

Wood Dust Sampling: Field Evaluation of Personal Samplers When Large Particles Are Present

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Recent recommendations for wood dust sampling include sampling according to the inhalable convention of International Organization for Standardization (ISO) 7708 (1995) Air quality—particle size fraction definitions for health-related sampling. However, a specific sampling device is not mandated, and while several samplers have laboratory performance approaching theoretical for an ‘inhalable’ sampler, the best choice of sampler for wood dust is not clear. A side-by-side field study was considered the most practical test of samplers as laboratory performance tests consider overall performance based on a wider range of particle sizes than are commonly encountered in the wood products industry. Seven companies in the wood products industry of the Southeast USA (MS, KY, AL, and WV) participated in this study. The products included hardwood flooring, engineered hardwood flooring, door skins, shutter blinds, kitchen cabinets, plywood, and veneer. The samplers selected were 37-mm closed-face cassette with ACCU-CAPTM, Button, CIP10-I, GSP, and Institute of Occupational Medicine. Approximately 30 of each possible pairwise combination of samplers were collected as personal sample sets. Paired samplers of the same type were used to calculate environmental variance that was then used to determine the number of pairs of samples necessary to detect any difference at a specified level of confidence. Total valid sample number was 888 (444 valid pairs). The mass concentration of wood dust ranged from 0.02 to 195 mg m⁻³. Geometric mean (geometric standard deviation) and arithmetic mean (standard deviation) of wood dust were 0.98 mg m⁻³ (3.06) and 2.12 mg m⁻³ (7.74), respectively. One percent of the samples exceeded 15 mg m⁻³, 6% exceeded 5 mg m⁻³, and 48% exceeded 1 mg m⁻³. The number of collected pairs is generally appropriate to detect a 35% difference when outliers (negative mass loadings) are removed. Statistical evaluation of the nonsimilar sampler pair results produced a finding of no significant difference between any pairing of sampler type. A practical consideration for sampling in the USA is that the ACCU-CAPTM is similar to the sampler currently used by the Occupational Safety and Health Administration for purposes of demonstrating compliance with its permissible exposure limit for wood dust,

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which is the same as for Particles Not Otherwise Regulated, also known as inert dust or nuisance dust (Method PV2121).

Keywords: ACCU-CAP™; Button sampler; CIP10-I sampler; GSP sampler; inhalable sampling; IOM sampler; wood dust

INTRODUCTION

Over a half million US workers (519 651) were employed in wood product industries in 2007 (US Census Bureau) and ~3.6 million workers in 25 member countries of the European Union are estimated to be exposed to occupational wood dust (Kauppinen *et al.*, 2006). The International Agency for Research on Cancer (IARC) found sufficient evidence of carcinogenicity of wood dust in humans and consequently, wood dust has been classified as a human carcinogen (Group 1; IARC, 1995). The Industrial Injuries Advisory Council (IIAC) in the UK more recently concluded the existence of an association between wood dust exposure and nasopharyngeal cancer based on a literature review (IIAC, 2007). In addition, exposure to wood dust has been implicated in nonmalignant respiratory diseases, including obstructive disease of the lower airways and reactive disease of the upper airways (Whitehead *et al.*, 1981a; Goldsmith and Shy, 1988; Enarson and Chan-Yeung, 1990). However, excess longitudinal decline in lung function with exposure to wood dust was not observed in two recent epidemiological surveys of wood processing industry workers (Innocenti *et al.*, 2006; Glindmeyer *et al.*, 2008). The European Scientific Committee on Occupational Exposure Limits (SCOEL, 2003) concluded that exposure to wood dust $>0.5 \text{ mg m}^{-3}$ induces pulmonary effects and should be avoided. The National Institute for Occupational Safety and Health Recommended Exposure Limit for wood dust is 1 mg m^{-3} . In 2005, the American Conference of Governmental Industrial Hygienists (ACGIH) adopted a threshold limit value (TLV)–time-weighted average (TWA) of 1 mg m^{-3} for all wood species except western red cedar. Oak and beech were classified as confirmed human carcinogen (A1), birch, mahogany, teak, and walnut were classified as suspected human carcinogen (A2), and all other wood dusts were categorized as not classifiable as a human carcinogen (A4), and this remains the current TLV (ACGIH, 2009). The wood dust TLV also calls for sampling according to the inhalable penetration convention of the ISO Standard 7708, (1995) Air quality—particle size fraction definitions for health-related sampling [International Organization for Standardization

(ISO)]. The collected sample is known as inhalable particulate matter (IPM) and it includes all particles likely to enter the nose or mouth of a normally breathing worker. In addition, in 2005, French National Decree 2003, 1254 came into force of law requiring exposures be maintained $<1 \text{ mg m}^{-3}$ IPM in that country. It should be noted that the French standard currently requires the use of a 37-mm closed-face cassette sampler (CFC—see below), although there has been doubts raised as to whether this correctly represents an IPM sample.

Most airborne wood dust particles are $>10 \mu\text{m}$ in size (Whitehead *et al.*, 1981b; Hinds, 1988; Pisaniello *et al.*, 1991; Tatum *et al.*, 2001; Harper *et al.*, 2004; Verma *et al.*, 2007) and relatively little mass of wood dust is found in smaller particle sizes. Occupational wood dust exposure levels are determined by gravimetric analysis, such as NIOSH method 0500 (NIOSH, 1994) and mass measurements are dominated by larger particles as a result of the cubic relation between diameter and mass. NIOSH Method 0500 does not list a sampler for IPM but instead refers to the 37-mm plastic CFC, which may not have the necessary efficiency for large particle collection to match the ISO inhalable mass fraction convention. Many studies have compared occupational wood dust exposure levels using samplers whose performance is expected to conform to the ISO inhalable convention and the CFC sampler (Martin and Zalk, 1998; Kim and Lee, 1996; Perrault *et al.*, 1996; Tatum *et al.*, 2001; Harper and Muller, 2002; Kauffer *et al.*, 2010) and all studies have confirmed that the CFC collects less sample mass than an IPM sampler sampling the same atmosphere. The numerical value for the current ACGIH IPM limit (ACGIH, 2001) was based on prior workplace studies, which used the CFC but with a correction factor applied to account for the different relative collection efficiencies of the CFC and an inhalable sampler known as the Institute of Occupational Medicine (IOM) sampler. This correction factor, 2.5 or 250%, was developed based on side-by-side studies of the CFC and IOM in a variety of industries (Werner *et al.*, 1996) with a variety of aerosols of different particle size distributions, and woodworking industries represented only a small part of the

overall data set. Ratios found in woodworking industries have tended to be >2.5, typically ~3.5, but this may be due in part to the influence of the orientation of the CFC sampler (Görner *et al.*, 2010; Kauffer *et al.*, 2010) and in part due to particles depositing on the walls of the CFC either during sampling or after sampling in transit to the laboratory (Harper and Demange, 2007). There is no current consensus on which, if any, of the currently available samplers should be recommended for sampling wood dust. While several samplers have laboratory performance approaching the ideal for an ‘inhalable’ sampler, the best choice of sampler for wood dust in field situations is not clear due to other issues including the collection of particles exceeding the upper limit of the inhalable convention and projectile particles that may enter the sampler but not necessarily enter the head airways. The present study aims to compare wood dust personal sampling with five different aerosol samplers in the field, in order to evaluate difference between them and possibly to inform decisions about the most appropriate sampler and sampling method for wood dust collection.

METHODS AND MATERIALS

Sampling sites

Seven different wood product industries of the Southeast USA (AL, KY, MS, and WV) participated in this study, including veneer, plywood, engineered hardwood floor, door skin, shutter, hardwood floor, and kitchen cabinet makers (Table 1). There were a total of 888 individual samples as 444 pairs. Most of the samples were collected where significant wood dust level was anticipated, for example, where the processes involved cutting, drilling, sanding, or sawing.

Samplers used for this study

Five different samplers were utilized for this study, including the CFC with ACCU-CAP™, Button, CIP10-I, GSP, and IOM samplers (Fig. 1a–e).

ACCU-CAP™ (SKC Inc., Omega Division, Eighty Four, PA, USA) is a 37-mm polyvinyl chloride (PVC) filter with 5-μm pore size sealed to the base of a clear PVC dome with a hole at the vertex. The ACCU-CAP™ is a one-piece filter capsule similar to that used (Moore *et al.*, 1990) in the OSHA method PV2121 (OSHA, 2003). When included in the CFC, the ACCU-CAP™ can prevent losses of sample that would otherwise be deposited on the interior wall of the cassette (Puskar *et al.*, 1992). The ACCU-CAP™ fits inside of the top and bottom pieces of a two-piece 37-mm CFC (SKC Inc.) with back up pad under the filter and with the hole at the vertex of the dome fitting to the entry hole of the top piece of the cassette. A 2 l min⁻¹ flow rate was used. The CFC was located on the body of the wearer at the commencement of sampling with the opening facing at an ~45° angle to the vertical as previously published (Buchan *et al.*, 1986). However, this position was not fixed with a holder, and at the end of sampling, the CFC was often observed to be pointing face outward from the body in the orientation shown by Kauffer *et al.* (2010) to be the most appropriate for inhalable sampling.

Button sampler (SKC Inc.) has a spherical shell inlet with numerous holes of nominally 381-μm diameter functioning as orifices. A 25-mm PVC filter was used in the Button for sampling at company A but it resulted in many pump failures because of excessive pressure drop across the filter at the recommended flow rate of 4 l min⁻¹ for sampling in accordance with the inhalable convention, which was consistent with previous studies (e.g. Reynolds *et al.*, 2009). Thus, the remainder of the sampling surveys used a 25-mm glass fiber filter (SKC Inc.)

Table 1. Wood products of participating companies and collected sample number

Company	Product	Wood type	Number of samples	Air sampled process
A	Veneer	Walnut, cherry, white oak, and maple	112	Cutting, debarking, slicing, and sorting
B	Plywood	Pine	106	Cutting and pressing
C	Engineered hardwood flooring	Red oak, maple, walnut, hickory, and poplar	120	Cutting, cleaning, and sorting
D	Door skins	Pine, oak, and gum	124	Chipping, cleaning, and sawing
E	Shutters	Western red cedar	128	Cutting, drilling, and sanding
F	Hardwood flooring	Red oak, white oak, and maple	154	Cutting and sorting
G	Kitchen cabinetry	Oak, maple, cherry, and MDF	144	Cutting and sanding

MDF, medium-density fibreboard.

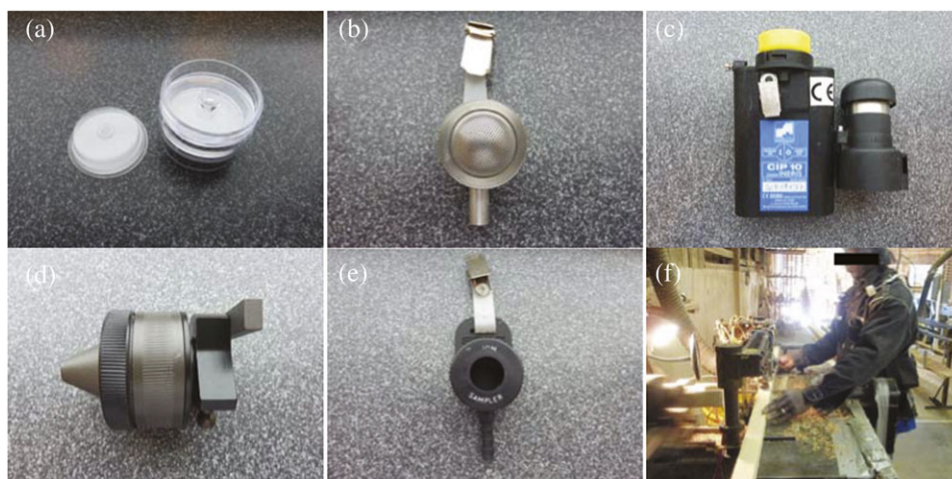


Fig. 1. Tested aerosol samplers. (a) ACCU-CAP™, (b) Button, (c) CIP10-I, (d) GSP, (e) IOM sampler, and (f) a picture of wood dust sampling (the worker was wearing ACCU-CAP™ and CIP10-I samplers).

instead. Validation of the gravimetric procedure has shown that glass fiber filters can be used in place of PVC filters with only slightly greater uncertainty in the result (McLister *et al.*, 2001). The uniform distribution of the holes on the curved inlet results in an even distribution of particles on the filter surface (Kalatoor *et al.*, 1995). Aizenberg *et al.* (2000) investigated the performance of the sampler for particles in the inhalable size range and showed the collection efficiency to lie between that of the CFC and IOM samplers. The Button sampler inlet screen should minimize the collection of particles $>100\ \mu\text{m}$ (i.e. those not covered by the ACGIH inhalable convention) in the same manner as the screen proposed for the IOM sampler by Aitken and Donaldson (1996). In laboratory and field studies (Aizenberg *et al.*, 2000), the sampler was shown to be relatively insensitive to wind speed and direction. The performance of the sampler in a second laboratory study was not as good a match to the inhalable convention, but the sampler was shown to have minimal internal wall losses of sample (Li *et al.*, 2000).

CIP10-I (Arelco ARC, Fontenay-sous-Bois, France) is a French sampler (Görner *et al.*, 1999) that also performed well when compared to the inhalable convention (Kenny *et al.*, 1997; Bartley, 1998). It operates at a flow rate of $10\ \text{l min}^{-1}$ by the unusual method of spinning a foam disk. Wood dusts are captured in the polyurethane foam and holder, which are weighed. There are practical issues for this sampler. First, there is an issue concerning the weight stability for foam so that at least equilibration overnight at constant temperature and humidity is recommended

by the manufacturer before weighing. Another issue is the large overall mass of the cup and foam combination. Finally, the CIP10-I has to be calibrated in the laboratory using assisted pressure drop compensation and the calibration can only be checked in the field by measuring the revolutions per minute (RPM) of the rotating cup. This situation has the potential for errors.

GSP sampler (Gesamtstaub-Probenahmesystem; GSMGesellschaft für Schadstoffmesstechnik, GmbH, Neuss-Norf, Germany) is a German sampler. Results from the collaborative study previously mentioned (Kenny *et al.*, 1997) suggested that the GSP sampler may operate sufficiently in accordance with the inhalable convention for it to be regarded as a possible inhalable sampling device (Bartley, 1998). These results were confirmed at the University of Cincinnati (Aizenberg *et al.*, 2000). A conductive plastic version of this sampler was evaluated for wood dust (Davies *et al.*, 1999). However, the conductive plastic cassette is also subject to weighing errors (Li *et al.*, 2000). The samplers used in this study were the original cast aluminum samplers. The flow rate was $3.5\ \text{l min}^{-1}$ and PVC filters were used to collect the wood dusts.

IOM sampler (SKC Inc.) was designed to collect the inhalable fraction of an aerosol (Mark and Vincent, 1986). The IOM showed good orientation averaged performance agreement with the ACGIH inhalable convention curve (within 10%) at averaged wind velocities of 0.5 and $1\ \text{m s}^{-1}$ (Bartley, 1998). However, several studies reported that the IOM is susceptible to significant bias in situations where there is a constant directional component to the

airflow (Roger *et al.*, 1998; Aizenberg *et al.*, 2000; Li *et al.*, 2000). Both particles collected on the filter as well as the particles collected on the inner inlet surfaces are analyzed since the entire stainless steel cassette is weighed. The IOM is characterized by a large 15-mm entry orifice and is operated facing orthogonally outward from the body at a flow rate of 2 l min^{-1} loaded with a 25-mm PVC filter. The IOM sampler has been criticized as a sampler for certain dusts because of the large entry inlet, which increases the potential for aspiration of particles $>100 \text{ }\mu\text{m}$ aerodynamic equivalent diameter (Lidén and Kenny, 1994; Aitken and Donaldson, 1996; Aizenberg *et al.*, 2000; Lidén *et al.*, 2000). Wood dust particles of this size are present in the breathing zone (Vaughan *et al.*, 1990; Martin and Zalk, 1998; Harper *et al.*, 2004) but they are not covered by the inhalable convention because the aspiration efficiency of the human mouth and nose for such large particles has been shown to be low (Breyse and Swift, 1990; Aitken and Donaldson, 1996; Hsu and Swift, 1999; Kennedy and Hinds, 2002; Dai *et al.*, 2006).

Combination of samplers

Workers were asked to wear commercial back braces and the samplers were located on the shoulder straps of these back braces, one on each side, with sides randomized for different pairs. Fifteen possible combination of the samplers were generated: ACCU-CAPTM/ACCU-CAPTM, ACCU-CAPTM/Button, ACCU-CAPTM/CIP10-I, ACCU-CAPTM/GSP, ACCU-CAPTM/IOM, Button/Button, Button/CIP10-I, Button/GSP, Button/IOM, CIP10-I/CIP10-I, CIP10-I/GSP, CIP10-I/IOM, GSP/GSP, GSP/IOM, and IOM/IOM. Combinations were selected at random for each sampling event so that no deliberate bias was introduced by having a large percentage of any one combination being associated with any particular site or task. The sample size was determined by the method used in our previous study (Harper and Muller, 2002). Paired samplers of the identical type were used to calculate environmental variance (true field coefficient of variation) that was then used to determine the number of pairs of samples necessary to detect any difference at a specified level of confidence.

Wood dust sampling

CFC with ACCU-CAPTM samplers were prepared by using a cassette closer (Omega Specialty Instruments; now SKC Inc., Omega Division) connected to a cylinder of nitrogen to apply an even force to close each cassette. Assembled cassettes were wrap-

ped with Omega Gel Bands (Omega Specialty Instruments) to prevent air leaks and leak checked using a field cassette leak tester (Omega Specialty Instruments). The CFC samplers were opened using an EZ Cassette Opener (SKC Inc.). The IOM samplers with stainless steel cassettes, GSP, and Button samplers were prepared in accordance with the manufacturer's instructions. The pumps were attached to a back belt (Safe-T-Lift, Style No. 70-110543; FLA Orthopedics, Inc.) around the waist of the participants (note that the belts are used in this case only for support of the sampling equipment and an endorsement of such belts for any other purposes is not intended) (Fig. 1f). Aircheck PCXR-4 pumps (SKC Inc.) were connected to the filter holders (other than the CIP10-I) by flexible tubing. The flow rates through the sampling trains were calibrated using a BIOS Dry Cal Meter (Product DCL-MH; BIOS International Corporation, Butler, NJ, USA), using sampler calibration adapters (e.g. IOM Calibration Adapter; SKC Inc.) where necessary. The flow rates were calibrated before and after each day of sampling on-site and to ensure the flow rate did not change significantly. The flow rate of the CIP10-I was calibrated in the NIOSH laboratory with a CIP10 Calibration bench (Arelco, ARC) and an on-site check by measurement of the rotational speed (RPM) of the cup was conducted before and after field sampling. Sampling times were adjusted according to the judgment of the on-site hygienist to obtain optimal particle deposition on the filters and the sampling time was between 1 and 4 h in most of the samples. Most of the wood workshops used kiln-dried wood with very low moisture contents, leading to dust with a very high electrostatic charge, requiring careful handling of the samplers and samples to avoid sample losses. All samples were capped and stored in ziplock bags during transportation from the sampling sites. Samples were taken by hand to the laboratory rather than shipped.

Gravimetric analysis

All filters and foams from the first three sites (A, B, and C) were equilibrated for a minimum 72 h to specific humidity conditions before and after sampling in a cabinet containing a saturated solution of sodium dichromate. Later, a purpose-built weighing room was used for gravimetric analysis for the following four sites (D, E, F, and G). The weighing room maintains constant relative humidity ($50\% \pm 2$) and temperature ($26^\circ\text{C} \pm 2$). All samples were weighed before and after sampling using a micro Balance (UMT2; Mettler-Toledo, Columbus, OH, USA) and weights were recorded on data sheets.

For IOM filter and stainless steel cassette and CIP10-I polyurethane form and rotating cup, measurements were made with a dual range analytical balance (AG245; Mettler-Toledo). A calibration check on the balances was performed and recorded each day of weighing and the balances were zeroed between each weighing. Filters were passed through an electrostatic bar (Mettler-Toledo) before they were weighed to dissipate static charge. Measurements were made after allowing exactly 120 s for balance stabilization. Field blanks were taken but samples were not corrected for field blanks, according to normal practice. Average mass difference from 156 field blank samples was -0.03 mg.

Statistical analysis

Analyses were performed on matched pairs after calculation of air concentrations using SAS 9.2 (SAS Institute, Cary, NC, USA). All variables were tested to ensure that they met statistical assumptions for all analyses. Log-transformed data were used where appropriate. In this sample, there were 995 nonmissing data points. Data were excluded if the mass loading and, therefore, the concentration was less than zero. Data were paired, and if either of the paired sampler concentrations were missing, both pairs were excluded. In the final data set, there were 444 pairs, giving 888 data points, resulting in a loss of 11% of the data.

Associations were performed using correlation analyses to see how well-paired samples correlated with each other. Pearson's correlation coefficients and their corresponding *P*-values were generated. Whether or not correlation coefficients were of statistical significance, the size of the coefficient itself may still be meaningful, with values approaching 1 or -1 indicating stronger associations (Table 2).

Mixed linear models were used to look for statistically significant differences between the paired samples. Random effect models were used to control for and test variation between the different sampling sites. Nonsignificant *P*-values ($P > 0.05$) indicate that the samplers were not statistically different, with mean values close to zero indicating similar readings.

Correction factors have also been calculated. These values show what the paired group's values would have to be multiplied by in order to make them match up as close as possible to each other, with the correction factor being applied to the sampler that comes last alphabetically (i.e. for samplers ACCU-CAPTM and Button, the correction factor would be applied to the Button sampler). Values close to one indicate that the samplers' data

were initially close together. Values for matched samplers (ACCU-CAPTM/ACCU-CAPTM, Button/Button, CIP10-I/CIP10-I, GSP/GSP, and IOM/IOM) are given to show the potential variability within sampler type, although they may not be meaningful in an applied sense.

RESULTS

Total valid sample number was 888 among 995 nonmissing data. There were originally 1257 data points, including 156 blank samples, but some data were excluded due to the pump failure (62 individual samples), mishandling samples (9 individual samples), and negative mass [35 individual samples, mainly from the CIP10-I sampler in site B, where malfunction possibly related to the weighing or calibration issues referred to previously is suspected since the observed sampling rate, even with outliers omitted, is markedly lower than at any other site (see Fig. 2 below)]. The mass concentration of wood dust ranged from 0.02 to 195 mg m⁻³ for total data set. Since the mass concentration of wood dust showed log normal distributions, the results can also be described using geometric mean (GM) and geometric standard deviation (GSD) as descriptive statistics. GM (GSD) and arithmetic mean (standard deviation) of wood dust were 0.98 mg m⁻³ (3.06) and 2.12 mg m⁻³ (7.74), respectively. One percent of samples exceeded 15 mg m⁻³, 6% of samples exceeded 5 mg m⁻³, and 48% of samples exceeded 1 mg m⁻³. The obtained data were separated by sampling sites, by each sampler, and by pair of sampler for comparison. It should be acknowledged that the data were not 8-h TWA results, but the range of results was consistent with that observed across European countries (Kauppinen, *et al.*, 2006).

Comparison by sampling sites

The GM of mass concentration of wood dust in seven different wood product industries ranged from 0.44 (Site B) to 2.08 mg m⁻³ (Site F). In accordance with the mixed model analysis of variance, Sites D and B were both significantly different from Site F, but all other pairs were not significantly different at the $P < 0.05$ level. The GM with GSD for each site by separating with each sampler is shown in Fig. 2. The highest GM of dust concentrations recorded at each site were (A) IOM (1.04 mg m⁻³), (B) GSP (0.56 mg m⁻³), (C) IOM (1.25 mg m⁻³), (D) IOM (0.79 mg m⁻³), (E) GSP (1.65 mg m⁻³), (F) CIP10-I and IOM (2.51 mg m⁻³; although the IOM had a larger variance), and (G) CIP10-I (2.01 mg m⁻³) samplers.

Table 2. Correlation coefficient, average difference, and correction factors between pair of the samplers

Sampler pairs	Sample number	Correlation coefficient (<i>P</i> value)	Average difference between the groups (mg m^{-3}) (<i>P</i> value) ^a	Correction factors ^b
ACCU-CAP™/ACCU-CAP™	29	0.26 (0.264) ^c	6.66 (0.324) ^c	0.16 ± 0.50^c
ACCU-CAP™/Button	30	0.94 (<0.001)	0.10 (0.885)	0.91 ± 0.46
ACCU-CAP™/CIP10-I	30	0.87 (<0.001)	1.08 (0.267)	0.60 ± 1.52
ACCU-CAP™/GSP	33	0.95 (<0.001)	0.98 (0.709)	0.77 ± 0.75
ACCU-CAP™/IOM	32	0.46 (0.008)	0.55 (0.270)	0.64 ± 0.38
Button/Button	30	0.96 (<0.001)	0.03 (0.933)	0.97 ± 0.34
Button/CIP10-I	32	0.22 (0.219)	0.33 (0.529)	0.82 ± 1.18
Button/GSP	30	0.32 (0.087)	0.29 (0.404)	0.79 ± 0.57
Button/IOM	24	0.80 (<0.001)	0.05 (0.913)	0.95 ± 0.52
CIP10-I/CIP10-I	24	0.94 (<0.001)	0.27 (0.714)	1.13 ± 1.02
CIP10-I/GSP	30	0.70 (<0.001)	0.11 (0.871)	1.08 ± 0.35
CIP10-I/IOM	28	0.99 (<0.001)	0.22 (0.924)	1.08 ± 0.39
GSP/GSP	30	0.91 (<0.001)	0.17 (0.734)	0.88 ± 0.26
GSP/IOM	31	0.91 (<0.001)	0.03 (0.982)	0.99 ± 0.39
IOM/IOM	31	0.74 (<0.001)	0.75 (0.573)	0.74 ± 0.74

^aAverage difference between the groups was calculated using absolute value difference between the sampler groups.

^bCorrection factors is applied to the second sampler to make average mass concentration same as the first sampler in sampler pairs; \pm indicates 95% confidence interval value.

^cCorrelation coefficient becomes 0.83 ($P < 0.001$), the average difference becomes 0.01 mg m^{-3} ($P = 0.98$), and the correction factor becomes 1.02 ± 0.50 with the removal of a single outlier.

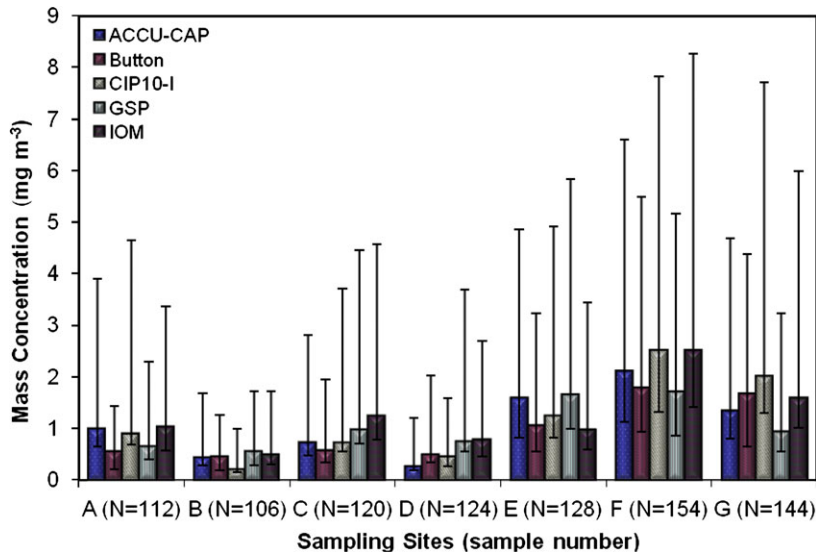


Fig. 2. GM and GSD wood dust mass concentration at each sampling site by each sampler. *N* is the combined sample number for all samplers.

Comparison by samplers

The box plot for wood dust mass concentration collected by each sampler at the seven different wood product industries with total sample number (*N*) is shown in Fig. 3. The median mass concentration of wood dust for ACCU-CAP™, Button,

CIP10-I, GSP, and IOM samplers were 0.90, 0.87, 1.08, 0.95, and 1.06, respectively. The mass concentrations of wood dust collected by ACCU-CAP™ ranged from 0.02 to 195 mg m^{-3} with GM (GSD) of 1.03 mg m^{-3} (3.25). The mass concentrations of wood dust collected by Button sampler

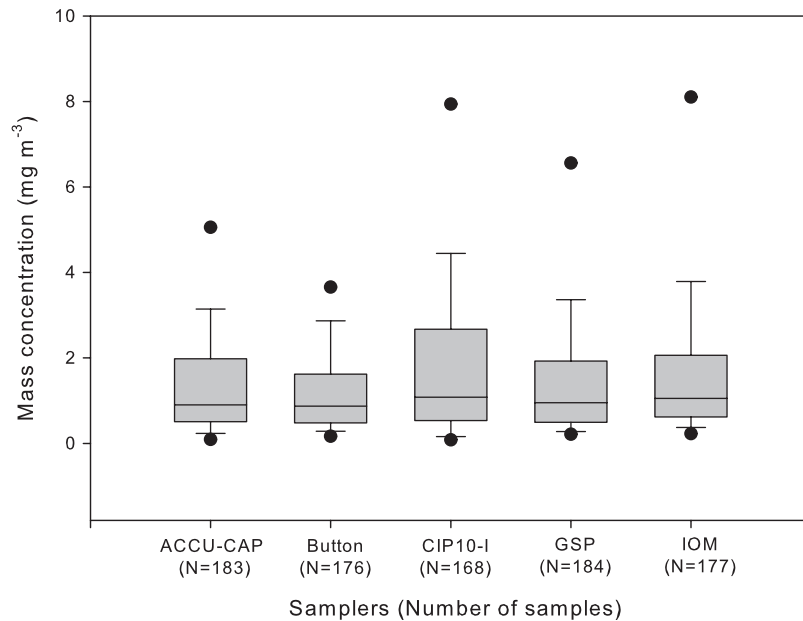


Fig. 3. The box plot of wood dust mass concentration (milligrams per cubic meter) by different samplers without outliers. The horizontal lines in the box plot from bottom to top indicate 10th, 25th, 50th (median), 75th, and 90th percentiles. The circles indicate the 5th (lower circle) and 95th (upper circle) percentiles. *N* is the number of samples for each sampler, and the total sample number is 888 when all samplers combined.

ranged from 0.03 to 7.68 mg m⁻³ with GM (GSD) of 0.88 mg m⁻³ (2.64). The mass concentrations of wood dust collected by CIP10-I sampler ranged from 0.02 to 49.5 mg m⁻³ with GM (GSD) of 1.01 mg m⁻³ (3.46). The mass concentrations of wood dust collected by GSP ranged from 0.07 to 29.9 mg m⁻³ with GM (GSD) of 0.90 mg m⁻³ (2.94) mg m⁻³. The mass concentrations of wood dust collected by IOM sampler (filter plus cassette) ranged from 0.14 to 38.7 mg m⁻³ with GM (GSD) of 1.28 mg m⁻³ (2.80). The mass concentrations from all five different samplers were not statistically significant different ($P > 0.05$). GM (GSD) of the IOM sampler with filter-only analysis was 1.06 mg m⁻³ (2.92) and the range of the concentration was between 0.04 and 37.39 mg m⁻³. The GM and arithmetic mean difference between the concentrations from IOM samples weighing the filter-only and samples weighing both the filter plus cassette were 0.28 and 0.24 mg m⁻³, respectively. The two groups are not significantly different ($P = 0.78$) and the filter-only concentrations were not statistically different from the other samplers.

Comparison by pairs of samplers

The box plot for wood dust mass concentration ratios between 15 different pairs of the samplers for all sampling sites is shown in Fig. 4. The median mass

concentration ratio for all pairs ranged from 0.62 (ACCU-CAPTM and CIP10-I pair) to 1.07 (CIP10-I and GSP pair). Table 2 shows correlation coefficient, average difference in mass concentration (from random effects model), and correction factors for each pair of samplers. All the pairs showed a strong and statistically significant ($P < 0.05$) correlation between samplers, except for pairs ACCU-CAPTM/ACCU-CAPTM, Button/CIP10-I, and Button/GSP. Average difference between the groups calculated from GM for each pair ranged from 0.03 (Button/ Button pair and GSP/IOM) to 6.66 (ACCU-CAPTM/ACCU-CAPTM) mg m⁻³ and all P -values are >0.05 , implying that the mass concentrations of wood dust from side-by-side sampling are not significantly different. The average difference between ACCU-CAPTM and ACCU-CAPTM is very large due to single large mass concentration (195 mg m⁻³). When this value is considered as outlier, the average difference between the samplers drops to just 0.01 mg m⁻³ and the correlation between pairs becomes significant with a coefficient of 0.83 ($P < 0.001$). The number of collected pairs is generally appropriate to detect a 35% difference with 80% power at an alpha level of 0.05. A difference of 35% may be considered more appropriate for field studies, as opposed to the NIOSH 25% criterion for laboratory studies, as field studies can have highly variable environmental

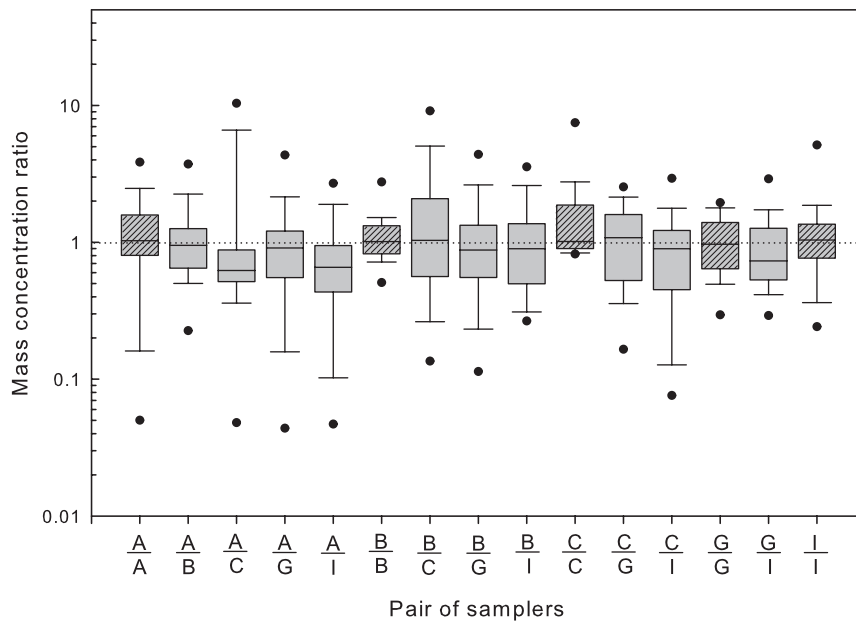


Fig. 4. Box plot of wood dust mass concentration ratios between the pairs of the samplers without outliers. The horizontal lines in the box plot from bottom to top indicate 10th, 25th, 50th (median), 75th, and 90th percentiles. The circles indicate the 5th (lower circle) and 95th (upper circle) percentiles. A is ACCU-CAP™, B is Button sampler, C is CIP10-I, G is GSP sampler, and I is IOM sampler (total number of pairs is 444).

conditions (Bartley *et al.*, 2007). Statistical evaluation of the nonsimilar sampler pair results produced a finding of no significant difference between any pairing of sampler type. Correction factors ranged from 0.16 (pair of ACCU-CAP™ and ACCU-CAP™ samplers) to 1.13 (pair of CIP10-I and CIP10-I samplers). Correction factor as defined here means that the average level of wood dust measured by the ACCU-CAP™ was the same as that of Button (for ACCU-CAP™ and Button samplers pair) when the stated correction factor (0.91) was applied to the concentration measured by the Button. The average differences between the identical sampler pairs (in Table 2) may reflect the ability of the sampler to collect more or less coarse particles. Variability increases for samplers, such as the IOM, expected to collect coarser particles.

DISCUSSION

Wood dust exposure levels

The average wood dust exposure levels in the present study were consistent with the previously reported studies, despite the fact that the sampling methods and sites were different in each study and the wood dust concentrations were not always TWA measurements (Moore *et al.*, 1990; Hamill

et al., 1991; Pisaniello *et al.*, 1991; Scheeper *et al.*, 1995; Martin and Zalk, 1998; Teschke *et al.*, 1999; Schlünssen *et al.*, 2001; Rando *et al.*, 2005; Scarselli *et al.*, 2007; Kalliny *et al.*, 2008; Galea *et al.*, 2009; Saejiw *et al.*, 2009; Yamanaka *et al.*, 2009; Kauffer *et al.*, 2010). Wood dust exposure levels may have decreased over recent decades possibly due to the changes in equipment, production methods, and upgrading engineering ventilation system for dust control (Teschke *et al.*, 1999; Galea *et al.*, 2009).

Sampler comparison

Many studies have compared and reported ratios between CFC and inhalable samplers. Moore *et al.* (1990) compared MSA respirable dust cassette that contains an aluminum inner capsule with the filter-only analysis from the standard 37-mm cassette and the mean mass concentration ratio was 2.2 implying that the filter-only analysis was underestimating exposure because of particles lost to the interior surfaces of the cassette. Werner *et al.* (1996) measured a variety of occupational aerosols including wood dust side-by-side sampling of CFC (filter-only) and IOM and they found that the ratio of IOM and CFC was 2.5. Martin and Zalk (1998) reported ratios of IOM to CFC (filter-only) in carpenter shops ranged from 1.8 to 4.1 (mean ratio 2.8) when the CFC concentrations were $>0.5 \text{ mg m}^{-3}$. Davies

et al. (1999) compared the GSP and CFC (filter-only) and found concentration ratios to be ~ 4.0 . Tatum *et al.* (2001) compared wood dust mass concentration between IOM and CFC (filter-only) and the average ratio was 1.85. Harper and Muller (2002) measured wood dust by side-by-side personal sampling using Button, CFC, and IOM samplers and median levels of sampler ratios for IOM/CFC, IOM/Button, and CFC/Button were 3.35, 3.15, and 1.2, respectively. Harper *et al.* (2004) also reported the ratios between Button, CFC, and IOM samplers with inclusion of particles deposited on the walls of the CFC sampler. The ratios IOM/CFC, IOM/Button, and CFC/Button were 2.16, 3.43, 1.59, respectively, with IOM/CFC and CFC/Button therefore significantly different from the earlier measurements of Harper and Muller (2002) where wall deposits had not been included. Hagström *et al.* (2008) reported that the ratio between IOM and 25-mm open-face cassette (filter-only) ranged from 0.67 to 17 with an average of 3.2 at wood pellets industries. Kauffer *et al.* (2010) reported ratios between selected aerosol samplers including ACCU-CAP™, Button, CIP10-I, and IOM samplers and the CFC (filter-only) in wood product industries and the ratios were 1.68, 1.46, 1.84, and 2.00, respectively. They also compared the various samplers to one another and found that concentrations measured by ACCU-CAP™, CIP10-I, and IOM samplers were not statistically different but the Button sampler concentrations were lower than the others, which was consistent with the study from Harper and Muller (2002). The difference ratios between the samplers might be attributed to the different size distribution of wood or occupational dust at sampling sites and different machining processes from each study. It is clear that inclusion of particles collected on the internal surfaces of the CFC provides results that are closer to those of samplers often considered to collect IPM, such as the IOM, the GSP, and the CIP10-I. Thus, the CFC with ACCU-CAP™ might behave more like an inhalable sampler as has previously been noted for samples of airborne metals (Demange *et al.*, 2002; Harper, 2006). Görner *et al.* (2010) reported that ACCU-CAP™ collected more airborne particles than CFC up to 10 and 20% in wind tunnel and calm air chamber, respectively. In this study, the ratio of ACCU-CAP™ and the IOM sampler was 1.56, which was not significantly different from 1. The closer match between the ACCU-CAP™ and IOM sampler found in this field study as well as the study of Kauffer *et al.* (2010) is probably because in the field particles are projected toward the worker in calm air rather than falling vertically

from above as in a calm air chamber (Lidén and Kenny, 1994). Our findings corroborate these results and show that using the ACCU-CAP™ decreases the difference of mass concentration between CFC and inhalable samplers to statistical nonsignificance. Note that other inserts designed for a similar purpose, such as the Woodchek (MSA, Inc., Pittsburgh, PA, USA) would likely provide the same result, although this was not tested in this study.

Sampler selection for wood dust measurement

While the most important consideration in sampling wood dust is the size fraction (i.e. inhalable, thoracic, or respirable), there are other important considerations, such as cost of sampling and acceptability to workers (comfort and placement of device) and industrial hygienists (ease of calibration and disposability). In the present study, all five samplers tested appear to provide similar results for wood dust measurement. Other factors become important in the final selection of sampler, when the choice is between samplers shown to have acceptably similar performance.

CONCLUSIONS

Wood dusts were collected in seven different wood product industries using five different total and inhalable samplers including the 37-mm CFC with ACCU-CAP™, Button, CIP10-I, GSP, and IOM, with side-by-side personal sampling. Approximately 30 replicates of each of 15 different combinations of samplers were collected as personal sample pairs. Statistical evaluation of the nonsimilar sampler pair results produced a finding of no significant difference between any pairing of sampler type implying that these samplers could be used interchangeably for personal wood dust exposure assessment.

A practical consideration for sampling in the USA is that the ACCU-CAP™ is similar to the sampler currently used by the OSHA for purposes of demonstrating compliance with its permissible exposure limit for wood dust, which is the same as for particles not otherwise regulated, also known as inert dust or nuisance dust (Method PV2121).

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Disclaimer—The findings and conclusions in this report are those of the author(s) and do not necessarily represent the official position of the Centers for Disease Control and Prevention/the Agency for Toxic Substances and Disease Registry. Mention of commercial equipment should not be construed to imply endorsement.

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