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Exposure to Heavy Metals and Hormone Levels in Midlife Women: the Study of Women's Health Across the Nation (SWAN)

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Abstract

Exposure to heavy metals may alter the circulating levels of sex hormones. However, epidemiologic studies on heavy metals and sex hormones have been limited, and results have been inconsistent. We assessed the associations of heavy metals assayed in urine, including arsenic, cadmium, lead, and mercury, with repeated measures of serum estradiol (E_2), follicle-stimulating hormone (FSH), testosterone, and sex hormone-binding globulin (SHBG) levels in the Study of Women's Health Across the Nation Multi-Pollutant Study. The sample included 1,355 White, Black, Chinese, and Japanese women, aged 45-56 years at baseline (1999-2000), whose serum hormone levels were repeatedly measured through 2017. Urinary metal concentrations were measured at baseline. Linear mixed effect models were used to calculate percent changes in serum hormone levels per doubling of urinary metal concentrations, adjusting for demographics, socioeconomic status, lifestyle, health-related factors, and urinary creatinine. After multivariable adjustment, a doubling of urinary metal concentration was associated with lower E_2 levels by 2.2%

Conflict of interest

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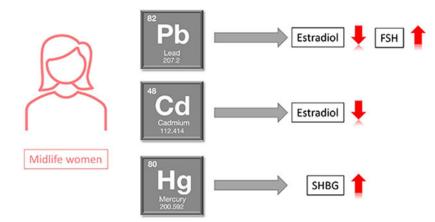
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Declaration of interests

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(95% CI: -4.0%, -0.3%) for mercury and 3.6% (95% CI: -5.7%, -1.6%) for lead; higher FSH levels by 3.4% (95% CI: 0.9%, 5.9%) for lead; and higher SHBG levels by 3.6% (95% CI: 1.3%, 5.9%) for cadmium. The overall joint effect using the Bayesian kernel machine regression showed that metal mixtures were inversely associated with E₂ and positively associated with FSH levels. No association was found between metals and testosterone levels. Results from this prospective cohort study demonstrate that environmental heavy metal exposures, including cadmium, mercury, and lead, may disturb circulating levels of E₂, FSH, and SHBG in midlife women.

Graphical Abstract



Keywords

Heavy metals; arsenic; cadmium; lead; mercury; sex hormones

1. Introduction

The menopausal transition is a significant health milestone for women, encompassing a time period during which physiologic changes mark progression toward the cessation of ovarian function. It is characterized by a shift in women's sex hormone profiles, including a relatively sharp decline in estradiol (E_2) and a gradual rise in follicle-stimulating hormone (FSH) levels toward the final menstrual period, owing to permanent changes in ovarian function (Randolph et al., 2011). By contrast, testosterone levels remained relatively stable (Randolph et al., 2003). Levels of E₂ or FSH over the menopausal transition may affect women's health in the midlife and as they age, as they influence risk of obesity (Sutton-Tyrrell et al., 2005; Wildman et al., 2012), type 2 diabetes mellitus (Park et al., 2017), cardiovascular disease (El Khoudary and Thurston, 2018), and osteoporosis (Neer, 2010). A comprehensive understanding of the factors associated with sex hormone profiles is of great importance for understanding and ultimately preventing health risks associated with ovarian aging. Evidence from epidemiologic studies indicates that genetic (Sowers et al., 2011), sociodemographic, and lifestyle factors (Randolph et al., 2011) may alter sex hormone profiles during this critical time window. However, the potential impacts of exposure to environmental chemicals are less well understood.

Heavy metals, including arsenic, cadmium, mercury, and lead, are widely dispersed in the environment, with drinking water, ambient air, food, and consumer products as possible exposure sources (Wang et al., 2019b). Arsenic, cadmium, mercury, and lead have been evaluated for their endocrine-disrupting properties (Dyer, 2007; Iavicoli et al., 2009; Sun et al., 2016). It is biologically plausible that exposure to these metals may affect hormone profiles through alteration in production and secretion of sex hormones, interaction with hormone receptors, interference with steroidogenesis, regulation of gonadal receptor expression, and/or direct alteration of gametogenesis (Iavicoli et al., 2009). Sex hormone-binding globulin (SHBG) is a glycoprotein binding globulin produced by the liver that transports sex steroids in human circulation and is therefore an important regulator of free fractions of E_2 and testosterone (Deswal et al., 2017). Dysregulation of SHBG due to heavy metal exposures may also potentially affect sex hormone profiles (Avvakumov et al., 2000; Tian et al., 2021).

Information on associations between heavy metals and sex hormones in human populations, however, is limited, and results have been inconsistent (Gallagher et al., 2010; Jackson et al., 2011; Kresovich et al., 2015; Krieg and Feng, 2011; Lei et al., 2015; Nagata et al., 2005; Pollack et al., 2011; Zheng et al., 2015). Most studies have been cross-sectional, with metal exposures and hormone levels measured at a single time-point, so that causal relationships were hard to be established. Additionally, studies of the associations between heavy metals and sex hormone levels during the critical window of midlife in women have been particularly limited.

Thus, to test the hypothesis that exposure to heavy metals may adversely affect sex hormone profiles, we investigated the associations of urinary concentrations of heavy metals, including arsenic, cadmium, mercury, and lead measured at baseline with serial serum levels of E_2 , FSH, testosterone, and SHBG in the Study of Women's Health Across the Nation (SWAN), a multi-site, multi-racial/ethnic prospective cohort study of midlife women.

2. Material and methods

2.1 Study population

SWAN is a prospective cohort study of midlife women designed to investigate physiological and psychosocial changes during the menopausal transition and other related health outcomes from mid-to-late life (Sowers et al., 2000). 3,302 women were recruited from seven study sites between 1996 and 1997, and each site included White women and one minority group (Black women at Boston, MA, Pittsburgh, PA, southeast Michigan, MI, and Chicago, IL sites; Hispanic women at the Newark, NJ site; Chinese women at the Oakland, CA site; and Japanese women at the Los Angeles, CA site). Inclusion criteria included between age 42 and 52, having an intact uterus and at least one ovary, without use of hormone therapy in the past three months, and having at least one menstrual period in the past three months. Institutional Review Board approval was obtained at each study site, and all participants provided signed informed consent at each study visit.

Women in this analysis were 1,400 participants in the SWAN Multi-Pollutant Study (SWAN-MPS) who had stored urine samples from the SWAN Repository between 1999

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and 2000 (Ding et al., 2020a; Wang et al., 2019). For these analyses, we excluded 18 with missing information on serum hormone levels, 23 with missing information on key covariates, and 4 participants using hormone therapy at the SWAN-MPS baseline, yielding a final analytic sample of 1,355 participants. In SWAN, FSH and SHBG were measured through the 15^{th} follow-up visit (2015-2017), E₂ through the 13^{th} follow-up visit (2011-2012), and testosterone through the 10^{th} follow-up visit (2006-2007). The timeframe of sex hormone measurements was shown in Figure S1. We censored observations (n=2,244) in follow-up visits when a participant was taking hormone therapy because the true untreated levels of the outcome parameters were unknown. A final sample of 1,355 women representing 10,645 observations was used for FSH and SHBG analyses, 9,754 for E₂ analyses, and 7,745 for testosterone analyses.

2.2 Serum sex hormones

 E_2 , FSH, testosterone, and SHBG were assayed from fasting serum samples at each visit. Optimally, serum samples were collected for each participant before 10:00 A.M. on one day during days 2-5 of a spontaneous menstrual cycle occurring within 60 days of recruitment at the baseline visit and approximately annually thereafter. If a follicular phase sample could not be obtained, a random fasting sample was taken within 90 days of the anniversary of the baseline visit. FSH was assayed in singlicate and measured with a two-site chemiluminometric immunoassay using the automated Ciba Corning Diagnostics ACS-180 analyzer (Bayer Diagnostics Corp., Norwood, MA); the inter- and intra-assay coefficients of variation (CVs) were 12% and 6%, respectively. E_2 was assayed in duplicate and measured with a modified, off-line ACS-180 (E2–6) immunoassay, with inter- and intra-assay CVs of 11% and 6%, respectively. Testosterone was assayed using the modified rabbit polyclonal anti-T ACS-180 immunoassay, with the average inter- and intra-assay CVs of 10% and 8%, respectively. SHBG was assayed using rabbit anti-SHBG antibodies, with inter- and intra-assay CVs of 10% and 6%, respectively.

2.3 Urinary metals

High-resolution inductively coupled plasma-mass spectrometry (ICP-MS) (Thermo Scientific iCAP RQ, Waltham, MA) was used to determine arsenic, cadmium, mercury, and lead concentrations in morning urine samples collected at SWAN-MPS baseline at the Applied Research Center of NSF International (Ann Arbor, Michigan), a part of the Michigan Children's Health Exposure Analysis Resource (M-CHEAR) Laboratory Hub. Details of analytic methods and quality control procedures have been described previously (Wang et al., 2019b). Limits of detection (LODs) and detection rates are presented in Table S1. Participants with metal concentrations below the LOD were assigned a value equal to the LOD divided by the square root of 2. Spearmen correlations between metal concentrations were calculated.

2.4 Covariates

Age, race/ethnicity (White, Black, Chinese, or Japanese), and education level (as high school or less, some college, or college degree or higher) were collected using a self-administered questionnaire at the SWAN baseline. Considering the possibility that metal exposures may affect BMI, we did not use time-varying BMI in the analysis in order to prevent potential

overadjustment bias (Wang et al., 2018). Urinary creatinine was measured using the Cobas Mira analyzer (Horiba ABX, Montpellier, France) at the SWAN-MPS baseline as a marker of urine dilution. At each study visit, smoking status (never smoked, former smoked only, or current smoking), menopause transition stage (pre-menopausal, early perimenopausal, late perimenopausal, postmenopausal, and unknown due to hormone therapy use), and parity (nulliparous, parous) were self-reported. Pre-menopausal was defined as having menstruation in past 3 months with no change in bleeding regularity; early perimenopausal was having menstruation in past 3 months but decreasing regularity between menses; late perimenopausal was no menstruation in past 3-11 months; and postmenopausal was no menstruation in past 12 or more months.

2.5 Statistical analysis

We used linear mixed effects models with random intercepts to evaluate the associations between baseline urinary metal concentrations and longitudinally measured serum hormone levels. Given the highly skewed distributions of hormone levels and metal concentrations, logarithmic transformations were applied so that shapes of exposure-outcome relationships were more closely log-linear. All regression coefficients and associated 95% confidence intervals (95% CIs) were back-transformed and expressed as percent change in hormone levels for a doubling increase in metal concentrations. Potential confounders were adjusted progressively. The base model included non-time-varying variables age (baseline), race/ ethnicity, study site, and urinary creatinine (log-transformed). The fully adjusted model further adjusted for non-time-varying variables education, BMI (baseline), and parity; and time-varying variables included follow-up time, smoking and menopause status. Multiple comparison for each hormone was addressed at a false discovery rate (FDR) of 0.05 using the Benjamini–Hochberg Method (Benjamini and Hochberg, 1995).

In secondary analyses, we quantified differences in hormone levels associated with exposure to metal mixtures using a two-stage modeling strategy (Wang et al., 2020), addressing correlations for both exposures and repeatedly measured outcomes. In stage 1, linear mixed effects models with random intercepts were fitted to estimate participant-specific hormone levels while accounting for correlations in hormone levels within each participant over time. In stage 2, the Bayesian kernel machine regression (BKMR) was used to examine the associations between metal mixtures and participant-specific hormone levels estimated from stage 1, while handling the correlations between metal exposures and potential non-linear relationships (Bobb et al., 2015). It should be noted that we used the two-stage modeling approach instead of specifying a random intercept within BKMR because of the computational limits. The R package 'bkmr' was used to implement BKMR (Bobb et al., 2015).

In other secondary analyses, we repeated analyses stratified by menopausal status and race/ethnicity to evaluate potential effect modifications. Additionally, we evaluated the associations between metal concentrations and log-transformed testosterone-to- E_2 ratio. Finally, we further adjusted for stress, dietary intake of fat, alcohol drinking, and physical activity. All data analyses were performed using SAS version 9.4 (SAS Institute Inc.) and R version 4.2.1 (www.R-project.org).

3. Results

3.1 Descriptive statistics

The mean (SD) age was 49.3 (2.6) years (Table 1). The proportions of each racial/ethnic group were 50.0% for white, 21.8% for Black American, 13.1% for Chinese, and 15.2% for Japanese. Most women had at least some college education and had never smoked. The geometric mean (GM) (geometric standard deviation, GSD) at baseline was 34.5 (2.7) pg/mL for E₂, 34.5 (2.8) mIU/mL for FSH, 32.3 (1.7) ng/dL for testosterone, and 35.0 (2.0) nM for SHBG. The detection rate and distributions of urinary metal concentrations is presented in Table S1. All four metals were positively correlated, with the strongest correlation between cadmium and lead (r=0.57, Figure S2).

3.2 Associations between sex hormones and metals

Significant inverse associations of serum E_2 levels with cadmium, mercury, and lead were observed in the base models, whereas the associations were attenuated and only significant for mercury and lead in the full model after adjusting for multiple comparison with FDRs < 0.05 (Table 2). After adjustments for age, race/ethnicity, study site, education, BMI at baseline, smoking, parity, menopausal status, and urinary creatinine (log-transformed), a doubling increase in urinary metal concentration was associated with 2.2% (95% CI: -4.0%, -0.3%) lower E_2 level for mercury, and 3.6% (95% CI: -5.7%, -1.6%) for lead.

In the base models, significant positive associations with FSH levels were found for cadmium, mercury, and lead (Table 3). Adjustment for additional covariates in the full model attenuated these associations, and significant association was only observed for lead after controlling for multiple comparison. After full adjustment for covariates, a doubling increase in urinary lead concentration was associated with 3.4% (95% CI: 0.9%, 5.9%) higher FSH level.

Significant positive associations of serum SHBG levels with arsenic, cadmium, and mercury were found in the base model, whereas results were significant only for cadmium in the full model after controlling for multiple comparison (Table 4). A doubling increase in urinary cadmium concentration was associated with 3.6% (95% CI: 1.3%, 5.9%) higher serum SHBG levels after fully adjusting for covariates. No association was observed between urinary metal concentrations and serum testosterone levels (Table S2).

3.3 Secondary analyses

In the secondary analysis evaluating the associations between metal mixtures and hormone levels, we found that lead was inversely associated with E_2 levels, as observed in the primary analyses (Figure S3). Similarly, a positive association was observed between lead and FSH (Figure S4) and between cadmium and SHBG (Figure S5). No associations between metals and testosterone were observed (Figure S6). Figure 1 shows the overall effects of metal mixtures as the estimated changes in log-transformed hormone levels, comparing concentrations of all metals simultaneously at different percentiles of their distributions to all metals at their 10th percentiles. Metal mixtures were inversely associated with E_2 and positively associated with FSH levels.

In the secondary analysis stratifying the associations by menopausal status (pre-menopausal, early and late perimenopausal, and postmenopausal) (Table S3), we found stronger inverse associations of E_2 with mercury and lead and a stronger positive association between FSH and lead in the early and late perimenopausal stages. Additionally, arsenic was inversely associated with E_2 and positively associated with SHBG in the early and late perimenopausal stages. Mercury was positively associated with FSH in the early and late perimenopausal stages. In stratified analysis by race/ethnicity, we found stronger positive associations between lead and FSH in Japanese women (Table S4). We also observed an inverse association between cadmium and testosterone in Black women.

In other secondary analyses, positive associations of testosterone-to- E_2 ratio (log-transformed) with mercury and lead were observed (Table S5). Similar associations were observed after additional adjustment for stress, dietary intake of fat, alcohol drinking, and physical activity (Table S6).

4. Discussion

This study examined associations between heavy metal exposures and longitudinal measures of sex hormone levels in midlife women as they aged through the menopausal transition. Higher urinary mercury and lead concentrations were associated with lower serum E_2 levels. Higher lead concentrations were associated with higher serum FSH levels. A positive association between cadmium concentration and serum SHBG levels was also observed. Stratification by menopausal status revealed stronger associations in early and late perimenopausal women.

Several hormones in the hypothalamic-pituitary-ovarian axis are markers of ovarian aging, including FSH and E_2 (Strauss and Barbieri, 2013). The progressive depletion of follicles and cessation of ovarian function during the menopausal transition lead to altered ovarian feedback in the hypothalamic-pituitary-ovarian axis, resulting in a reduction in E2 and subsequent elevation in FSH (Strauss and Barbieri, 2013). The observed associations of heavy metals with FSH and E2 in the present study, particularly in the early and late perimenopausal stages, suggested that heavy metals may influence ovarian aging. Previous studies have shown that lower E2 in postmenopausal is associated with cardiovascular disease (Zhao et al., 2018), lower bone mineral density and osteoporosis (Ettinger et al., 1998), depression (Bromberger et al., 2010), and other conditions (Notelovitz et al., 2000). This finding is also in agreement with our recent report on the association between heavy metal exposures and earlier age at natural menopause in SWAN (Wang et al., 2021). Our findings are of public health importance, given the fact that the alterations in hormone profiles accelerate ovarian aging and may lead to reproductive, physiologic, psychological, and behavioral changes that shape women's midlife and future health (El Khoudary et al., 2019).

One major finding of this study was that urinary lead was associated with higher FSH and lower E_2 levels, suggesting its potential effect on accelerated ovarian aging. Evidence from epidemiologic studies supports that lead exposures may disrupt sex hormones in women. For example, three large population-based, cross-sectional studies found that blood lead

concentration was associated with elevated serum FSH levels among premenopausal women (Chen et al., 2016; Krieg and Feng, 2011; Lee et al., 2019); however, in other studies with smaller samples, associations between lead and FSH were not observed (Jackson et al., 2011; Pollack et al., 2011). No previous studies have reported a significant association between lead and E₂ (Chen et al., 2016; Jackson et al., 2011; Pollack et al., 2011). Although the biological pathways that may underlie the association of lead with FSH and E_2 are still largely unknown; findings from mechanistic studies suggest that lead may affect hormone profiles in several ways. For example, in an in vitro study of human ovarian granulosa cells, lead treatment led to reduced production of p450 aromatase messenger RNA and cytochrome p450 aromatase, which is required for the transformation of androgen to E₂ (Taupeau et al., 2003). Lead can also directly disrupt the hypothalamic-pituitary axis. A study of lead exposure in rats reported that long-term, low-dose treatment of lead significantly upregulated the production of gonadotropin-releasing hormone, which further triggers the secretion of FSH (Sokol et al., 2002). Additionally, lead has been suggested to increase concentrations of homocysteine, which play a role in the stimulation of FSH release (Schafer et al., 2005).

Our finding of a positive association between cadmium and SHBG levels is in line with the positive associations between blood cadmium and serum SHBG reported in 869 adult men in NHANES (Kresovich et al., 2015) and in a U.S. prospective cohort of 251 healthy premenopausal women (Kim et al., 2021). SHBG is a transporter of both E_2 and testosterone in human circulation, and its levels are important in the regulation of free and albumin bound levels of these hormones (Deswal et al., 2017). The observed positive association between cadmium and SHBG might indicate reduced bioavailable free and albumin-bound E_2 levels due to cadmium exposure. Alternatively, the observed association could be the result of competitive binding of cadmium to SHBG, which in turn stimulates SHBG production (Avvakumov et al., 2000). SHBG is produced primarily in liver. Cadmium can induce hepatic necroinflammation and lead to non-alcoholic fatty liver disease (NAFLD) and metabolic disorders (Souza-Arroyo et al., 2022). SHBG is considered a hepatokine and is involved in the development of NAFLD and metabolic impairments (Qu and Donnelly, 2020). The observed associations may also be due to the potential liver toxicity of cadmium among midlife women. We also observed urinary cadmium was positively associated with FSH in the early and late perimenopausal stages. In the National Health and Nutrition Examination Survey (NHANES) III, urinary cadmium was associated with higher serum FSH levels in premenopausal women (Gallagher et al., 2010). In contrast, a prospective study of 252 premenopausal women in the U.S. found a positive association between blood cadmium and serum FSH (Pollack et al., 2011). However, in a large cross-sectional study of postmenopausal women in South Korea, no significant association between urinary cadmium and serum FSH was found (Lee et al., 2019). Cadmium has been identified as an endocrine disruptor linked with alteration in ovarian cell morphology in human ovarian granulosa cell cultures (Paksy et al., 1997) and has been found to suppress the secretion of progesterone in animal models (Zhang et al., 2008), which in turn upregulates FSH levels (Ding et al., 2020).

Urinary mercury was inversely associated with E_2 levels in our study. The general population can be exposed to methylmercury (organic form) primarily through consumption

of seafood and to inorganic mercury through occlusal surfaces of teeth that are filled with mercury-containing amalgams (Mutter, 2011) or through skin lightening products (Park and Zheng, 2012). In humans, urinary mercury mainly reflects inorganic mercury (ATSDR, 1999). Epidemiologic studies examining the relationship between mercury and E_2 have been few, and none has reported a significant association (Jackson et al., 2011; Lee et al., 2019; Pollack et al., 2011). Our findings add a reference for an association of mercury, particularly the inorganic form, with reduced E_2 in midlife women. The potential impact of mercury exposure on E_2 is supported by several experimental animal studies, although the underlying mechanism is still unclear. In a study in rats, exposure to inorganic mercury has also been detected within ovarian follicles and in the corpora lutea of rats and golden hamsters after chronic exposure accompanied by decreased ovulation (Lamperti and Printz, 1974; Stadnicka, 1980).

Urinary arsenic was inversely associated with E_2 and positively associated with SHBG levels in early and late perimenopausal women. In a case-control study of women with primary ovarian insufficiency and healthy controls, urinary arsenic was positively associated with the odds of primary ovarian insufficiency and inversely associated with E_2 levels (Pan et al., 2020). Urinary arsenic was also associated with higher odds of infertility in another case-control study of infertile and pregnant women (Lei et al., 2015). In studies of rats, exposure to inorganic arsenic has been found to inhibit ovarian activities of 3β and 17β -hydroxysteroid dehydrogenase, induce degeneration of luminal epithelial, stromal and myometrial cells of the rat uterus, and downregulate the estrogen receptor and estrogenresponsive genes in the estrogen signaling pathway, leading to reduced serum E_2 levels (Chatterjee and Chatterji, 2010; Chattopadhyay and Ghosh, 2010).

The primary strength of our study is its use of a large, prospective cohort with prospective and reliable measures of serum hormones, menopausal status, and covariates over 17 years of follow-up as women transitioned through menopause. To our knowledge, this is the first large prospective cohort study investigating the associations of heavy metal exposures with repeated measures of sex hormones in midlife women. Standard annualized measures also provided reliable estimates of hormone levels and ensure temporality between exposure and outcome.

Several limitations should be considered as well. First, all metal concentrations were determined in urine samples, and urinary concentrations may not reflect the total body burden from all forms of these metals or their exposure sources. Second, we relied on baseline measures of all metals to examine their associations with longitudinal measures of hormones. Existing evidence demonstrates long-term consistency in arsenic exposure, and urinary arsenic provides the most accurate exposure assessment (Navas-Acien et al., 2009). Urinary cadmium has been proposed as a biomarker for long-term exposure because cadmium is not rapidly excreted and has a half-life that ranges from years to decades (Vacchi-Suzzi et al., 2016). Urinary lead adjusted for urine dilution has been used as a biomarker of lead from bone resorption; bone lead is a biomarker of long-term exposure and has been suggested as an endogenous and primary source of lead in mid-and late life (Wang et al., 2019a). Nevertheless, reliance on one-time measurement of metal concentrations

might underestimate the associations between metals and hormones. Third, in SWAN, E_2 levels were assessed in the early follicular phase, which does not represent the peak E_2 levels in premenopausal women. Thus, the impact of metal exposures on E_2 could still be underestimated.

5. Conclusions

The results from this prospective cohort study demonstrate that environmental heavy metal exposures, including cadmium, mercury, and lead, may disturb circulating levels of E_2 , FSH, and SHBG in midlife women. Given the known link between sex hormones and health endpoints, our findings provide important information about potential factors of environmental origins that may affect women's midlife health and healthy aging. Our findings also highlight the need to avoid exposure to heavy metals by reducing possible environmental sources and modifying lifestyle exposure to improve midlife and older women's health.

Supplementary Material

Refer to Web version on PubMed Central for supplementary material.

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Abbreviations:

BMI	body mass index	
CV	coefficients of variation	
E ₂	estradiol	
FSH	follicle-stimulating hormone	
GM	geometric mean	
GSD	geometric standard deviation	
LOD	limit of detection	
SHBG	sex hormone-binding globulin	
SWAN	AN Study of Women's Health Across the Nation	
SWAN-MPS	Study of Women's Health Across the Nation Multi-Pollutant Study	

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Highlights

• Heavy metal exposures disrupted sex hormone levels in midlife women.

- Lead was associated with lower E₂ and higher FSH levels.
- Mercury was associated with lower E₂ levels.
- Cadmium was associated with higher SHBG levels.

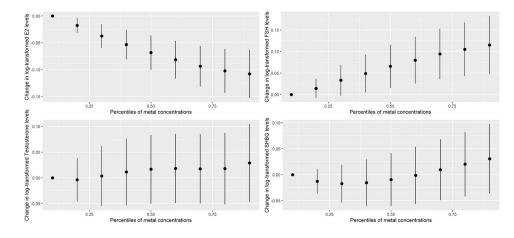


Figure 1.

Associations between heavy metal concentrations and sex hormone levels, estimated by Bayesian kernel machine regression. The plots show the estimated changes and 95% confidence intervals (95% CIs) in log-transformed hormone levels, comparing concentrations of all metals simultaneously at different percentiles of their distributions to all metals at their 10th percentiles. The models were adjusted for age, race/ethnicity, study site, urinary creatinine (log-transformed), education, baseline body mass index, smoking, parity, menopausal status, stress, dietary intake of fat, alcohol drinking, and physical activity.

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Table 1.

Descriptive characteristics at the SWAN-MPS baseline (n=1,355).

Characteristics	Mean (SD), GM (GSD), or n (%)
Age, mean (SD), years	49.3 (2.6)
Race/ethnicity, n (%)	
White	677 (50.0)
African American	295 (21.8)
Chinese	177 (13.1)
Japanese	206 (15.2)
Study site, n (%)	
Southeast Michigan	248 (18.3)
Boston	226 (16.7)
Oakland	302 (22.3)
Los Angeles	354 (26.1)
Pittsburgh	225 (16.6)
Education, n (%)	
High school or less	245 (18.2)
Some College	434 (32.2)
College	333 (24.7)
Post-college	336 (24.9)
Smoking status, n (%)	
Never	854 (63.0)
Former	359 (26.5)
Current	142 (10.5)
Parity, n (%)	
Nulliparous	262 (19.3)
Parous	1093 (80.7)
Menopausal status, n (%)	
Premenopausal	166 (12.3)
Early perimenopausal	749 (55.3)
Late perimenopausal	125 (9.2)
Postmenopausal	302 (22.3)
Unknown ^a	13 (1.0)
Body mass index, mean (SD), kg/m ²	27.95 (7.3)
Estradiol, GM (GSD), pg/mL	34.5 (2.7)
Follicle-stimulating hormone, GM (GSD), mIU/mL	34.5 (2.8)
Testosterone, GM (GSD), ng/dL	32.3 (1.7)
Sex hormone-binding globulin, GM (GSD), nM	35.0 (2.0)

Note: SD: standard deviation; GM: geometric mean; GSD: geometric standard deviation.

 $^a\!\mathrm{Menopausal}$ status unknown due to hysterectomy or hormone the rapy.

Table 2.

Percent changes (95% CIs) in serum levels of estradiol (E_2) per doubling increase in urinary metal concentrations.

Metals	Base model ^a	Full Model ^b
	Percent changes (95% CI)	Percent changes (95% CI)
Arsenic	1.3 (-2.8, 0.3)	-0.8 (-2.2, 0.5)
Cadmium	-3.5 (-5.5, -1.5)	-1.5 (-3.3, 0.3)
Mercury	-3.2 (-5.3, -1.1)	-2.2 (-4.0, -0.3)
Lead	-6.6 (-8.8, -4.3)	-3.6 (-5.7, -1.6)

^aBase model: adjustment for age, race/ethnicity, study site, and urinary creatinine (log-transformed).

b Full model: base model with additional adjustment for education, baseline body mass index, smoking, parity, follow-up time, and time-varying menopausal status.

Table 3.

Percent changes (95% CIs) in serum levels of follicle stimulating hormone (FSH) per doubling increase in urinary metal concentrations.

Metals	Base model ^a	Full Model ^b
	Percent changes (95% CI)	Percent changes (95% CI)
Arsenic	0.1 (-1.8, 2.1)	-0.9 (-2.4, 0.6)
Cadmium	5.3 (2.7, 8.0)	1.8 (-0.3, 4.0)
Mercury	3.4 (0.7, 6.2)	0.8 (-1.3, 2.9)
Lead	7.8 (4.7, 11.1)	3.4 (0.9, 5.9)

 a Base model: adjustment for age, race/ethnicity, study site, and urinary creatinine (log-transformed).

 $b_{\rm Full}$ model: base model with additional adjustment for education, baseline body mass index, smoking, parity, follow-up time, and time-varying menopausal status.

Table 4.

Percent changes (95% CIs) in serum levels of sex hormone-binding globulin (SHBG) per doubling increase in urinary metal concentrations.

Metals	Base model ^{<i>a</i>}	Full Model ^b
	Percent changes (95% CI)	Percent changes (95% CI)
Arsenic	2.0 (0.3, 3.8)	1.5 (-0.1, 3.2)
Cadmium	4.8 (2.4, 7.2)	3.6 (1.3, 5.9)
Mercury	3.9 (1.5, 6.4)	1.4 (-0.8, 3.7)
Lead	2.5 (-0.2, 5.3)	1.5 (-1.1, 4.1)

 a Base model: adjustment for age, race/ethnicity, study site, and urinary creatinine (log-transformed).

 $b_{\rm Full}$ model: base model with additional adjustment for education, baseline body mass index, smoking, parity, follow-up time, and time-varying menopausal status.