 24.6 min, Table 2) observed in  $\alpha$ -terpineol/O<sub>3</sub> experiments indicating that

the  $\alpha$ -terpineol reaction products dominate the POC + O<sub>3</sub> gas-phase emission profile. The other products that were observed (glyoxal, methylglyoxal, and 4-oxopentanal) are likely formed through multiple reaction pathways with other terpenes, including  $\alpha$ -terpineol, as previously observed in both the gas-phase and surface-phase reactions (Wells, 2005; Ham et al., 2006; Forester et al., 2007a,b; Ham and Wells, 2008, 2009).

In the TALD + O<sub>3</sub> and TGCD + O<sub>3</sub> experiments, a number of oxidation products (retention times 9.9, 17.2 & 17.9, 19.9, 21.1, Table 3) were generated that were not due to reactions of O<sub>3</sub> with  $\alpha$ -terpineol. Both the 9.9 min peak (glycolaldehyde) and 21.1 min peak (5-methyl-4-hexenal) are likely products from geraniol, based on mechanisms in the literature and previous geraniol + O<sub>3</sub> gas-phase studies (Nunes et al., 2005; Forester et al., 2007b). The peaks observed at 17.2 and 17.9 min (6-hydroxy-6-methyl-2-heptanone) are due to the dihydromyrcenol + O<sub>3</sub> reaction as previously observed in surface-phase reactions on both glass and vinyl flooring tiles (Ham and Wells, 2009). The peak observed at 19.9 min (6-hydroxy-4-methylhexanal) is an oxidation product that was also observed in previous citronellol/O<sub>3</sub> gas-phase studies (Ham et al., 2006).

Previous studies of terpene alcohol reactions on surfaces have indicated that the OH moiety of these species orients itself to the surface resulting in a change in the rate and reaction pathways (Voges et al., 2007; Ham and Wells, 2008, 2009). In these series of experiments surface coverage from the solution mixtures was calculated to be on the order of 4–12 terpenoid molecules thick. It is possible that these terpenoid layers above the first OH bound layer exist in a terpenoid rich solution as “islands” or “pools” where O<sub>3</sub> may penetrate to initiate reactions. If the OH bound layer is responsible for the observed products, it is anticipated that increasing the number of terpene alcohols (TAL < TALD < TGCD) in a mixture applied to a surface and initiating a reaction with ozone will result in a possible reduction in some reaction product emission profiles, i.e., peak areas, due to surface site competition. When these mixtures were applied to the surface, those peaks attributable to  $\alpha$ -terpineol/O<sub>3</sub> reactions did not change in peak area. This indicates that  $\alpha$ -terpineol may have a more favorable attachment to the surface compared to the other terpene alcohols.

This observation could be explained by investigating some physical characteristics of the terpenoid compounds in mixtures as they are applied to the surface. This analysis could potentially be used to predict the reactive chemistry of other mixtures. Recently, the octanol–air partition coefficients ( $K_{OA}$ ) were calculated for several semivolatile organic compounds (SVOCs) to characterize chemical migration or redistribution in indoor environments (Weschler and Nazaroff, 2008). Those results showed that  $\log K_{OA}$ 's greater than 10 are indicative of strongly absorbing species that could persist for hundreds to thousands of hours in indoor environments (Meylan and Howard, 2005; Weschler and Nazaroff, 2008). Viewing the vinyl surface as simply another environmental surface, the calculated  $\log K_{OA}$ 's for  $\alpha$ -terpineol, dihydromyrcenol, geraniol, citronellol,  $\alpha$ -pinene and limonene are 6.6, 6.3, 6.8, 6.5, 3.6, and 4.3, respectively (EPA, 2000). Additionally, the Henry's law constants (HLC) for  $\alpha$ -terpineol, dihydromyrcenol, geraniol, citronellol,  $\alpha$ -pinene and limonene were calculated:  $1.6 \times 10^{-5}$ ,  $4.1 \times 10^{-5}$ ,  $5.9 \times 10^{-5}$ ,  $5.7 \times 10^{-5}$ ,  $1.1 \times 10^{-1}$ ,  $3.8 \times 10^{-1}$  atm<sup>3</sup> mol<sup>-1</sup>, respectively (EPA, 2000). While the calculated  $\log K_{OA}$ 's for the terpene alcohols are similar, the HLCs for these terpenoids indicated that  $\alpha$ -terpineol's lower value suggests that its emission will be slower thus enhancing its reaction probability with ozone. Furthermore, the lower  $\log K_{OA}$  values and higher HLCs for limonene and  $\alpha$ -pinene coincide with no observation of carbonyl oxidation products from limonene/O<sub>3</sub> or  $\alpha$ -pinene/O<sub>3</sub> reactions on the surface. These calculations are consistent with the observed reaction product emission data. When  $\alpha$ -terpineol was removed from

the mixture (ALGCD/O<sub>3</sub> experiment), no peaks were observed at retention times of 19.2, 20.5, and 24.4 or 24.6 min indicating that those products are specific to  $\alpha$ -terpineol + O<sub>3</sub>. Additionally, peaks associated with geraniol, citronellol, and dihydromyrcenol did not change in peak area compared to TGCD + O<sub>3</sub> experiments indicating the competition for surface sites at this particular concentration of terpenoid compounds does not affect gas-phase emissions.

All data described in the previous paragraphs were collected at 50% relative humidity to simulate a typical indoor environment. However, humidity can have a significant effect on the formation/observation of reaction products as compounds partition between the gas and aerosol phases (Jonsson et al., 2006; Vesna et al., 2009; Warren et al., 2009). To investigate the influence of relative humidity on product emissions, POC + O<sub>3</sub>,  $\alpha$ -terpineol + O<sub>3</sub>, and TGCD + O<sub>3</sub> experiments were conducted using urethane-coated vinyl flooring tiles and 5% relative humidity. Additionally, contrasting the emission profiles from the simplest (one component), four-component and an actual consumer product mixture was done to provide more insight into the impact multiple components, as well as humidity, have on surface ozone reactions. In these experiments, an increase in reaction product peak area was observed for all oxidation products listed in Table 2 from POC/O<sub>3</sub>,  $\alpha$ -terpineol + O<sub>3</sub>, and TGCD + O<sub>3</sub> experiments. Fig. 2a and b show the increase in peak areas for both the 19.2 (6-hydroxyhept-5-en-2-one) and 20.5 (5-(1-hydroxy-1-methyl ethyl)-2-methylcyclohex-2-en-1-one) minute products at 5% compared to 50% relative humidity. The increase in peak area suggests that at the lower humidity reaction products partition more into the gas-phase as water is less available to form aggregate clusters. It could also mean that these terpene alcohols compete for surface sites with water which will affect the surface concentration (density) of carbon–carbon double bonds available for reaction with ozone. Furthermore, at lower humidity it is possible that the sprayed terpenoid solution evaporates faster resulting in more  $\alpha$ -terpineol molecules reacting at the air–surface interface. However, when the surface rinsate was analyzed no difference in reaction product peak area was observed between 5% and 50% relative humidity, suggesting that relative humidity does not significantly impact terpene alcohol surface attachment.

#### 4. Conclusions

The results from these experiments demonstrated that the presence of  $\alpha$ -terpineol in a POC or other simulated mixture can be a predictor of the emissions from terpene/ozone surface reactions. This can be clearly seen in Fig. 1 where POC + O<sub>3</sub>,  $\alpha$ -terpineol + O<sub>3</sub> and ALGCD + O<sub>3</sub> are compared. The presence of terpene alcohols on surfaces may be the main source of carbonyl reaction product emissions in indoor environments. This information could be used to predict consumer product emissions and gives insight into more accurate indoor occupant exposure assessments.

#### Disclaimer

The findings and conclusions in this report are those of the authors and do not necessarily represent the official position of the Centers for Disease Control.

#### References

- Aschmann, S.M., Arey, J., Atkinson, R., 2002. OH radical formation from the gas-phase reactions of O-3 with a series of terpenes. *Atmos. Environ.* 36, 4347–4355.
- Aschmann, S.M., Tuazon, E.C., Arey, J., Atkinson, R., 2003. Products of the gas-phase reaction of O-3 with cyclohexene. *J. Phys. Chem. A* 107, 2247–2255.
- Atkinson, R., 1997. Gas-phase tropospheric chemistry of volatile organic compounds: 1. Alkanes and alkenes. *J. Phys. Chem. Ref. Data* 26, 215–290.

- Atkinson, R., Arey, J., 2003. Gas-phase tropospheric chemistry of biogenic volatile organic compounds: a review. *Atmos. Environ.* 37, S197–S219.
- Atkinson, R., Arey, J., Aschmann, S.M., Corchnoy, S.B., Shu, Y., 1995. Rate constants for the gas-phase reactions of cis-hexen-1-ol, cis-3-hexenylacetate, trans-2-hexenal, and linalool with OH and NO<sub>3</sub> radicals and O<sub>3</sub> at 296 ± 2 K, and OH radical formation yields from the O<sub>3</sub> reactions. *Int. J. Chem. Kinet.* 27, 941–955.
- Baker, J., Arey, J., Atkinson, R., 2004. Kinetics of the gas-phase reactions of OH radicals, NO<sub>3</sub> radicals and O<sub>3</sub> with three C-7-carbonyls formed from the atmospheric reactions of myrcene, ocimene and terpinolene. *J. Atmos. Chem.* 48, 241–260.
- Cain, W.S., Schmidt, R., Wolkoff, P., 2007. Olfactory detection of ozone and Limonene: reactants in indoor spaces. *Indoor Air* 17, 337–347.
- Coleman, B.K., Destailats, H., Hodgson, A.T., Nazaroff, W.W., 2008. Ozone consumption and volatile byproduct formation from surface reactions with aircraft cabin materials and clothing fabrics. *Atmos. Environ.* 42, 642–654.
- EPA, 2000. EPIWEB 4.0. Washington D.C., U.S. Environmental Protection Agency. <<http://www.epa.gov/opptintr/exposure/pubs/episuite.htm>>, (accessed 10.02.10).
- Fan, Z.H., Liou, P., Weschler, C., Fiedler, N., Kipen, H., Zhang, J.F., 2003. Ozone-initiated reactions with mixtures of volatile organic compounds under simulated indoor conditions. *Environ. Sci. Technol.* 37, 1811–1821.
- Flemmer, M.M., Ham, J.E., Wells, J.R., 2007. Field and laboratory emission cell automation and control system for investigating surface chemistry reactions. *Rev. Sci. Instrum.* 78, 014101.
- Forester, C.D., Ham, J.E., Wells, J.R., 2007a. β-Ionone reactions with ozone and OH radical: rate constants and gas-phase products. *Atmos. Environ.* 41, 8758–8771.
- Forester, C.D., Ham, J.E., Wells, J.R., 2007b. Geraniol (2,6-dimethyl-2,6-octadien-8-ol) reactions with ozone and OH radical: rate constants and gas-phase products. *Atmos. Environ.* 41, 1188–1199.
- Fruekilde, P., Hjorth, J., Jensen, N.R., Kotzias, D., Larsen, B., 1998. Ozonolysis at vegetation surfaces: a source of acetone, 4-oxopentanal, 6-methyl-5-hepten-2-one, and geranyl acetone in the troposphere. *Atmos. Environ.* 32, 1893–1902.
- Gunschera, J., Andersen, J.R., Schulz, N., Salthammer, T., 2009. Surface-catalysed reactions on pollutant-removing building products for indoor use. *Chemosphere* 75, 476–482.
- Ham, J.E., Wells, J.R., 2008. Surface chemistry reactions of α-terpineol [(R)-2-(4-methyl-3-cyclohexenyl)isopropanol] with ozone and air on a glass and a vinyl tile. *Indoor Air* 18, 394–407.
- Ham, J.E., Wells, J.R., 2009. Surface chemistry of dihydromyrcenol (2,6-dimethyl-7-octen-2-ol) with ozone on silanized glass, glass, and vinyl flooring tiles. *Atmos. Environ.* 43, 4023–4032.
- Ham, J.E., Proper, S.P., Wells, J.R., 2006. The gas-phase chemistry of citronellol with ozone and OH radical: rate constants and products. *Atmos. Environ.* 40, 726–735.
- Jonsson, A.M., Hallquist, M., Ljunstrom, E., 2006. Impact of humidity on the ozone initiated oxidation of limonene, D3-carene, and α-pinene. *Environ. Sci. Technol.* 40, 188–194.
- Jonsson, A.M., Hallquist, M., Saathoff, H., 2007. Volatility of secondary organic aerosols from the ozone initiated oxidation of alpha-pinene and limonene. *J. Aerosol Sci.* 38, 843–852.
- Ljubic, I., Sabljic, A., 2002. Theoretical study of the mechanism and kinetics of gas-phase ozone additions to ethene, fluoroethene, and chloroethene: a multireference approach. *J. Phys. Chem. A* 106, 4745–4757.
- Ljubic, I., Sabljic, A., 2005. Ozonolysis of fluoroethene: theoretical study of unimolecular decomposition paths of primary and secondary fluorozonide. *J. Phys. Chem. A* 109, 2381–2393.
- Meylan, W.M., Howard, P.H., 2005. Estimating octanol-air partition coefficients with octanol-water partition coefficients and Henry's law constants. *Chemosphere* 61, 640–644.
- Moise, T., Rudich, Y., 2002. Reactive uptake of ozone by aerosol-associated unsaturated fatty acids: kinetics, mechanism, and products. *J. Phys. Chem. A* 106, 6469–6476.
- Nazaroff, W.W., Cass, G.R., 1986. Mathematical-modeling of chemically reactive pollutants in indoor air. *Environ. Sci. Technol.* 20, 924–934.
- Nazaroff, W.W., Weschler, C.J., 2004. Cleaning products and air fresheners: exposure to primary and secondary air pollutants. *Atmos. Environ.* 38, 2841–2865.
- Nojgaard, J.K., 2010. Indoor measurements of the sum of the nitrate radical, NO<sub>3</sub>, and nitrogen pentoxide, N<sub>2</sub>O<sub>5</sub> in Denmark. *Chemosphere* 79, 898–904.
- Nojgaard, J.K., Bilde, M., Stenby, C., Nielsen, O.J., Wolkoff, P., 2006. The effect of nitrogen dioxide on particle formation during ozonolysis of two abundant monoterpenes indoors. *Atmos. Environ.* 40, 1030–1042.
- Nunes, F.M.N., Veloso, M.C.C., Pereira, P., de Andrade, J.B., 2005. Gas-phase ozonolysis of the monoterpenoids (S)-(+)-carvone, (R)-(-)-carvone, (-)-carveol, geraniol and citral. *Atmos. Environ.* 39, 7715–7730.
- OEHHA, 1999. Determination of acute reference exposure levels for airborne contaminants. Air Toxicology and Epidemiology Section, California Office of Environmental Health Hazard Assessment, Sacramento, CA, USA. <<http://www.oehha.org/air/allrels.html>>, (accessed 07.04.10).
- Pandrangi, L.S., Morrison, G.C., 2008. Ozone interactions with human hair: ozone uptake rates and product formation. *Atmos. Environ.* 42, 5079–5089.
- Sarwar, G., Corsi, R., 2007. The effects of ozone/limonene reactions on indoor secondary organic aerosols. *Atmos. Environ.* 41, 959–973.
- Sarwar, G., Corsi, R., Kimura, Y., Allen, D., Weschler, C.J., 2002. Hydroxyl radicals in indoor environments. *Atmos. Environ.* 36, 3973–3988.
- Singer, B.C., Destailats, H., Hodgson, A.T., Nazaroff, W.W., 2006. Cleaning products and air fresheners: emissions and resulting concentrations of glycol ethers and terpenoids. *Indoor Air* 16, 179–191.
- Tamas, G., Weschler, C.J., Bako-Biro, Z., Wyon, D.P., Strom-Tejsten, P., 2006. Factors affecting ozone removal rates in a simulated aircraft cabin environment. *Atmos. Environ.* 40, 6122–6133.
- Vesna, O., Sax, M., Kalberer, M., Gaschen, A., Ammann, M., 2009. Product study of oleic acid ozonolysis as function of humidity. *Atmos. Environ.* 43, 3662–3669.
- Vibenholt, A., Norgaard, A.W., Clausen, P.A., Wolkoff, P., 2009. Formation and stability of secondary ozonides from monoterpenes studied by mass spectrometry. *Chemosphere* 76, 572–577.
- Voges, A.B., Stokes, G.Y., Gibbs-Davis, J.M., Lettan, R.B., Bertin, P.A., Pike, R.C., Nguyen, S.T., Scheidt, K.A., Geiger, F.M., 2007. Insights into heterogeneous atmospheric oxidation chemistry: development of a tailor-made synthetic model for studying tropospheric surface chemistry. *J. Phys. Chem. C* 111, 1567–1578.
- Warren, B., Malloy, Q.G.J., Yee, L.D., Cocker, D.R., 2009. Secondary organic aerosol formation from cyclohexene ozonolysis in the presence of water vapor and dissolved salts. *Atmos. Environ.* 43, 1789–1795.
- Wells, J.R., 2005. Gas-phase chemistry of alpha-terpineol with ozone and OH radical: rate constants and products. *Environ. Sci. Technol.* 39, 6937–6943.
- Wells, J.R., Morrison, G.C., Coleman, B.K., 2008. Kinetics and reaction products of ozone and surface-bound squalene. *J. ASTM Int.* 5, JA1101629.
- Weschler, C.J., 2001. Reactions among indoor pollutants. *J. Sci. World* 1, 443–457.
- Weschler, C.J., Nazaroff, W.W., 2008. Semivolatile organic compounds in indoor environments. *Atmos. Environ.* 42, 9018–9040.
- Weschler, C.J., Shields, H.C., 1996. Production of the hydroxyl radical in indoor air. *Environ. Sci. Technol.* 30, 3250–3258.
- Weschler, C.J., Brauer, M., Koutrakis, P., 1992. Indoor ozone and nitrogen-dioxide – a potential pathway to the generation of nitrate radicals, dinitrogen pentoxide, and nitric-acid indoors. *Environ. Sci. Technol.* 26, 179–184.
- Weschler, C.J., Shields, H.C., Naik, D.V., 1994. Indoor chemistry involving O<sub>3</sub>, NO and NO<sub>2</sub> as evidenced by 14 months of measurements at a site in Southern California. *Environ. Sci. Technol.* 28, 2120–2132.
- Wisthaler, A., Tamas, G., Wyon, D.P., Strom-Tejsten, P., Space, D., Beauchamp, J., Hansel, A., Mark, T.D., Weschler, C.J., 2005. Products of ozone-initiated chemistry in a simulated aircraft environment. *Environ. Sci. Technol.* 39, 4823–4832.
- Yu, J.Z., Flagan, R.C., Seinfeld, J.H., 1998. Identification of products containing –COOH, –OH, and –C=O in atmospheric oxidation of hydrocarbons. *Environ. Sci. Technol.* 32, 2357–2370.